Food and Agriculture Organization of the United Nations

## Pesticide residues in food 2016

## Special Session of the Joint FAO/WHO Meeting on Pesticide Residues

## REPORT

 2016
# Pesticide residues in food 2016 

## Joint FAO/WHO Meeting on Pesticide Residues

Peport of the special session of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Core Assessment Group on Pesticide Pesidues Geneva, Switzerland, 9-13 May 2016

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## T, toxicological evaluation

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Geneva, 9-13 May 2016

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[^2]|  | ABBREVIATIONS |
| :---: | :---: |
| ADI | acceptable daily intake |
| AHS | Agricultural Health Study |
| AMPA | aminomethylphosphonic acid |
| ARfD | acute reference dose |
| BMD | benchmark dose |
| bw | body weight |
| CAS | Chemical Abstracts Service |
| CCPR | Codex Committee on Pesticide Residues |
| CI | confidence interval |
| CYP | cytochrome P450 |
| DCF | diet correction factor |
| DNA | deoxyribonucleic acid |
| $\mathrm{F}_{0}$ | parental generation |
| $\mathrm{F}_{1}$ | first filial generation |
| FAO | Food and Agriculture Organization of the United Nations |
| GEMS/Food | Global Environment Monitoring System - Food Contamination Monitoring and Assessment Programme |
| GLP | good laboratory practice |
| HR | highest residue in the edible portion of a commodity found in trials used to estimate a maximum residue level in the commodity |
| HR-P | highest residue in a processed commodity calculated by multiplying the HR of the raw commodity by the corresponding processing factor |
| IARC | International Agency for Research on Cancer |
| IEDI | international estimated daily intake |
| IESTI | international estimate of short-term dietary intake |
| ISO | International Organization for Standardization |
| IUPAC | International Union of Pure and Applied Chemistry |
| JMPR | Joint FAO/WHO Meeting on Pesticide Residues |
| $\mathrm{LC}_{50}$ | median lethal concentration |
| $\mathrm{LD}_{50}$ | median lethal dose |
| LOAEL | lowest-observed-adverse-effect level |
| MDCA | malathion dicarboxylic acid |
| MIC | minimum inhibitory concentration |
| MMCA | malathion monocarboxylic acid |
| MRL | maximum residue limit |
| NC | no national consumption data available |


| NES | not elsewhere specified |
| :--- | :--- |
| NHL | non-Hodgkin lymphoma |
| NOAEC | no-observed-adverse-effect concentration |
| NOAEL | no-observed-adverse-effect level |
| OR | odds ratio |
| PP | processed product |
| ppm | parts per million |
| RAC | relative risk agricultural commodity |
| RR | supervised trials median residue <br> STMR |
| STMR-P | multiplying the STMR of the raw commodity by the corresponding processing <br> factor |
| TAF | toxicity adjustment factor |
| WHO | World Health Organization |

## USE OF JMPR REPORTS AND EVALUATIONS BY REGISTRATION AUTHORITIES

Most of the summaries and evaluations contained in this report are based on unpublished proprietary data submitted for use by JMPR in making its assessments. A registration authority should not grant a registration on the basis of an evaluation unless it has first received authorization for such use from the owner of the data submitted for the JMPR review or has received the data on which the summaries are based, either from the owner of the data or from a second party that has obtained permission from the owner of the data for this purpose.

## PESTICIDE RESIDUES IN FOOD

## REPORT OF THE MAY 2016 JOINT FAO/WHO MEETING OF EXPERTS

## 1. INTRODUCTION

A Joint Meeting of the Food and Agriculture Organization of the United Nations (FAO) Panel of Experts on Pesticide Residues in Food and the Environment and the World Health Organization (WHO) Core Assessment Group on Pesticide Residues (JMPR) was held at WHO Headquarters, Geneva (Switzerland), from 9 to 13 May 2016.

The meeting was opened by Dr Kazuaki Miyagishima, Director of the Department of Food Safety and Zoonoses, WHO, who welcomed participants on behalf of the Directors General of WHO and FAO. Dr Miyagishima stated that the meeting was convened to re-evaluate three compounds for which new studies had become available since their last full assessments. He reminded the participants of the importance of the functional separation between risk assessment and risk management and of the role that JMPR plays as the expert risk assessment body providing scientific advice to Codex and to Member States. He urged the participants to be guided by JMPR's standing rules and procedures based on the weight of evidence approach. Dr Miyagishima thanked the participants for devoting significant time and effort to the work of JMPR, including the preparatory work of paramount importance that had taken place in the past months. He reminded the experts that they were invited as independent experts acting in their own individual capacities and not as representatives of their countries or organizations. He also reminded the participants of the confidential nature of the meeting, in order to allow experts to freely express their opinions.

During the meeting, the WHO Core Assessment Group was responsible for reviewing epidemiological, toxicological and related data in order to establish acceptable daily intakes (ADIs) and acute reference doses (ARfDs), where necessary. As no residue data were requested, the FAO Expert was responsible for estimating the dietary exposures (both short-term and long-term) to the pesticides reviewed and, on this basis, performed dietary risk assessments in relation to their ADIs or ARfDs.

The Meeting re-evaluated three pesticides, established ADIs and ARfDs and recommended them for use by the Codex Committee on Pesticide Residues (CCPR). The Meeting also considered issues related to the evaluation of genotoxicity and epidemiological studies in relation to the risk assessment of chemicals.

### 1.1 DECLARATION OF INTERESTS

The Secretariat informed the Meeting that all experts participating in the May 2016 JMPR had completed declaration of interest forms and that no conflicts had been identified.

## 2. GENERAL CONSIDERATIONS

### 2.1 GENERAL CONSIDERATIONS ON THE EVALUATION OF GENOTOXICITY STUDIES

A large number of genotoxicity studies were evaluated during the present meeting. These were identified through direct submission to JMPR, searches of the publicly available literature and requests to the International Agency for Research on Cancer (IARC) Monographs Secretariat and industry groups. The studies evaluated included unpublished (primarily guideline) studies submitted to support pesticide registration as well as peer-reviewed studies published in the scientific literature. The number, quality and relevance of studies differed widely for each chemical and necessitated that a somewhat different approach be used to evaluate each pesticide. As a general strategy, the studies were separated into categories based largely on phylogenetic relevance and significance of the genetic end-point measured. The categories used were human biomonitoring, in vivo mammals, in vitro mammalian cells, in vitro bacteria, phylogenetically distant organisms, metabolites in vivo and metabolites in vitro. The evaluation was conducted for the pesticide active ingredient, its formulation products and prominent metabolites, as data were available. For the three pesticides evaluated, the human biomonitoring studies were most often confounded by exposures to other pesticides or considered to have other limitations. Among the genotoxicity studies, in vivo studies in mammals were given the greatest weight, compared with cell culture studies or investigations in phylogenetically distant organisms. Studies of gene mutations and chromosomal alterations were also given more weight than studies measuring other less serious or transient types of genotoxic damage. With regard to route of exposure, studies in which chemicals were administered by the oral route were considered to be of most relevance for evaluating low-level dietary exposures.

Following an evaluation and weighting of the studies, taking the criteria described above and the quality of the studies into account, an overall weight of evidence approach was used to reach conclusions about the genotoxicity of the individual pesticides. An important aspect of the evaluation was whether the genotoxic effect would be likely to occur in humans exposed to low levels of the pesticide present as residues in food.

The Meeting recommended that a guidance document be developed for the evaluation of genotoxicity studies, taking the experience gained from this meeting into account.

### 2.2 METHODS FOR THE EVALUATION OF EPIDEMIOLOGICAL EVIDENCE FOR RISK ASSESSMENT

## Identification of compound/cancer sites and screening of papers

There is a large body of literature regarding pesticide exposures and non-cancer outcomes (neurodevelopmental, neurodegenerative and reproductive outcomes, among other health outcomes), but the assessment of the epidemiological evidence on diazinon, glyphosate and malathion was restricted to studies of cancer outcomes. This restriction was partly driven by feasibility reasons: a clinically relevant adverse effect size (or an acceptable level of risk) for a non-cancer outcome must be defined, and the methodologies for hazard identification and characterization based on observational epidemiological findings of non-carcinogenic adverse effects are less well established than those for cancer. ${ }^{1}$

[^3]The IARC Monographs on malathion, diazinon and glyphosate referred to a total of 45 epidemiological studies. ${ }^{1}$ Databases were searched for any relevant articles published after the studies cited in these Monographs using the following search terms: [(diazinon OR glyphosate OR malathion) AND cancer] and [(diazinon OR glyphosate OR malathion) AND (NHL OR lymphoma OR leukemia OR "lung cancer" OR "prostate cancer")] in PubMed (limited to Humans; published in the last 5 years) and Scopus (limited to 2014-2016). Two studies published since the publication of the IARC Monographs that evaluated at least one of malathion, diazinon or glyphosate were identified in relation to cancer outcomes. ${ }^{2}$ An additional study on prostate cancer, ${ }^{3}$ which was not included in the IARC Monographs, was also identified.

The pre-agreed evaluation process shown in Fig. 1 was used to (1) select compound/cancer site combinations to include in this evaluation; (2) screen papers for inclusion/exclusion in this evaluation (Tier 1 screening criteria); and (3) evaluate the information available for risk assessment. In this process, it was noted that there were stand-alone analyses for specific subtypes of nonHodgkin lymphoma (NHL). The risk for subtypes of NHL was not evaluated separately, as there was insufficient evidence (too few studies or small numbers of cases); the risk for other haematopoietic and lymphoid tumours was also not evaluated separately, as the positive associations identified by IARC were for total NHL.


[^4]
## Evaluation of evidence for the compound/cancer site associations

Several aspects of each study and of all studies combined were considered in this evaluation, including factors that decrease the level of confidence in the body of evidence, such as risk of bias, unexplained inconsistency and imprecision; and factors that increase the level of confidence, such as large magnitude of effect, dose-response and consistency. ${ }^{1}$ The findings for each study were summarized in tables, and risk estimates for non-quantitative exposure assessment (predominantly ever versus never use) were summarized in forest plots.

## Evaluation of information available for risk assessment/hazard characterization

To evaluate overall evidence for dose-response relationships, risk estimates were plotted against quantitative exposure measures (for studies that had used these). The most commonly used quantitative exposure metric was days of use per year. Where studies had used other quantitative exposure metrics (e.g. lifetime days of exposure), data were requested from the authors on median "days of use per year" for the participants in each of the original exposure categories, although this information was not always forthcoming. These additional data allowed the translation and plotting of risk estimates from different studies on the same exposure scale (days of use per year).

[^5]
## 3. EVALUATION OF DATA FOR ACCEPTABLE DAILY INTAKE AND ACUTE REFERENCE DOSE FOR HUMANS

### 3.1 DIAZINON (22)

## TOXICOLOGY

Diazinon is the common name approved by the International Organization for Standardization (ISO) for $O, O$-diethyl $O$-2-isopropyl-6-methylpyrimidin-4-yl phosphorothioate (International Union of Pure and Applied Chemistry [IUPAC]), with the Chemical Abstracts Service (CAS) number 333-41-5.

Diazinon is a contact organophosphorus insecticide with a wide range of insecticidal activity. It is effective against adult and juvenile forms of flying insects, crawling insects, acarians and spiders. Diazoxon, the biologically active metabolite of diazinon, inhibits the activity of cholinesterases.

Diazinon is used mainly as a pesticide in agriculture and as a drug in veterinary medicine. Thus, the major source of diazinon residues in edible crops is from its use as an agricultural pesticide; residues in meat, offal and other animal products arise from its use as a veterinary drug containing active ingredient.

Diazinon has been evaluated by JMPR on several occasions since the first evaluation in 1963. In the most recent evaluation, in 2006, the Meeting established an ADI of $0-0.005 \mathrm{mg} / \mathrm{kg}$ body weight (bw), based on a no-observed-adverse-effect level (NOAEL) of $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day for inhibition of erythrocyte acetylcholinesterase activity in a 92-day repeated-dose toxicity study in rats. The 2006 Meeting reaffirmed the ARfD of $0.03 \mathrm{mg} / \mathrm{kg}$ bw, established by the 2001 JMPR , based on a NOAEL of $2.5 \mathrm{mg} / \mathrm{kg}$ bw observed in a study of acute neurotoxicity in rats.

Diazinon was scheduled within the periodic review programme of CCPR for 2021. The compound was placed on the agenda by the JMPR Secretariat following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously considered toxicological data in addition to new published or unpublished toxicological studies and published epidemiological studies on cancer outcomes. Several study reports evaluated at previous JMPR meetings were not available to the present Meeting, as they were not submitted in the sponsor's dossier; for these studies, the evaluations in this report were summarized from the 1993 JMPR monograph without further review.

All critical unpublished studies contained statements of compliance with good laboratory practice (GLP), unless otherwise specified. The studies on human volunteers were conducted in accordance with the principles expressed in the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

Following oral administration to rats, diazinon was almost completely absorbed and rapidly eliminated, mainly in the urine. There was no evidence of accumulation.

Diazinon is metabolized by P 450 to diazoxon, the active metabolite. The main degradative pathway includes the oxidase/hydrolase-mediated cleavage of the ester bond, leading to the pyrimidinol derivative 2-isopropyl-6-methyl-4( 1 H$)$-pyrimidinone, which is further oxidized to more polar metabolites.

## Toxicological data

The oral median lethal dose $\left(\mathrm{LD}_{50}\right)$ for diazinon in rats ranged from 300 to greater than $2150 \mathrm{mg} / \mathrm{kg}$ bw, whereas the dermal $\mathrm{LD}_{50}$ was greater than $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The inhalation median lethal concentration ( $\mathrm{LC}_{50}$ ) was $3.1 \mathrm{mg} / \mathrm{L}$ in rats. Diazinon produced mild skin and eye irritation in rabbits. It caused skin sensitization in the guinea-pig Magnusson and Kligman maximization test.

The most sensitive end-point observed in all species given single and repeated doses of diazinon was inhibition of cholinesterase activity. Brain acetylcholinesterase activity was generally decreased at doses higher than those that inhibited erythrocyte acetylcholinesterase activity. Clinical signs of cholinergic toxicity occurred at doses causing more than $50 \%$ inhibition of brain acetylcholinesterase activity. Female rats were more sensitive than male rats.

Many repeated-dose toxicity studies are available. In both rats and dogs, no effects other than those related to cholinesterase inhibition have been observed at the lowest-observed-adverse-effect level (LOAEL); in general, effects observed at the highest doses can be considered secondary to the cholinergic toxicity. In these studies, NOAELs ranged from 0.02 to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day, and LOAELs ranged from 1 to $15 \mathrm{mg} / \mathrm{kg}$ bw per day, based on erythrocyte acetylcholinesterase inhibition (i.e. $>20 \%$ ), with brain acetylcholinesterase inhibition (i.e. $>10 \%$ ) generally appearing at the next higher dose and clinical cholinergic signs appearing at doses above $23 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 28-day acetylcholinesterase inhibition study, rats received diazinon by dietary administration at a concentration of $0,0.3,30,300$ or 3000 parts per million ( ppm ) (equal to $0,0.02$, $2.3,23$ and $213 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.02,2.4,23$ and $210 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.3 ppm (equal to $0.02 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 30 ppm (equal to $2.3 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a short-term toxicity study, rats were fed diazinon at a concentration of 0 or 2 ppm (equivalent to 0 and $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 7 days or at a concentration of 0 or 25 ppm (equivalent to 0 and $2.5 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 30 days. The NOAEL was 2 ppm (equivalent to $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day), based on inhibition of erythrocyte acetylcholinesterase activity at 25 ppm (equivalent to $2.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 3-month toxicity study, rats were given diets containing diazinon at a concentration of 0 , $0.5,5,250$ or 2500 ppm (equal to $0,0.03,0.3,15$ and $168 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.04$, $0.4,19$ and $212 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 5 ppm (equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at 250 ppm (equal to $15 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 3-month toxicity study, rats were fed diets containing diazinon at a concentration of $0,0.3,30,300$ or 3000 ppm (equal to $0,0.017,1.7,17$ and $177 \mathrm{mg} / \mathrm{kg}$ bw per day for males and 0 , $0.019,1.9,19$ and $196 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.3 ppm (equal to $0.017 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 30 ppm (equal to $1.7 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a third 3-month toxicity study, female rats were fed diets containing diazinon at a concentration of $0,5,10$ or 15 ppm (equivalent to $0,0.5,1$ and $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 92 days. In the second phase, female rats were fed diets containing diazinon at a concentration of $0,1,2,3$ or 4 ppm (equivalent to $0,0.1,0.2,0.3$ and $0.4 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 42 days. In the third phase, female rats were fed diets containing diazinon at a concentration of $0,0.1$, $0.5,1$ or 2 ppm (equivalent to $0,0.01,0.05,0.1$ and $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 35 days. The NOAEL in the first phase was 5 ppm (equivalent to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day), based on inhibition of erythrocyte acetylcholinesterase activity at 10 ppm (equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day) after dosing for 92 days. The NOAEL for females in the second and third phases were the highest tested doses of 4 ppm (equivalent to $0.4 \mathrm{mg} / \mathrm{kg}$ bw per day) and 2 ppm (equivalent to $0.2 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per day) after dosing for 42 and 35 days, respectively.

In a fourth 3-month toxicity study, rats were fed diets containing diazinon at a concentration of $0,5,125$ or 2000 ppm (equal to $0,0.3,7.8$ and $198 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.3,8.9$ and $247 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 5 ppm (equal to $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 125 ppm (equal to $7.8 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 90-day repeated-dose neurotoxicity study, rats were dosed in the diet at $0,25,125$ or 1000 ppm (equal to $0,1.7,8.4$ and $69.1 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1.8,9.3$ and $82.4 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). A NOAEL could not be identified, as erythrocyte acetylcholinesterase activity was inhibited at $1.7 \mathrm{mg} / \mathrm{kg}$ bw per day, the lowest dose tested.

In considering the NOAELs and LOAELs identified in the 28-day and 3-month (neuro)toxicity studies in rats measuring the inhibition of acetylcholinesterase activity, the Meeting concluded that the extent of acetylcholinesterase inhibition was not dependent on duration of dosing once steady state had been achieved (within 4 weeks). The overall NOAEL for the 28-day and 3month (neuro)toxicity studies in rats was 5 ppm , based on inhibition of erythrocyte acetylcholinesterase activity at the overall LOAEL of 10 ppm . In studies where feed consumption data were used to calculate test substance intake, 5 ppm was equal to $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day. These substance intake data are considered to be more accurate than those calculated using a default conversion factor, in which the NOAEL of 5 ppm is equivalent to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 90-day toxicity study, dogs were given diets containing diazinon at a concentration of 0 , $0.1,0.5,150$ or 300 ppm (equal to $0,0.0034,0.020,5.9$ and $10.9 \mathrm{mg} / \mathrm{kg}$ bw per day for males and 0 , $0.0037,0.021,5.6$ and $11.6 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.5 ppm (equal to $0.020 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at a dietary concentration of 150 ppm (equal to $5.6 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 90 -day toxicity study, dogs were given diazinon at $0,0.3,3$ or $10 \mathrm{mg} / \mathrm{kg}$ bw per day by gelatine capsule. The NOAEL was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at $3 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 1-year toxicity study in dogs given diazinon in the diet at a concentration of $0,0.1,0.5$, 150 or 300 ppm (equal to $0,0.0032,0.015,4.7$ and $7.7 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.0037$, $0.020,4.5$ and $9.1 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 0.5 ppm (equal to $0.015 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte (males and females) and brain (females only) acetylcholinesterase activities at 150 ppm (equal to $4.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL for the 90-day and 1-year toxicity studies in dogs was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte and brain acetylcholinesterase activities at $3 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a pre-GLP carcinogenicity study in mice that was considered adequate to evaluate carcinogenicity but not chronic toxicity, diazinon was administered at a dietary concentration of 0 , 100 or 200 ppm (equivalent to 0,15 and $30 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) over 103 weeks. No treatment-related tumours were observed.

In another pre-GLP carcinogenicity study in mice, diazinon was administered at a dietary concentration of $0,100,200,300$ (males) or 400 (females) ppm (equal to $0,16,31 \mathrm{and} 46 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per day for males and $0,22,43$ and $86 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively) for 104 weeks. Cholinesterase activity was not measured in this study. The NOAEL for chronic toxicity was 200 ppm (equal to $31 \mathrm{mg} / \mathrm{kg}$ bw per day), based on depression of body weight and lower feed consumption at 300 ppm (equal to $46 \mathrm{mg} / \mathrm{kg}$ bw per day). No treatment-related tumours were observed.

In a pre-GLP carcinogenicity study in rats that was considered adequate to evaluate carcinogenicity but not chronic toxicity, diazinon was administered at a dietary concentration of 0 , 400 or 800 ppm (equivalent to 0,20 and $40 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) over 103 weeks. No treatment-related tumours were observed.

In a chronic toxicity study, rats received diazinon in the diet at a concentration of 0 (untreated and vehicle controls), $0.1,1.5,125$ or 250 ppm (equal to $0,0.004,0.06,5$ and $10 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.005,0.07,6$ and $12 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively) for 98/99 weeks. The NOAEL was 1.5 ppm (equal to $0.06 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at 125 ppm (equal to $5 \mathrm{mg} / \mathrm{kg}$ bw per day). From the available data, there was no evidence of a tumorigenic response; however, the group size $(N=20)$ was too small to allow a conclusion to be reached on carcinogenicity.

In a combined chronic toxicity and carcinogenicity study in rats, diazinon was fed in the diet at concentrations adjusted to achieve target concentrations of $0,0.025,0.1,1.5$ and $22.5 \mathrm{mg} / \mathrm{kg}$ bw per day for 104 weeks. The NOAEL for long-term toxicity was $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte acetylcholinesterase activity at $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day. No treatment-related tumours were observed.

The overall NOAEL for chronic toxicity in rats was $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte acetylcholinesterase activity at $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that diazinon is not carcinogenic in mice or rats.
Given the similarity of the sensitivities of mammalian species, an overall NOAEL in all studies of repeated-dose (neuro)toxicity in rats and dogs could be identified. The overall NOAEL was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, on the basis of inhibition of acetylcholinesterase activity in erythrocytes at $1 \mathrm{mg} / \mathrm{kg}$ bw per day.

In studies submitted by the sponsors, diazinon was tested for genotoxicity in an adequate range of assays, both in vitro and in vivo. In addition, many studies with diazinon were described in the published literature, but most of these were considered by the Meeting as inappropriate to evaluate the genotoxicity of diazinon, as they had major deficiencies in study design or reliability (e.g. lack of statistical analysis, testing of mixtures of diazinon with other chemicals and similarity between negative and positive control values). Overall, these studies provided no convincing evidence of genotoxic effects.

The Meeting concluded that diazinon is unlikely to be genotoxic.
In the multigeneration and developmental toxicity studies, cholinesterase activity was not measured.

In a two-generation study on reproductive toxicity, rats received diazinon in the diet at a concentration of $0,10,100$ or 500 ppm over the course of two generations ( $\mathrm{F}_{0}$ and $\mathrm{F}_{1}$ ). Mean diazinon intakes for the $F_{0}$ generation during the premating period were $0,0.77,7.48$ and $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.77,7.48$ and $40.26 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively. The NOAEL for reproductive effects was 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day), based on prolonged gestation duration, decrease in the number of pregnancies, and reduced fertility and mating indices at 500 ppm (equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for parental effects was 10 ppm (equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day), based on reduced parental body weight gain at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for offspring toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on reduced viability of pups and pup weights at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day).

In another two-generation study on reproductive toxicity, rats received diazinon in the diet at a concentration of $0,0.1,1.0$ or $10 \mathrm{mg} / \mathrm{kg}$ (equivalent to $0,0.0067,0.067$ and $0.67 \mathrm{mg} / \mathrm{kg}$ bw per day, assuming concentrations are in $\mathrm{mg} / \mathrm{kg}$ feed or ppm ) over the course of two generations ( $\mathrm{F}_{0}$ and $\mathrm{F}_{1}$ ). A rationale for the dose selection was not provided. There were no treatment-related effects observed in $F_{0}$ or $F_{1}$ parental animals or pups. The NOAEL for reproductive, parental and offspring toxicity was 10 ppm (equivalent to $0.67 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

In a range of studies on estrogenic and androgenic activities, no estrogenic, androgenic or antiandrogenic activity was observed at concentrations relevant to human exposure via the diet.

Overall NOAELs from the multigeneration studies in rats were identified. The overall NOAEL for reproductive effects was 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 500 ppm (equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for parental toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for offspring toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a study of developmental toxicity evaluated by the 1993 JMPR, rats were administered diazinon via gavage at a dose of $0,15,50$ or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. A marked decrease in maternal feed consumption correlating with weight loss at the beginning of the treatment period and a slightly higher incidence of incomplete ossification at different sites in the fetuses were observed at 100 $\mathrm{mg} / \mathrm{kg}$ bw per day. As limited information was available from the previous JMPR monograph, the Meeting was unable to identify a NOAEL for this study.

In a study of developmental toxicity, rats were administered diazinon via gavage at a dose of 0 , 10,20 or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was $20 \mathrm{mg} / \mathrm{kg}$ bw per day, based on body weight loss on gestation days $6-10$, reduced body weight/body weight gains throughout treatment and decreased feed consumption on gestation days $6-9$ at $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo/fetal toxicity was $20 \mathrm{mg} / \mathrm{kg}$ bw per day, based on an increased incidence of rudimentary 14 th ribs at $100 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a study of developmental toxicity, rabbits were dosed with diazinon via gavage at $0,7,25$ or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was $25 \mathrm{mg} / \mathrm{kg}$ bw per day, based on mortality, tremors, convulsions, hypoactivity, anorexia and reduced body weight gain observed at 100 $\mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo/fetal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

In another developmental toxicity study, diazinon was administered to pregnant rabbits by gavage at a dose level of $0,2.5,10$ or $40 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was 10 $\mathrm{mg} / \mathrm{kg}$ bw per day, based on clinical signs, decreased body weight and reduced feed consumption. The NOAEL for embryo/fetal toxicity was $10 \mathrm{mg} / \mathrm{kg}$ bw per day, based on decreased fetal weight at $40 \mathrm{mg} / \mathrm{kg}$ bw per day.

The overall NOAEL for maternal toxicity in developmental toxicity studies in rabbits was 25 $\mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $40 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall NOAEL for embryo/fetal toxicity was $10 \mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $40 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that diazinon is not teratogenic.
In a limited acute neurotoxicity study in which acetylcholinesterase activity was not measured, rats were dosed with diazinon at $0,100,300$ or $500 \mathrm{mg} / \mathrm{kg}$ bw by gavage. The NOAEL was 100 $\mathrm{mg} / \mathrm{kg}$ bw, based on systemic toxicity and clinical signs of neurotoxicity observed at 300 or 500 $\mathrm{mg} / \mathrm{kg}$ bw. In another acute toxicity study, rats were administered a single dose of diazinon by gavage at $0,2.5,150,300$ or $600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, on the basis of depressed erythrocyte acetylcholinesterase activity and behavioural changes at $150 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. In a third study, rats were administered a single dose of diazinon by gavage at 100,250 or $500 \mathrm{mg} / \mathrm{kg}$ bw for males or $0,0.05,0.12,0.25,2.5,25$ or $250 \mathrm{mg} / \mathrm{kg}$ bw for females. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg}$ bw, on the basis of inhibition of brain and erythrocyte acetylcholinesterase activities in females at $25 \mathrm{mg} / \mathrm{kg}$ bw.

In a study that investigated the time course of acute inhibition of acetylcholinesterase activity, rats were given a single dose of diazinon by gavage at $0,2.5,150,300$ or $600 \mathrm{mg} / \mathrm{kg}$ bw, and brain and blood samples were collected at 3,9 and 24 hours after dosing. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg}$ bw, based on inhibition of brain and erythrocyte acetylcholinesterase activities at $150 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. Inhibition was observed beginning at 3 hours post-dosing, with maximal inhibition at 9 hours postdosing.

The overall NOAEL in all studies of acute toxicity was $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, on the basis of inhibition of acetylcholinesterase activity in erythrocytes and in the brain at $25 \mathrm{mg} / \mathrm{kg}$ bw in rats of both sexes.

Three studies were performed on delayed neurotoxicity in the hen. Oral doses of diazinon technical ranging from 10 to $100 \mathrm{mg} / \mathrm{kg}$ bw were administered to hens. Inhibition of cholinesterase activity was observed from $20 \mathrm{mg} / \mathrm{kg}$ bw, but there was no evidence that diazinon caused acute delayed neurotoxicity in the hen.

No specific studies on immunotoxicity were submitted. A study in the open literature with intraperitoneal injection of diazinon in mice was not informative. The submitted repeated-dose toxicity studies do not indicate an immunotoxic potential for diazinon after oral exposure.

## Toxicological data on metabolites and/or degradates

No toxicological data were available on any metabolites of diazinon other than diazoxon, which is the active metabolite of diazinon. However, the Meeting concluded that none of the other metabolites would be of toxicological concern at the levels present in the diet.

## Human data

In a study of acute toxicity in male volunteers given ascending doses of diazinon (seven volunteers per group given $0.03,0.12,0.20$ or $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$; one volunteer given $0.30 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ), acetylcholinesterase activity was not inhibited in erythrocytes at $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, the second highest dose tested. The highest dose ( $0.30 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ) was not informative, as it was tested in a single volunteer only. Plasma cholinesterase activity was inhibited by more than $20 \%$ at doses above 0.12 $\mathrm{mg} / \mathrm{kg}$ bw.

Repeated-dose studies in four male volunteers given diazinon for 28-37 days showed that, although there was some inhibition of plasma cholinesterase activity at the highest tested dose of $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day (actual administered doses varied slightly, i.e. $0.03,0.027,0.022 / 0.027$ and $0.026 \mathrm{mg} / \mathrm{kg}$ bw per day), no inhibition of erythrocyte acetylcholinesterase activity was observed.

Diazinon was evaluated in four male volunteers who received diazinon in capsules at 0.025 $\mathrm{mg} / \mathrm{kg}$ bw per day for 37-43 days. There were no consistent treatment-related effects on erythrocyte acetylcholinesterase activity, blood chemistry or urine analysis. No clinical effects were reported. The NOAEL was $0.025 \mathrm{mg} / \mathrm{kg}$ bw per day, the only dose tested.

The overall NOAEL from repeated-dose studies in humans was $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day.
Several epidemiological studies on cancer outcomes following occupational exposure to diazinon were available. The review of these studies focused on the occurrence of three cancer types: NHL, leukaemia and lung cancer (see section 2.2). One prospective cohort study was available, the Agricultural Health Study (AHS), with a large sample size and detailed exposure assessment. Cohort studies are considered a powerful design, as recall bias is avoided. All other studies were casecontrol studies, usually retrospective, which are more prone to recall and selection biases.

There was no significant evidence of a positive association of NHL with diazinon exposure and no evidence of an exposure-response relationship in the AHS. In a large pooled case-control study, the unadjusted estimates showed a significant elevated risk of NHL (relative risk [RR] $=1.7 ; 95 \%$ confidence interval $[\mathrm{CI}]=1.2-2.5$ ) associated with ever versus never use of diazinon. However, these risks were attenuated and/or no longer significant when proxy respondents were excluded and analyses were mutually adjusted for other pesticides (malathion, fonofos). Although increasing risk across exposure duration categories was observed, which was suggestive of a duration-response pattern, confidence intervals were non-significant, wide and overlapping between categories. Two other studies reported elevated risks of NHL for ever versus never use of diazinon or high versus low
diazinon use, but confidence intervals were wide, reflecting uncertainty in the risk estimates, and chance could not be excluded as an explanation for the findings. Overall, there was no convincing evidence of a positive association between NHL and exposure to diazinon.

A significantly increased risk of leukaemia in the highest exposure category (> 38.8 lifetime days of diazinon exposure; $\mathrm{RR}=3.36 ; 95 \% \mathrm{CI}=1.08-10.49$ ) and a significant exposure-response relationship were observed in the AHS. Findings for intensity-weighted lifetime exposure days demonstrated a similar pattern, but did not reach significance. Two other studies reported nonsignificantly elevated risks of leukaemia for high versus low diazinon use and ever versus never use of diazinon, with a non-significant dose-response relationship observed using days of use per year. Overall, there is weak evidence of a positive association between leukaemia and exposure to diazinon from the AHS only. It is noted that the number of diazinon-exposed cases was low or not reported in all three available studies.

A significant $60 \%$ excess risk of lung cancer in the highest exposure category ( $>38.8$ lifetime days of diazinon exposure) and a significant trend across exposure categories were observed in the AHS. Findings for intensity-weighted lifetime exposure days demonstrated a similar pattern, but did not reach significance. A separate analysis of ever use of diazinon versus never use from the AHS found no evidence of elevated risk of lung cancer among spouses of farmers/pesticide applicators; however, there were only 15 exposed cases. One other study reported a non-significant elevated risk of lung cancer for ever versus never use of diazinon (based on 17 exposed cases). Overall, there is weak evidence of a positive association between lung cancer and exposure to diazinon from the AHS cohort study only.

In view of the lack of genotoxicity and the absence of carcinogenicity in mice and rats and considering the available epidemiological data from occupational exposure, the Meeting concluded that diazinon is unlikely to pose a carcinogenic risk to humans via exposure from the diet.

The Meeting concluded that the existing database on diazinon was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The Meeting identified inhibition of acetylcholinesterase activity as the most sensitive end-point after single or repeated doses of diazinon in all species. After considering all previously evaluated data and the new studies, the Meeting established an ADI of $0-0.003 \mathrm{mg} / \mathrm{kg}$ bw, based on the overall NOAEL of $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day from all repeated-dose toxicity studies, and using a safety factor of 100. This ADI was supported by the NOAEL of $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested, identified in repeated-dose studies that involved a limited number of male volunteers, with application of a safety factor of 10 .

In 2006, the Meeting established an ADI of $0-0.005 \mathrm{mg} / \mathrm{kg}$ bw, based on the highest NOAEL of $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day for inhibition of erythrocyte acetylcholinesterase activity at $1 \mathrm{mg} / \mathrm{kg}$ bw per day in a 92 -day repeated-dose toxicity study in rats and using a safety factor of 100 . In this study, the dietary concentrations of diazinon were converted to units of milligrams per kilogram body weight per day using a default conversion factor; the present Meeting considers this less reliable than the conversion using feed consumption data.

The Meeting reaffirmed the ARfD of $0.03 \mathrm{mg} / \mathrm{kg}$ bw established by the 2006 JMPR. This ARfD was based on the NOAEL of $2.5 \mathrm{mg} / \mathrm{kg}$ bw identified in studies of acute (neuro)toxicity in rats, and using a safety factor of 100 . This ARfD was supported by the NOAEL of $0.21 \mathrm{mg} / \mathrm{kg}$ bw, the highest dose tested, identified in the study in which a limited number of male volunteers were given a single dose of diazinon, with application of a safety factor of 10 .

A toxicological monograph was prepared.

## Levels relevant to risk assessment of diazinon

| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
| Mouse | Two-year study of carcinogenicity ${ }^{\mathrm{a}, \mathrm{b}}$ | Toxicity | 200 ppm , equal to 31 $\mathrm{mg} / \mathrm{kg}$ bw per day | 300 ppm , equal to 46 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 300 ppm , equal to 46 $\mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
| Rat | Acute (neuro)toxicity studies ${ }^{\mathrm{d}, \mathrm{e}}$ (acetylcholinesterase inhibition) | Toxicity | $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ | $25 \mathrm{mg} / \mathrm{kg}$ bw |
|  | Four-week or 3-month studies of (neuro)toxicity ${ }^{\text {a,e }}$ | Toxicity | 5 ppm , equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{f}}$ | 10 ppm , equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-year studies of toxicity and carcinogenicity ${ }^{\text {a,e }}$ | Toxicity <br> Carcinogenicity | $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{f}}$ <br> 800 ppm , equivalent to $40 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{c}}$ | $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-generation studies of reproductive toxicity ${ }^{\text {a,b, }}$ | Reproductive toxicity | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day | 500 ppm , equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Parental toxicity | 10 ppm , equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | 10 ppm , equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity study ${ }^{\text {b,d }}$ | Maternal toxicity <br> Embryo and fetal toxicity | $20 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $20 \mathrm{mg} / \mathrm{kg}$ bw per day | $100 \mathrm{mg} / \mathrm{kg}$ bw per day $100 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity studies ${ }^{\text {b,d,e }}$ | Maternal toxicity <br> Embryo and fetal toxicity | $25 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $10 \mathrm{mg} / \mathrm{kg}$ bw per day | $40 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $40 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Dog | Ninety-day and 1-year studies of toxicity ${ }^{\text {a,e }}$ | Toxicity | $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {f }}$ | $3 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rat, dog | Repeat-dose (neuro)toxicity studies ${ }^{\text {e }}$ | Toxicity | 5 ppm , equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day | 10 ppm , equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Human | Acute toxicity study ${ }^{\text {d }}$ | Toxicity | $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}^{\text {c }}$ | - |
|  | Four/five-week studies of toxicity ${ }^{\text {d,e }}$ | Toxicity | $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |

[^6]Estimate of acceptable daily intake (ADI)
$0-0.003 \mathrm{mg} / \mathrm{kg}$ bw

## Estimate of acute reference dose (ARfD)

$0.03 \mathrm{mg} / \mathrm{kg}$ bw

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

## Critical end-points for setting guidance values for exposure to diazinon

| Absorption, distribution, excretion and metabolism in mammals |  |
| :--- | :--- |
| Rate and extent of oral absorption | Nearly complete and rapid $(\sim 90 \%$ at $10 \mathrm{mg} / \mathrm{kg}$ bw within 24 h$)$ |
| Dermal absorption | No data |
| Distribution | Widely distributed at low concentrations |
| Potential for accumulation | No potential for accumulation |
| Rate and extent of excretion | Predominantly in urine ( $86-93 \%$ at $10 \mathrm{mg} / \mathrm{kg}$ bw within 24 h$)$ <br> Metabolism in animals |
| Rapidly degraded to diazoxon and subsequently mainly via <br> oxidase/hydrolase-mediated cleavage of the ester bond, and <br> further oxidation at the isopropyl substituent to yield hydroxy <br> pyrimidinols |  |
| Toxicologically significant compounds in <br> animals and plants | Parent compound and diazoxon |

Acute toxicity

| Rat, $\mathrm{LD}_{50}$, oral | 300 to $>2150 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Rat, $\mathrm{LD}_{50}$, dermal | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| Rat, $\mathrm{LC}_{50}$, inhalation | $3.1 \mathrm{mg} / \mathrm{L}$ |
| Rabbit, dermal irritation | Mildly irritating |
| Rabbit, ocular irritation | Mildly irritating |
| Guinea-pig, dermal sensitization | Sensitizing (Magnusson and Kligman maximization test) |
| Repeat-dose studies of (neuro)toxicity |  |
| Target/critical effect | Acetylcholinesterase inhibition |
| Overall oral NOAEL | $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day (rat, dog) |
| Lowest relevant dermal NOAEL | $3 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rat) |
| Lowest relevant inhalation NOAEC | $0.46 \mathrm{mg} / \mathrm{m}^{3}(21$ days; rat) |

Long-term studies of carcinogenicity
Carcinogenicity Not carcinogenic in mice or rats ${ }^{a}$
Genotoxicity
No evidence of genotoxicity by the oral route ${ }^{a}$

| Reproductive toxicity |  |  |
| :---: | :---: | :---: |
| Target/critical effect | Mortality, reduced parental body weig of pups and pup weights, prolonged ge decrease in number of pregnancies, and mating indices | reduced viability duration, ed fertility and |
| Lowest relevant parental NOAEL | $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Lowest relevant offspring NOAEL | $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Lowest relevant reproductive NOAEL | $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Developmental toxicity |  |  |
| Target/critical effect | Clinical signs, reduced maternal body consumption, and reduced fetal weight | and feed |
| Lowest relevant maternal NOAEL | $25 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) |  |
| Lowest relevant embryo/fetal NOAEL | $10 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) |  |
| Neurotoxicity ${ }^{\text {b }}$ |  |  |
| Acute neurotoxicity NOAEL | $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ (acetylcholinesterase inh | rat) |
| Developmental neurotoxicity NOAEL | No data |  |
| Acute delayed neurotoxicity | No evidence (hens) |  |
| Human data | Acetylcholinesterase inhibition: |  |
|  | Acute toxicity NOAEL: $0.21 \mathrm{mg} / \mathrm{kg}$ bw, | st dose tested |
|  | Subchronic toxicity NOAEL: 0.03 mg / dose tested ( $4 / 5$ weeks) | per day, highest |
| ${ }^{\text {a }}$ Unlikely to pose a carcinogenic risk to humans via exposure from the diet. <br> ${ }^{\mathrm{b}}$ Ninety-day neurotoxicity study in rats is covered by the overall NOAEL for repeated-dose studies of |  |  |
| Summary |  |  |
| Value | Study | Safety factor |
| ADI $\quad 0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ | Repeated-dose toxicity studies (rat, dog) | 100 |
| ARfD $\quad 0.03 \mathrm{mg} / \mathrm{kg}$ bw | Acute (neuro)toxicity studies (rat) | 100 |

## DIETARY RISK ASSESSMENT

## Long-term dietary exposure

The ADI for diazinon is $0-0.003 \mathrm{mg} / \mathrm{kg}$ bw. The international estimated daily intakes (IEDIs) for diazinon were estimated for the 17 Global Environment Monitoring System - Food Contamination Monitoring and Assessment Programme (GEMS/Food) cluster diets using the supervised trials median residue (STMR) or STMR in a processed commodity (STMR-P) values estimated by the 1996 (animal commodities), 1999 (pome fruit, cabbage head) and 2006 (cranberries) JMPRs. An STMR value for tomato was estimated using the data reported in the 1993 JMPR evaluation monograph. For all other commodities, the maximum residue limits (MRLs) were used, as STMR values were not available. The results are shown in Annex 3. The IEDI ranged from $3 \%$ to $50 \%$ of the
maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of diazinon from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The ARfD for diazinon is $0.03 \mathrm{mg} / \mathrm{kg}$ bw. The international estimate of short-term dietary intake (IESTI) was calculated. The calculation employed highest residue (HR) values where these could be identified in the relevant JMPR reports; otherwise, the MRL was used. In the case of meat, the Meeting noted that residues in fat are approximately 15 times higher than those in muscle and used the MRL value of $2 \mathrm{mg} / \mathrm{kg}$ for fat and $0.1333 \mathrm{mg} / \mathrm{kg}$ for muscle. The results are shown in Annex 4. The IESTI represented a maximum of $100 \%$ of the ARfD for both children and the general population. The Meeting concluded that the short-term dietary exposure to diazinon residues from uses considered by JMPR was unlikely to present a public health concern.

### 3.2 GLYPHOSATE (158)

## TOXICOLOGY

Glyphosate is the ISO-approved common name for $N$-(phosphonomethyl)glycine (IUPAC), with CAS number 1071-83-6. It is a broad-spectrum systemic herbicide.

Glyphosate was previously evaluated by JMPR for toxicology in 1986, 1997 (evaluation of the metabolite aminomethylphosphonic acid, or AMPA), 2004 and 2011 (evaluation of new plant metabolites in genetically modified maize and soya beans).

Glyphosate was last re-evaluated for toxicology within the periodic review programme of CCPR in 2004. The compound was reviewed by the present Meeting following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously considered toxicological data in addition to new published or unpublished toxicological studies and published epidemiological studies on cancer outcomes. The evaluation of the biochemical aspects and systemic toxicity of glyphosate was based on previous JMPR evaluations, updated as necessary with additional information. The particular focus of the current meeting was on genotoxicity, carcinogenicity, reproductive and developmental toxicity and epidemiological studies on cancer outcomes. The scope was restricted to the active ingredient.

All critical unpublished studies contained statements of compliance with GLP, unless otherwise specified. The studies on human volunteers were conducted in accordance with the principles expressed in the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

In studies with radiolabelled glyphosate in rats, glyphosate was rapidly absorbed from the gastrointestinal tract following oral intake, but only to a limited extent (about 20-30\%). Elimination was fast and virtually complete within $72-168$ hours, with the majority being excreted during the first 48 hours. Most of the excretion occurred in faeces, largely as unabsorbed dose, and in the urine. Biliary excretion of glyphosate was negligible. Less than $1 \%$ of the administered dose was retained in tissues 168 hours post-administration. Highest residues were detected in bone, followed by kidney and liver. This pattern of absorption, distribution and elimination was independent of dose, treatment regimen and sex of the test animals. Peak plasma concentrations of radiolabel were observed at 6 and 2 hours after administration in male and female rats, respectively. The estimated half-life for wholebody elimination of the radiolabel was about 5.9-8.3 hours.

There was very little biotransformation of glyphosate; the only metabolite, AMPA, accounted for $0.2-0.7 \%$ of the administered dose in excreta; the rest was unchanged glyphosate.

## Toxicological data

Glyphosate has low acute oral toxicity in mice $\left(\mathrm{LD}_{50}>2000\right.$ to $>10000 \mathrm{mg} / \mathrm{kg}$ bw; no lethality at $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ) and rats $\left(\mathrm{LD}_{50} 5600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}\right)$, low acute dermal toxicity in rats $\left(\mathrm{LD}_{50}>2000\right.$ $\mathrm{mg} / \mathrm{kg} \mathrm{bw})$ and rabbits $\left(\mathrm{LD}_{50}>5000 \mathrm{mg} / \mathrm{kg}\right.$ bw$)$, and low acute inhalation toxicity in rats $\left(\mathrm{LC}_{50}>5.48\right.$ $\mathrm{mg} / \mathrm{L})$. Glyphosate was not irritating to the skin of rabbits. Glyphosate produced moderate to severe eye irritation in rabbits, with irreversible corneal opacity in one study as a consequence of the low pH of the test material in solution. Glyphosate was not sensitizing in guinea-pigs or mice as determined by the Magnusson and Kligman maximization test, the Buehler test and the local lymph node assay.

In short-term studies of toxicity in different species, the most notable effects were clinical signs related to gastrointestinal irritation, decreased body weight, salivary gland changes (hypertrophy and increase in basophilia of cytoplasm of acinar cells), histological findings in the caecum and hepatotoxicity.

In short-term studies in mice, reduced body weight was seen at a dietary concentration of 50000 ppm (equal to $9710 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for decreased body weight was 10000 ppm (equal to $1221 \mathrm{mg} / \mathrm{kg}$ bw per day). Effects on the salivary glands were observed in mice in only one study out of four at 6250 ppm (equal to $1065 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for the salivary gland effects in mice was 3125 ppm (equal to $507 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL in shortterm studies in mice was 3125 ppm (equal to $507 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 6250 ppm (equal to $1065 \mathrm{mg} / \mathrm{kg}$ bw per day).

In 90-day toxicity studies in rats, common findings included soft faeces, diarrhoea, reduced body weight gain and decreased food utilization at dietary concentrations of 20000 ppm (equal to $1262.1 \mathrm{mg} / \mathrm{kg}$ bw per day) and above. The lowest NOAEL was $371.9 \mathrm{mg} / \mathrm{kg}$ bw per day. A decrease in urine pH was frequently noted owing to the acidic nature of the compound and excretion as glyphosate in the urine. In two 90-day dietary toxicity studies, an increase in caecum weight (at 10000 ppm , equal to $569 \mathrm{mg} / \mathrm{kg}$ bw per day) and histological findings in the caecum (at 50000 ppm , equal to $3706 \mathrm{mg} / \mathrm{kg}$ bw per day) were observed. In rats, effects on the salivary gland were seen in two out of seven 90-day studies starting at 12500 ppm (equal to $811 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAELs for effects on the salivary gland were 300 and $410 \mathrm{mg} / \mathrm{kg}$ bw per day. The overall NOAEL in short-term studies in rats was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall LOAEL was 10000 ppm (equal to $569 \mathrm{mg} / \mathrm{kg}$ bw per day).

In four 90-day toxicity studies in dogs, the most notable effects were loose stools, decreased body weight and reduced feed consumption. In one study, there were no treatment-related effects at doses up to 40000 ppm (equal to $1015 \mathrm{mg} / \mathrm{kg}$ bw per day). The lowest NOAEL and LOAEL were $300 \mathrm{mg} / \mathrm{kg}$ bw per day and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively.

Seven 1-year toxicity studies in dogs are available. In one study, changes in faeces were observed at $100 \mathrm{mg} / \mathrm{kg}$ bw per day and above. The NOAEL was $30 \mathrm{mg} / \mathrm{kg}$ bw per day. However, these results were not reproduced in four other studies with administration via capsules at 300 or 500 $\mathrm{mg} / \mathrm{kg}$ bw per day. In the remaining six studies, the NOAELs ranged from 8000 ppm (equal to 182 $\mathrm{mg} / \mathrm{kg}$ bw per day) to $500 \mathrm{mg} / \mathrm{kg}$ bw per day, and the LOAELs ranged from 30000 ppm (equal to 926 $\mathrm{mg} / \mathrm{kg}$ bw per day) to $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

The overall NOAEL in the 90-day and 1-year toxicity studies in dogs was 15000 ppm (equal to $448 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 30000 ppm (equal to $926 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting compiled the tumour incidence data for all relevant mouse and rat studies in order to undertake statistical analysis and investigate any potential pattern of occurrence across studies. In addition, incidences of tumours of lymphatic tissues were summarized, as these were identified as possible targets of relevance from the review of epidemiological cancer studies. However, the Meeting recognized that the relationship between tumours of lymphatic tissues in rodents and humans has not been clearly established.

Nine carcinogenicity studies in mice were available. Two studies were considered to be of insufficient quality to be included in the assessment. Effects such as loose stools, reduced body weights and decreased feed consumption were noted in most of the studies. The overall NOAEL for systemic toxicity in mice was 1600 ppm (equal to $153 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 8000 ppm (equal to $787 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that there is equivocal evidence of induction of lymphomas in male mice in three out of seven studies and in female mice in one out of seven studies at high doses (500040000 ppm , equal to $814-4348 \mathrm{mg} / \mathrm{kg}$ bw per day). The Meeting also noted that in the other three studies in which even higher doses (up to 50000 ppm , equal to $7470 \mathrm{mg} / \mathrm{kg}$ bw per day) had been used, no effect was observed.

The Meeting concluded that there is some indication, by a trend test, and not by pairwise comparison, of induction of kidney adenomas in male mice in four out of seven studies. The Meeting noted that the increases were marginal and occurred at the highest dose only and that other studies that used appreciably higher doses did not find any excess. However, the Meeting noted that kidney adenomas are uncommon in male mice.

Eleven combined chronic toxicity and carcinogenicity studies in rats were available. One study was considered to be inadequate for carcinogenicity assessment due to its exposure duration (12 months). Toxicities variously reported in some of these studies included increased incidences of clinical signs, reduced body weights, degenerative lens changes (cataracts) in males, microscopic findings in the salivary gland, increased incidence of basophilia of parotid acinar cells, and microscopic findings in liver, prostate and kidneys. The overall NOAEL for systemic toxicity in rats was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall LOAEL was $300 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting discussed the increased incidence of a variety of tumours observed in one or, in one case, two of the 10 studies in rats. The Meeting concluded that these findings were incidental, based on the following considerations:

- interstitial cell tumours of the testes: occurred in only one study; and other studies that used appreciably higher doses did not find any excess;
- pancreatic islet cell adenoma: occurred in only one study in males only; other studies that used appreciably higher doses did not find any excess; there was no dose-response relationship; and the incidence in controls was unusually low (less than the lower bound of the historical control data); the Meeting also noted that there was a negative dose-response relationship in females;
- thyroid C-cell tumours: occurred in only one study; other studies that used appreciably higher doses did not find any excess; and these tumours are considered not to be relevant for humans;
- skin keratoma: occurred in two studies in males only; other studies that used appreciably higher doses did not find any excess; in one study, there was no dose-response relationship; and in the other study, only the test for trend was statistically significant, not the pairwise test at any dose;
- lymphoma (in spleen and kidney): no evidence of induction in any of the studies.

The Meeting concluded that there is no reliable evidence for treatment-related tumours in rats at doses up to 32000 ppm (equal to $1750 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that glyphosate is not carcinogenic in rats but could not exclude the possibility that it is carcinogenic in mice at very high doses.

Glyphosate and its formulation products have been extensively tested for genotoxic effects using a variety of tests in a wide range of organisms. While no mutational effects have been detected in bacterial test systems, DNA damage and chromosomal effects have commonly been seen in cell culture models and in organisms that are phylogenetically distant from humans. However, these effects have not been seen in vivo in orally treated mammalian models. The overall weight of evidence indicates that administration of glyphosate and its formulation products at doses as high as $2000 \mathrm{mg} / \mathrm{kg}$ bw by the oral route, the route most relevant to human dietary exposure, was not associated with genotoxic effects in an overwhelming majority of studies conducted in mammals, a model considered to be appropriate for assessing genotoxic risks to humans.

The Meeting concluded that glyphosate is unlikely to be genotoxic at anticipated dietary exposures.

Seven reproductive toxicity studies in rats were available. No evidence of reproductive toxicity was observed at doses up to 30000 ppm (equal to $1983 \mathrm{mg} / \mathrm{kg}$ bw per day). In one study, an
increased incidence of histopathological findings in the parotid (males) and submaxillary salivary glands in females was observed in both generations at 10000 ppm (equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL was 3000 ppm (equal to $197 \mathrm{mg} / \mathrm{kg}$ bw per day). In a separate study, an increased incidence of loose stools and caecum distension was observed in both generations at 30000 ppm (equal to $2150 \mathrm{mg} / \mathrm{kg}$ bw per day), and the NOAEL was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day). Slight reductions in pup weight or weight gain were observed in most studies, but were confined to very high, parentally toxic dose levels. In addition, a significant delay in sexual maturation in male pups ( $\mathrm{F}_{1}$ ) was seen at 15000 ppm (equal to $1063 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for parental toxicity was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 10000 ppm (equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for offspring toxicity was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 10000 ppm (equal to $985 \mathrm{mg} / \mathrm{kg}$ bw per day).

No evidence of teratogenicity was observed in four developmental toxicity studies in rats at doses up to $3500 \mathrm{mg} / \mathrm{kg}$ bw per day. There was some variation in the extent of toxicity observed in the four studies. The lowest NOAEL for maternal toxicity was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, based on loose stools and reduced body weights seen at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day. The lowest NOAEL for embryo and fetal toxicity was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, based on delayed ossification and an increased incidence of fetuses with skeletal anomalies observed at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

Seven developmental toxicity studies in the rabbit were available. Maternal toxicity was primarily manifested as an increased incidence of soft stools and diarrhoea at doses of $175 \mathrm{mg} / \mathrm{kg}$ bw per day and above. The overall NOAEL for maternal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day. In three studies, the occurrences of a variety of low-incidence fetal effects (e.g. cardiac malformation, absent kidney) were slightly increased at higher dose levels. These increases are considered secondary to maternal toxicity. The overall NOAEL for embryo and fetal toxicity was $250 \mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $450 \mathrm{mg} / \mathrm{kg}$ bw per day. The Meeting considered that these effects were secondary to local irritation from unabsorbed glyphosate in the colon administered by gavage dosing and concluded that they were not relevant for establishing health-based guidance values.

The Meeting concluded that glyphosate is not teratogenic.
Glyphosate was tested in a range of validated in vivo and in vitro assays for its potential to interact with the endocrine system. The studies that the Meeting considered adequate for the evaluation clearly demonstrate that there is no interaction with estrogen or androgen receptor pathways or thyroid pathways.

There was no evidence of neurotoxicity in an acute neurotoxicity study in rats at doses up to $2000 \mathrm{mg} / \mathrm{kg}$ bw. The NOAEL for systemic toxicity was $1000 \mathrm{mg} / \mathrm{kg}$ bw, based on a single death and general signs of toxicity at $2000 \mathrm{mg} / \mathrm{kg}$ bw. In a 90 -day neurotoxicity study in rats, no evidence of neurotoxicity or systemic toxicity was seen at doses up to 20000 ppm (equal to $1546.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

No evidence of immunotoxicity was seen in a 28 -day dietary study in female mice at doses up to 5000 ppm (equal to $1448 \mathrm{mg} / \mathrm{kg}$ bw per day).

Effects on the salivary glands were observed in several repeated-dose toxicity studies in rats. The pH of glyphosate in solution is low, and it has been shown that exposure to organic acids can cause such changes in salivary glands. Therefore, the changes are likely secondary to the effects caused by the pH of the test compound in solution.

In many of the long-term repeated-dose studies reviewed, glyphosate was reported to have an impact on the gastrointestinal tract at high doses. Although this is not uncommon with high-dose chemical substance administration, this was investigated further, as glyphosate is known to be poorly absorbed in mammalian models, and alterations in gut microbiota profiles, specifically reductions in the beneficial microbiota and increases in pathogenic bacteria, are known to have impacts on carcinogenesis. There is evidence from livestock species that pathogenic bacteria are more resistant to glyphosate, whereas beneficial microbiota are more sensitive, and thus more vulnerable.

This is an emerging area of scientific investigation. The extent to which glyphosate adversely affects the normal functioning of the microbiota in the human gastrointestinal tract or the gastrointestinal tract of mammalian models is unclear. However, it is unlikely, given the available information on minimum inhibitory concentration (MIC) values, that this would occur from glyphosate residues in the diet.

## Toxicological data on metabolites and/or degradates

AMPA is the only identified metabolite found in the urine and faeces of orally treated rats. AMPA was of low acute oral and dermal toxicity in rats $\left(\mathrm{LD}_{50}>5000\right.$ and $>2000 \mathrm{mg} / \mathrm{kg}$ bw, respectively) and was not sensitizing in guinea-pigs, as determined by the Magnusson and Kligman maximization test. In a 90-day study of toxicity in rats, the NOAEL was $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested. AMPA administered orally in mammalian test systems showed no evidence of genotoxicity. Largely negative results were seen in studies in vitro. The Meeting concluded that AMPA is unlikely to be genotoxic in vivo by the oral route. In a study of developmental toxicity in rats, no evidence for embryo or fetal toxicity was observed; the NOAEL for maternal and embryo/fetal toxicity was 1000 $\mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

Following single gavage administration of radiolabelled $N$-acetyl-glyphosate, a plant-specific metabolite, at $15 \mathrm{mg} / \mathrm{kg}$ bw in rats, about $66.1 \%$ of the administered dose was excreted in urine ( $61.3 \%$ within 12 hours post-dosing), $26.4 \%$ in faeces ( $25.8 \%$ within 48 hours post-dosing), $2.79 \%$ in cage wash and wipe, and $0.23 \%$ in residual carcass. Radioactivity was eliminated rapidly from blood and plasma, with half-life values of 20.1 and 15.6 hours, respectively. Unchanged $\left[{ }^{14} \mathrm{C}\right] \mathrm{N}$-acetylglyphosate recovered in urine and faeces represented over $99 \%$ of the administered radioactivity. Glyphosate, a metabolite of $N$-acetyl-glyphosate, was detected in faeces and represented less than $1 \%$ of the total radioactivity.

The acute oral toxicity $\mathrm{LD}_{50}$ of $N$-acetyl-glyphosate in rats is greater than $5000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, expressed as the free acid. In a 90-day toxicity study in rats, the NOAEL was 18000 ppm (equal to $1157 \mathrm{mg} / \mathrm{kg}$ bw per day).
$N$-Acetyl-glyphosate was tested for genotoxicity in vitro and in vivo in an adequate range of assays; it was not found to be genotoxic in mammalian or microbial test systems.

The Meeting concluded that $N$-acetyl-glyphosate is unlikely to be genotoxic.
$N$-Acetyl-AMPA, another plant-specific metabolite, was of low acute oral toxicity; the $\mathrm{LD}_{50}$ was greater than $5000 \mathrm{mg} / \mathrm{kg}$ bw in rats.
$N$-Acetyl-AMPA was tested for genotoxicity in vitro and in vivo in an adequate range of assays; it was not found to be genotoxic in mammalian or microbial test systems.

The Meeting concluded that $N$-acetyl-AMPA is unlikely to be genotoxic.

## Human data

Routine medical surveillance of workers in production and formulation plants revealed no adverse health effects attributable to glyphosate. In operators applying glyphosate products, cases of eye, skin and/or respiratory tract irritation have been reported. Acute intoxication was reported in humans after accidental or intentional ingestion of concentrated glyphosate formulations, resulting in gastrointestinal, cardiovascular, pulmonary and renal effects and, occasionally, death. The acute toxicity of glyphosate formulations was likely caused by the surfactant in these products.

Several epidemiological studies on cancer outcomes following occupational exposure to glyphosate were available. The evaluation of these studies focused on the occurrence of NHL, as outlined in section 2.2. One meta-analysis and one prospective cohort study, the AHS, with a large sample size and detailed exposure assessment, were available. Cohort studies are considered a
powerful design, as recall bias is avoided. All other studies were case-control studies, usually retrospective, which are more prone to recall and selection biases.

The AHS cohort study found no evidence of a positive association of NHL with glyphosate exposure or an exposure-response relationship. Elevated risks were reported in various case-control studies. A significant elevated risk of NHL associated with ever versus never use of glyphosate (odds ratio $[\mathrm{OR}]=2.1 ; 95 \% \mathrm{CI}=1.1-4.0$ ) was reported. Ever use of glyphosate was not associated with risk of NHL in the cross-Canada case-control study of pesticides and health, but when analysing days of use per year, there was a significant elevated risk in the highest usage category ( $\mathrm{OR}=2.12$; $95 \%$ $\mathrm{CI}=1.20-3.73$; for $>2$ days/year glyphosate use). There was, however, no indication of an exposure-response relationship across exposure usage categories. In another case-control study, a significant increased risk of NHL associated with ever use ( $\mathrm{OR}=2.02 ; 95 \% \mathrm{CI}=1.10-3.71$ ) as well as the highest usage category ( $\mathrm{OR}=2.36 ; 95 \% \mathrm{CI}=1.04-5.37$; for greater than 10 days/year glyphosate use) was observed, with some suggestion of an exposure-response gradient. Two smaller case-control studies with few exposed cases and limited statistical power reported a non-significant elevated risk and no association, respectively, for risk of NHL and ever use of glyphosate. The metaanalysis, including the AHS, found a significant $50 \%$ excess risk ratio for ever versus never use of glyphosate.

Overall, there is some evidence of a positive association between glyphosate exposure and risk of NHL from the case-control studies and the overall meta-analysis. However, it is notable that the AHS, which is the only cohort study and is large and of high quality, found no evidence of association at any exposure level.

In view of the absence of carcinogenic potential in rodents at human-relevant doses and the absence of genotoxicity by the oral route in mammals, and considering the epidemiological evidence from occupational exposures, the Meeting concluded that glyphosate is unlikely to pose a carcinogenic risk to humans via exposure from the diet.

The Meeting concluded that the existing database on glyphosate was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The Meeting reaffirmed the group ADI for the sum of glyphosate, AMPA, $N$-acetyl-glyphosate and $N$-acetyl-AMPA of $0-1 \mathrm{mg} / \mathrm{kg}$ bw on the basis of the NOAEL of $100 \mathrm{mg} / \mathrm{kg}$ bw per day for effects on the salivary gland in a long-term study of toxicity and carcinogenicity in rats and application of a safety factor of 100 . The Meeting noted that these effects may be secondary to local irritation due to the low pH of glyphosate in solution, but was unable to establish this unequivocally.

The Meeting concluded that it was not necessary to establish an ARfD for glyphosate, AMPA, $N$-acetyl-glyphosate and $N$-acetyl-AMPA in view of their low acute toxicity, the absence of relevant developmental toxicity in rats and rabbits that could have occurred as a consequence of acute exposure, and the absence of any other toxicological effect that would be elicited by a single dose.

A toxicological monograph was prepared.

Levels relevant to risk assessment of glyphosate

| Species | Study | Effect | NOAEL | LOAEL |
| :--- | :--- | :--- | :--- | :--- |
| Mouse | Eighteen- to 24-month | Toxicity | 1600 ppm, equal to | 8000 ppm, equal to |
|  | studies of toxicity and |  | $153 \mathrm{mg} / \mathrm{kg}$ bw per <br> carcinogenicity ${ }^{\text {a,b }}$ |  |
|  | day $^{\mathrm{c}}$ | $787 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per |  | day |


| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Carcinogenicity | The Meeting could not exclude the possibility that glyphosate is carcinogenic in mice at very high doses. |  |
| Rat | Acute neurotoxicity study ${ }^{\text {a }}$ | Neurotoxicity | $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}^{\text {c }}$ | - |
|  | Two-year studies of toxicity and carcinogenicity ${ }^{b}$ | Toxicity | $100 \mathrm{mg} / \mathrm{kg}$ bw per day | $300 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 32000 ppm , equal to $1750 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  | Two-generation studies of reproductive toxicity ${ }^{\text {a,b }}$ | Reproductive toxicity | 30000 ppm , equal to $1983 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  |  | Parental toxicity | 6000 ppm , equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day | 10000 ppm , equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | 6000 ppm , equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day | 10000 ppm , equal to $985 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity studies ${ }^{\text {b,d }}$ | Maternal toxicity | $300 \mathrm{mg} / \mathrm{kg}$ bw per day | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $300 \mathrm{mg} / \mathrm{kg}$ bw per day | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity studies ${ }^{\text {b,d }}$ | Maternal toxicity ${ }^{\text {e }}$ | $100 \mathrm{mg} / \mathrm{kg}$ bw per day | $175 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity ${ }^{\text {e }}$ | $250 \mathrm{mg} / \mathrm{kg}$ bw per day | $450 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Dog | Thirteen-week and 1-year studies of toxicity ${ }^{\text {b,f }}$ | Toxicity | 15000 ppm , equal to $448 \mathrm{mg} / \mathrm{kg}$ bw per day | 30000 ppm , equal to $926 \mathrm{mg} / \mathrm{kg}$ bw per day |
| AMPA |  |  |  |  |
| Rat | Thirteen-week study of toxicity ${ }^{\text {d }}$ | Toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  | Developmental toxicity study ${ }^{\text {d }}$ | Maternal toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  |  | Embryo and fetal toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |

${ }^{\text {a }}$ Dietary administration.
${ }^{\text {b }}$ Two or more studies combined.
${ }^{\text {c }}$ Highest dose tested.
${ }^{\text {d }}$ Gavage administration.
${ }^{\mathrm{e}}$ Secondary to local irritation of the colon.
${ }^{f}$ Capsule administration.

## Estimate of acceptable daily intake (ADI)

$0-1 \mathrm{mg} / \mathrm{kg}$ bw (for sum of glyphosate, $N$-acetyl-glyphosate, AMPA and $N$-acetyl-AMPA)

## Estimate of acute reference dose (ARfD)

## Unnecessary

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

## Critical end-points for setting guidance values for exposure to glyphosate

Absorption, distribution, excretion and metabolism in mammals
Rate and extent of oral absorption Rapidly, but only to a limited extent (about 20-30\%)
Dermal absorption
Distribution
Potential for accumulation
Rate and extent of excretion

Metabolism in animals
Toxicologically significant compounds in animals and plants

## About 1-3\%

Widely distributed (low levels occurring in all tissues)
No evidence of accumulation
Rapid and nearly complete in 48 h (about 20-30\% in urine and about $60-70 \%$ in faeces)

Very limited ( $<0.7 \%$ ), by hydrolysis leading to AMPA
Parent compound, AMPA, $N$-acetyl-glyphosate, $N$-acetylAMPA

Acute toxicity
Rat, $\mathrm{LD}_{50}$, oral $5600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$
Rat, $\mathrm{LD}_{50}$, dermal $\quad>2000 \mathrm{mg} / \mathrm{kg}$ bw
Rat, $\mathrm{LC}_{50}$, inhalation $\quad>5.48 \mathrm{mg} / \mathrm{L}$
Rabbit, dermal irritation Not irritating
Rabbit, ocular irritation Moderately to severely irritating
Guinea-pig, dermal sensitization Not sensitizing (Magnusson and Kligman test, Buehler test)
Mouse, dermal sensitization Not sensitizing (local lymph node assay)
Short-term studies of toxicity
Target/critical effect Clinical signs (loose stools, diarrhoea), liver, salivary glands and reduced body weights

Lowest relevant oral NOAEL
$300 \mathrm{mg} / \mathrm{kg}$ bw per day ( 90 days; rat)
Lowest relevant dermal NOAEL $\quad>5000 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rabbit)
Lowest relevant inhalation NOAEC
No data
Long-term studies of toxicity and carcinogenicity
Target/critical effect Reduced body weights, loose stools, liver (toxicity), salivary glands (organ weight, histology), eye (cataracts, lens fibre degeneration)

Lowest relevant NOAEL
$100 \mathrm{mg} / \mathrm{kg}$ bw per day (2 years; rat)
Carcinogenicity Not carcinogenic in rats; could not exclude possibility of carcinogenicity in mice at very high doses ${ }^{\text {a }}$

## Genotoxicity

No genotoxic potential via oral route in mammals ${ }^{\mathrm{a}}$

## Reproductive toxicity

Target/critical effect Reduced body weights and delayed development (absence of maternal toxicity)
Lowest relevant parental NOAEL
Lowest relevant offspring NOAEL
$417 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)
$417 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)

Lowest relevant reproductive NOAEL
Developmental toxicity

Target/critical effect
Lowest relevant maternal NOAEL
Lowest relevant embryo/fetal NOAEL
$1983 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)

Neurotoxicity
Acute neurotoxicity NOAEL
Subchronic neurotoxicity NOAEL
Developmental neurotoxicity NOAEL
Slight increase in malformations at maternally toxic doses
$100 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) ${ }^{\text {b }}$
$250 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) ${ }^{\text {b }}$

Other toxicological studies
Immunotoxicity
No immunotoxicity; NOAEL $1448 \mathrm{mg} / \mathrm{kg}$ bw per day, highest dose tested (28 days; mouse)
Studies on toxicologically relevant metabolites
Toxicological studies on AMPA, N -acetyl-glyphosate and N -acetyl-AMPA reveal the metabolites to be less toxic than the parent compound

Human data
Medical surveillance of workers in plants producing and formulating glyphosate did not reveal any adverse health effects. In operators applying glyphosate products, cases of eye, skin and/or respiratory irritation have been reported. Cases of acute intoxication have been observed after accidental or intentional ingestion of glyphosate formulation.

[^7]
## Summary

|  | Value | Study | Safety factor |
| :--- | :--- | :--- | :--- |
| ADI | $0-1 \mathrm{mg} / \mathrm{kg}$ bw | Two-year studies of toxicity (rat) | 100 |
| ARfD | Unnecessary | - | - |

## DIETARY RISK ASSESSMENT

## Long-term dietary exposure

The ADI for glyphosate is $0-1 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The IEDIs for glyphosate were estimated for the 17 GEMS/Food cluster diets using the STMR or STMR-P values estimated by JMPR. The results are
shown in Annex 3. The IEDI ranged from $0 \%$ to $1 \%$ of the maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of glyphosate from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The Meeting concluded that it was unnecessary to establish an ARfD for glyphosate, and therefore an IESTI for glyphosate was not calculated. The Meeting therefore concluded that short-term dietary exposure to glyphosate residues is unlikely to present a risk to consumers.

### 3.3 MALATHION (49)

## TOXICOLOGY

Malathion is the ISO-approved common name for $S$-1,2-bis(ethoxycarbonyl)ethyl $O, O$-dimethyl phosphorothioate (IUPAC), with the CAS number 121-75-5.

Malathion is a non-systemic organophosphorus insecticide whose mode of pesticidal action is the inhibition of cholinesterase activity. It is used to control insects on agricultural crops and stored commodities and for vector control.

The toxicity of malathion was evaluated by JMPR in 1963, 1965, 1966, 1997 and 2003. Malathion was listed in the periodic review programme of CCPR but was not yet scheduled for review. The compound was reviewed by the present Meeting following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously submitted toxicological data in addition to new published and unpublished toxicological studies and published epidemiological studies on cancer outcomes. All critical unpublished studies contained certificates of compliance with GLP, unless otherwise specified. Human volunteer studies were conducted according to the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

In a study conducted in rats using $\left[{ }^{14} \mathrm{C}\right]$ malathion, gastrointestinal absorption was at least $77 \%$ in males and $86 \%$ in females. The majority (up to $90 \%$ ) of radioactivity was excreted in urine within 24 hours. Less than $1 \%$ of radioactivity was detected in tissues, with the highest proportions in the liver, skin, fat and gastrointestinal tract. There was no evidence that malathion or its metabolites accumulated in any tissue.

Malathion is extensively metabolized via desulfuration, oxidation, hydrolysis, dealkylation and demethylation reactions. In particular, the oxidative desulfuration of malathion in the liver generates malaoxon, which is a more potent inhibitor of acetylcholinesterase compared with malathion. The major metabolites detected in rat urine ( $>80 \%$ of urinary radioactivity) were $\alpha$ - and $\beta$ monocarboxylic acids (MMCA) and the dicarboxylic acid (MDCA) of malathion. Other urinary metabolites include desmethyl malathion, $O, O$-dimethyl phosphorothioic acid, fumaric acid, 2mercaptosuccinic acid, $O, O$-dimethyl phosphorodithioic acid, monoethyl fumarate and malaoxon. Malaoxon was observed only in urine samples and accounted for less than $2 \%$ of total urinary radioactivity. Similar metabolites were detected in human studies.

Published in vitro studies have further investigated the metabolism of malathion. In human liver microsomes, the metabolism of malathion to malaoxon was catalysed by CYP1A2, CYP2B6 or CYP3A4, their respective contributions depending on the concentration of malathion. Isomalathion, a storage impurity, was a potent non-competitive inhibitor of hepatic carboxylesterase activity, important for the formation of MMCA by human liver microsomes.

Estimates of in vitro dermal absorption through human skin ranged from $1.44 \%$ to $8.74 \%$ and from $8 \%$ to $20.7 \%$. In a volunteer study, dermal absorption was $4.48 \%$ following a single application and $3.53 \%$ following a second application.

## Toxicological data

Consistent with other organophosphorus insecticides, the most sensitive toxicological effect following acute and repeated exposures to malathion is the inhibition of acetylcholinesterase activity in erythrocytes and brain. At higher doses, cholinergic signs become evident.

In rats, the oral $\mathrm{LD}_{50}$ ranged from 1539 to $8227 \mathrm{mg} / \mathrm{kg}$ bw, the dermal $\mathrm{LD}_{50}$ was greater than $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ and the inhalation $\mathrm{LC}_{50}$ was greater than $5.2 \mathrm{mg} / \mathrm{L}$. The dermal $\mathrm{LD}_{50}$ in rabbits was $8790 \mathrm{mg} / \mathrm{kg}$ bw. Malathion was slightly irritating to rabbit skin and eyes. In a Buehler test conducted in guinea-pigs, malathion did not cause skin sensitization, whereas malathion caused skin sensitization in the guinea-pig maximization test. Malathion was not sensitizing in the mouse local lymph node assay.

In a 14 -day range-finding study conducted in juvenile rats, which tested gavage malathion doses of $0,250,450$ and $600 \mathrm{mg} / \mathrm{kg}$ bw per day, salivation occurred at 450 and $600 \mathrm{mg} / \mathrm{kg}$ bw per day. In males, erythrocyte and brain acetylcholinesterase activities were reduced at every dose, whereas in females, erythrocyte and brain acetylcholinesterase activities were reduced at 450 and 600 $\mathrm{mg} / \mathrm{kg}$ bw per day.

In a 28 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000$ and 10000 ppm (equal to $0,9.2,46.1,457.5$ and $947.8 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,9.4,47.4,461.3$ and $910.1 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $46.1 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of erythrocyte and brain acetylcholinesterase activities at 5000 ppm (equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day). Nasal toxicity, consisting of goblet cell depletion and hyperplasia of the olfactory epithelium, was noted at the highest dose.

In a 30 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,50,100,500,10000$ and 20000 ppm (equal to $0,5.1,10.4,51.9,1036$ and $2008 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,5.7,11.6,57.6,1134$ and $2193 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 10000 ppm (equal to $1036 \mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL from these two 1-month repeated-dose toxicity studies in rats was 500 ppm (equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day), with an overall LOAEL of 5000 ppm (equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 90 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000,10000$ and 20000 ppm (equal to $0,7,34,340,680$ and $1390 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,8,39,384,784$ and $1597 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $34 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 5000 ppm (equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 90 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000$ and 10000 ppm (equal to $0,7.2,35.0,353.6$ and $733.8 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,7.5,35.9,363.1$ and $719.0 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 100 ppm (equal to $7.2 \mathrm{mg} / \mathrm{kg}$ bw per day) for goblet cell depletion at 500 ppm (equal to $35.0 \mathrm{mg} / \mathrm{kg}$ bw per day). This is considered to be an atypical result, as the effect is likely to have arisen through non-dietary exposure.

In a 13 -week neurotoxicity study in rats, which tested dietary malathion concentrations of 0 , 50,5000 and 20000 ppm (equal to $0,4,352$ and $1486 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,4,395$ and $1575 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 50 ppm (equal to $4 \mathrm{mg} / \mathrm{kg}$ bw per day), based on the inhibition of erythrocyte acetylcholinesterase activity at 5000 ppm (equal to 352 $\mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL for the 90 -day (neuro)toxicity studies in rats was 500 ppm (equal to 34 $\mathrm{mg} / \mathrm{kg}$ bw per day) for effects at 5000 ppm (equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 28 -day range-finding study in dogs in which malathion was administered orally in capsules at doses of $0,125,250$ and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, inhibition of erythrocyte acetylcholinesterase occurred at 250 and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, with deaths, cholinergic signs and reduced body weight and feed consumption occurring at the highest dose.

In a 12-month repeated-dose toxicity study in dogs in which malathion was administered orally in capsules at doses of $0,62.5,125$ and $250 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $125 \mathrm{mg} / \mathrm{kg}$ bw per day for reduced body weight and haematological changes at $250 \mathrm{mg} / \mathrm{kg}$ bw per day. Inhibition of erythrocyte acetylcholinesterase activity occurred at every dose but was of marginal toxicological significance in the absence of brain acetylcholinesterase inhibition.

In a 3-week repeated-dose dermal toxicity study in rabbits, which tested malathion doses of 0 , 50,300 and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $300 \mathrm{mg} / \mathrm{kg}$ bw per day for the inhibition of brain acetylcholinesterase activity at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 21-day repeated-dose dermal toxicity study in rabbits, which tested malathion doses of 0 , $75,100,150$ and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $150 \mathrm{mg} / \mathrm{kg}$ bw per day for the inhibition of brain acetylcholinesterase activity at $500 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 13 -week repeated-dose inhalational toxicity study in which rats were exposed whole body to an aerosol malathion concentration of $0,0.1,0.45$ or $2.0 \mathrm{mg} / \mathrm{L}$, a no-observed-adverse-effect concentration (NOAEC) was not determined, as laryngeal hyperplasia and degeneration and/or hyperplasia of the olfactory epithelium occurred at every concentration.

In an 18 -month pre-GLP study conducted in mice, which tested dietary malathion concentrations of 0,8000 and 16000 ppm (equivalent to 0,1200 and $2400 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), a NOAEL for chronic toxicity was not identified, because clinical signs during the second year of exposure and reduced body weight occurred at both doses. Although no treatmentrelated tumours were observed, this study was considered unreliable for assessing carcinogenicity because of the small number of concurrent control mice $(n=10)$ compared with the treated groups ( $n=50$ ).

In a second 18 -month study conducted in mice, which tested dietary malathion concentrations of $0,100,800,8000$ and 16000 ppm (equal to $0,17,143,1476$ and $2978 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,21,167,1707$ and $3448 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for chronic toxicity was 800 ppm (equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 8000 ppm (equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day). Increases in liver carcinomas in males at the low dose and second highest dose were not considered treatment related because of the lack of a dose-response relationship, the lack of corroboration in females and the fact that liver carcinomas are a common age-related tumour in this strain of mouse (B6C3F1). The NOAEL for carcinogenicity was 800 ppm (equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day) for an increased incidence of liver adenomas at 8000 ppm (equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day).

In an 80 -week pre-GLP study conducted in rats, which tested dietary malathion concentrations of 0,4700 and 8150 ppm (equivalent to 0,1200 and $2400 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), it was not possible to identify a NOAEL for chronic toxicity because of the lack of reporting detail. While there was an increase in proliferative lesions of the thyroid in both sexes at both doses, these increases were not statistically significant in males and were significant in females only in a trend test and not by pairwise comparison when compared with groups of pooled controls. Overall, this study is not considered acceptable for the assessment of carcinogenicity because of the small number of rats in the concurrent control group ( 15 versus 50 in the treated groups) and the short duration of exposure.

In a subsequent 24 -month pre-GLP study conducted in rats, which tested dietary malathion concentrations of $0,100,1000$ and 5000 ppm (equivalent to $0,5,50$ and $250 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively, as calculated by a previous Meeting), the NOAEL was 100 ppm (equivalent to $5 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of erythrocyte acetylcholinesterase activity at 1000 ppm (equivalent to $50 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for carcinogenicity was 5000 ppm (equivalent to $250 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

In a 24 -month chronic toxicity and carcinogenicity study in rats, which tested dietary malathion concentrations of $0,100,500,6000$ and 12000 ppm (equal to $0,7,29,359$ and $729 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,8,35,415$ and $868 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for
chronic toxicity was 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) for reduced red cell parameters, inhibition of brain acetylcholinesterase activity and the occurrence of nasal toxicity at 6000 ppm (equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day). The nasal toxicity was characterized by olfactory epithelial degeneration, hyperplasia and cyst formation, goblet cell hyperplasia, congestion, oedema and inflammation. Four nasal adenomas were observed, one in each sex at the two highest doses. In females, but not males, the incidence of liver adenomas was increased slightly at 6000 and 12000 ppm, but the incidences were within the performing laboratory's historical control range. A NOAEL of 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) was identified for carcinogenicity, based on the increase in nasal adenomas at 6000 ppm (equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that there is some evidence that malathion is carcinogenic in rats and mice.

The Meeting noted that the mouse liver adenomas observed in the second 18 -month study occurred at doses exceeding the maximum tolerated dose and were not replicated in other mouse studies. The increases in liver adenomas in rats observed in the 24 -month chronic toxicity and carcinogenicity study occurred only in females and were within the performing laboratory's historical control range. Whereas the rodent liver adenomas were co-incident with liver hypertrophy, there were no findings in these or other studies to suggest a possible mode of action, such as liver enzyme induction or cytotoxicity. Malathion showed no peroxisome proliferator-activated receptor alpha or gamma activity and also showed no aryl hydrocarbon receptor activity. Overall, the Meeting considered that there was equivocal evidence to suggest a tumorigenic response in the liver, but this had a clear threshold and was likely to be secondary to the effects on the liver of prolonged exposure to very high dietary concentrations of malathion.

Based on consistent observations of nasal toxicity in dietary studies of various durations ranging from 28 days to 2 years and in a short-term inhalational toxicity study, the Meeting concluded that the formation of nasal adenomas in rats was due to a local mechanism of irritancy and cytotoxicity caused by prolonged exposure of the nasal epithelium to high concentrations of malathion absorbed via inhaled food particles or as a vapour arising from food. This produces a state of reactive hyperplasia, a causative factor in tumour formation. Scenarios of prolonged, direct and excessive exposure of human nasal tissue to malathion or malathion metabolites following ingestion of residues is unlikely, and therefore these tumours would not occur in humans following exposure to malathion in the diet.

Malathion has been extensively tested for genotoxicity using a broad range of in vitro and in vivo assays. In 1997, the Meeting evaluated the available unpublished and published genotoxicity studies and noted that the majority of studies indicated that malathion is not genotoxic, although a small number of studies indicated that it can induce chromosomal aberrations and sister chromatid exchanges in vitro. However, there was no evidence that malathion induced chromosomal aberrations in vivo. Therefore, the 1997 Meeting concluded that malathion does not induce genotoxic damage in vivo. The 2003 Meeting evaluated supplementary genotoxicity studies and found that malathion caused chromosomal aberrations in cultured human lymphocytes and gene mutations in the mouse lymphoma assay at cytotoxic concentrations, but did not cause unscheduled DNA synthesis in vivo in male rats. The 2003 Meeting reaffirmed its previous conclusion that although the results of some in vitro tests were positive, malathion was not considered to induce genotoxic damage in vivo.

In addition to the studies considered at previous meetings, the current Meeting considered a number of new published and unpublished genotoxicity studies, including studies that involved the assessment of genotoxic damage in exposed workers. Many of the published studies do not provide adequate experimental detail, do not specify the purity of the malathion tested or were conducted on commercial formulations, or used in vivo test systems or exposure routes less relevant to the risk assessment of dietary residues of pesticides. The following discussion is limited to studies that evaluated technical malathion or malathion at purities above $90 \%$ and provided adequate experimental and data analysis details to allow interpretation of the findings.

Using standard genotoxicity test systems, malathion was not mutagenic in assays using prokaryotes or lower eukaryotes when tested with or without metabolic activation. In contrast, in in vitro assays using either human or non-human cells, malathion was generally positive for the induction of (1) chromosomal damage, as measured by increased frequencies of chromosomal aberrations or micronuclei; (2) mutations; and (3) DNA damage, as measured by increases in DNA migration in the alkaline comet assay and increased frequencies of sister chromatid exchanges. Negative findings were reported for the induction of micronuclei in Molt-4 T-lymphocytes, unscheduled DNA synthesis in WI-38 cells and primary rat liver hepatocytes, and mutations in a mouse lymphoma assay (reported to be equivocal without metabolic activation and negative with metabolic activation).

Using in vivo non-mammalian systems, malathion was active for micronucleus induction in a bird model and for induction of reciprocal translocations and sex-linked recessive lethals in one Drosophila melanogaster study, but not for sex-linked recessive lethals, sex chromosome loss or wing spot mutations in another study.

Based on the criteria mentioned in section 2.1, very few of the 34 in vivo mammalian study/end-point combinations were considered adequate for this review. In reports submitted by the sponsor, malathion was negative in a rat liver unscheduled DNA synthesis study when administered by gavage, in a rat bone marrow chromosomal aberration study when administered by gavage and in a mouse bone marrow erythrocyte micronucleus assay when administered intraperitoneally. However, the unscheduled DNA synthesis assay is insensitive for detecting genotoxic compounds; the micronucleus assay, as conducted, suffers from concerns about scoring criteria; and the chromosomal aberration test appears to be significantly underpowered, based on the frequency of chromosomal aberrations detected among control and treated animals. A negative mouse dominant lethal test was also reported when malathion was administered in feed for 7 weeks, and a negative mouse bone marrow chromosomal aberration study was reported in intraperitoneally treated mice. In contrast, malathion-induced micronuclei and chromosomal aberrations were reported in bone marrow immature erythrocytes and proliferating cells, respectively. A positive alkaline comet assay using blood leukocytes sampled from rats treated intraperitoneally once a day for 5 days was reported.

The Meeting evaluated a number of human studies that examined genotoxicity end-points. Patients treated for acute intoxication with a malathion-based product exhibited increased levels of chromosomal damage in lymphocytes. The frequency of micronuclei and glycophorin A mutations in erythrocytes or micronuclei in lymphocytes was not increased in workers exposed selectively to malathion. However, DNA damage and chromosomal aberrations have been reported in workers exposed to a mixture of pesticides, including malathion. These studies are of limited value for examining the specific effect of malathion on genotoxicity end-points in humans.

The Meeting noted that malathion has been reported to have genotoxic activity in multiple assay systems at multiple genetic end-points. In several studies where evaluated, reactive oxygen species appear to have been responsible for the increased damage, as demonstrated by the detection of malathion-induced 8 -hydroxy-2'-deoxyguanosine and increased malondialdehyde concentrations in isolated human peripheral blood mononuclear cells treated in vitro, an effect attenuated by cotreatment with $N$-acetylcysteine or curcumin; by increased intracellular levels of reactive oxygen species and reduced levels of catalase, superoxide dismutase and glutathione in rat PC12 cells treated in vitro, an effect ameliorated by co-treatment with vitamin E; and by the detection of oxidative damage using the comet assay in isolated rat lymphocytes treated in vitro with malathion. Supportive of this hypothesis, malathion appears to selectively induce markers of oxidative stress in Tox21/ToxCast high-throughput screening assays. The Meeting concluded that the observed genotoxic effects occur secondary to the formation of reactive oxygen species, which will exhibit a threshold.

The Meeting concluded that malathion is unlikely to be genotoxic at anticipated dietary exposures.

In the multigeneration and developmental toxicity studies, cholinesterase activity was not measured.

In a two-generation reproductive toxicity study conducted in rats, which tested dietary malathion concentrations of $0,550,1700,5000$ and 7500 ppm (equal to $0,43,130,393$ and 595 $\mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,50,152,438$ and $655 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for both reproductive toxicity and parental toxicity was 7500 ppm (equal to $595 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested. The NOAEL for offspring toxicity was 1700 ppm (equal to 130 $\mathrm{mg} / \mathrm{kg}$ bw per day) for reduced pup weights at 5000 ppm (equal to $393 \mathrm{mg} / \mathrm{kg}$ bw per day).

Two published studies reported potential testicular toxicity in rats exposed to malathion orally, but these studies had a number of methodological limitations that reduced their utility. Further, the reported observations are not corroborated by the preceding GLP-compliant multigenerational rat study in which no effects on the testes were observed.

A variety of in vivo and in vitro assays in mammalian and non-mammalian models indicated that malathion is unlikely to affect the endocrine system.

In a pilot developmental toxicity study in rats, which tested gavage malathion doses of 0,300 , 600,800 and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 15 of gestation, no embryo or fetal toxicity occurred, whereas maternal toxicity occurred at and above $600 \mathrm{mg} / \mathrm{kg}$ bw per day. In the main developmental toxicity study in rats, which tested gavage doses of $0,200,400$ and $800 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 15 of gestation, the NOAEL for maternal toxicity was $400 \mathrm{mg} / \mathrm{kg}$ bw per day for clinical signs and reduced body weight gain and feed consumption at $800 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo and fetal toxicity was $800 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

In a range-finding developmental toxicity study in rabbits, which tested gavage malathion doses of $0,25,50,100,200$ and $400 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 18 of gestation, no embryo or fetal toxicity occurred, whereas maternal toxicity occurred at 200 and $400 \mathrm{mg} / \mathrm{kg}$ bw per day. In the main study, which tested malathion doses of $0,25,50$ and $100 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 18 of gestation, the NOAEL for maternal toxicity was $25 \mathrm{mg} / \mathrm{kg}$ bw per day for a marginal effect on body weight gain at $50 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo and fetal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

The Meeting concluded that malathion is not teratogenic.
In a study conducted in hens, there was no evidence that malathion caused delayed peripheral neuropathy.

In an acute neurotoxicity study in rats, which tested gavage malathion doses of $0,500,1000$ and $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, the NOAEL was $1000 \mathrm{mg} / \mathrm{kg}$ bw for reduced erythrocyte acetylcholinesterase activity in females and reduced ambulatory activity in males at $2000 \mathrm{mg} / \mathrm{kg}$ bw.

A 13-week neurotoxicity study in rats is described above together with the other 13-week toxicity studies in rats, and an overall NOAEL is identified for these studies.

In a developmental neurotoxicity study in rats, which tested gavage malathion doses of $0,5,50$ and $150 \mathrm{mg} / \mathrm{kg}$ bw per day from day 6 of gestation to day 10 of lactation, the NOAEL for both maternal toxicity and offspring toxicity was $50 \mathrm{mg} / \mathrm{kg}$ bw per day for clinical signs at $150 \mathrm{mg} / \mathrm{kg}$ bw per day.

Administration of malathion from day 6 of gestation to day 21 of lactation had no effect on the thickness of the corpus callosum in rat pups at doses up to $150 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that malathion is neurotoxic.
Studies in rats have examined the time to peak effect and compared the effects of malathion and malaoxon on the inhibition of acetylcholinesterase activity. The time to peak effect in juvenile rats following dosing with malathion ranged from 30 to 90 minutes for the inhibition of erythrocyte acetylcholinesterase activity and from 60 to 90 minutes for the inhibition of brain
acetylcholinesterase activity. Malaoxon was a more potent inhibitor of acetylcholinesterase activity compared with malathion. Comparison of benchmark doses (BMDs) following acute oral dosing indicated that the toxicity adjustment factor (TAF) for malaoxon was 21.5 in males and 17.4 in females for the inhibition of erythrocyte acetylcholinesterase activity and 14.8 in males and 11.0 in females for the inhibition of brain acetylcholinesterase activity. Comparison of BMDs for the inhibition of erythrocyte acetylcholinesterase activity from chronic toxicity studies indicated that TAFs for malaoxon ranged from 37 to 38 in males and from 65 to 69 in females.

In a 6 -week immunotoxicity study in female rats, which tested dietary malathion concentrations of $0,50,100,700$ and 7000 ppm (equal to $0,8.9,17.6,126.8$ and $1215.8 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), the NOAEL for immunotoxicity was 7000 ppm (equal to $1215.8 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

The Meeting concluded that malathion is not immunotoxic.
An extensive literature search did not identify any potential adverse effects on intestinal microbiota or any evidence that intestinal microbiota can metabolize malathion.

## Toxicological data on metabolites, degradates and/or impurities

Current FAO specifications for malathion prescribe maximum limits for isomalathion (CAS No. 3344-12-5), malaoxon (CAS No. 152-20-05), O,O,S-trimethyl phosphorothioate (CAS No. 2953-299 ) and $O, S, S$-trimethyl phosphorodithoate (CAS No. 152-18-1).

Toxicity tests were conducted on malaoxon, isomalathion, desmethyl malathion, desmethyl malathion monocarboxylic acid, MMCA, MDCA and desmethyl malaoxon dicarboxylic acid.

## Malaoxon

The oral $\mathrm{LD}_{50}$ in rats for malaoxon was $50 \mathrm{mg} / \mathrm{kg}$ bw.
In a 14 -day range-finding study in rats, which tested malaoxon at dietary concentrations of 0 , $10,25,100,2500$ and 3500 ppm (equal to $0,1.1,3.0,12.1,293$ and $387 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1.1,3.1,12.5,281.6$ and $294.7 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), inhibition of erythrocyte acetylcholinesterase activity occurred at and above 100 ppm (equal to $12.1 \mathrm{mg} / \mathrm{kg}$ bw per day). At the two highest doses, inhibition of brain acetylcholinesterase activity and reduced body weight gain and feed consumption occurred.

In a 103 -week carcinogenicity study conducted in mice, which tested dietary malaoxon concentrations of 0,500 and 1000 ppm (estimated by a previous Meeting to be equal to 0,75 and 150 $\mathrm{mg} / \mathrm{kg}$ bw per day, respectively), survival and body weight were reduced at the highest dose. There were no treatment-related neoplastic or non-neoplastic lesions. In a parallel study conducted in rats, which tested the same dietary concentrations of malathion (equal to 0,25 and $50 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), the combined incidence of C-cell adenomas and carcinomas of the thyroid in females was increased, although this was comparable to historical control values. The incidence of gastric ulcers, commonly observed in the forestomach, was increased in treated rats.

In a 24 -month toxicity study in rats, which tested malaoxon at dietary concentrations of 0,20 , 1000 and 2000 ppm (equal to $0,1,57$ and $110 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1,68$ and 140 $\mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for chronic toxicity was 20 ppm (equal to 1 $\mathrm{mg} / \mathrm{kg}$ bw per day), based on mortality and the inhibition of brain acetylcholinesterase activity at 1000 ppm (equal to $57 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for carcinogenicity was 2000 ppm (equal to $110 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested. Similar to studies conducted on malathion, inflammatory changes in the nasal mucosa occurred at 1000 and 2000 ppm ; these changes were likely attributable to inhaled food particles containing malaoxon, resulting in tissue injury and inflammation of the nasal cavity, with secondary effects on the lungs and middle ear.

The Meeting concluded that malaoxon is not carcinogenic in mice or rats.
Malaoxon was negative for mutagenicity in bacterial assays and in lower eukaryotes, both with and without metabolic activation. Malaoxon was reported to be active for induction of sister chromatid exchanges but not chromosomal aberrations in Chinese hamster ovary cells, with or without metabolic activation. An increase in sister chromatid exchanges when tested in the absence of metabolic activation only was also reported; it was also reported that malaoxon was more potent than malathion in this assay. Malaoxon was also reported to induce DNA damage as measured by the comet assay in rat adrenal gland PC12 cells when tested in the absence of metabolic activation only and was mutagenic in mouse lymphoma (L5178Y) cells in the absence but not the presence of metabolic activation. In this study, there seemed to be a preference for the induction of small colonies, generally considered to be indicative of chromosomal damage rather than gene mutations.

Malaoxon induced DNA damage in isolated lymphocytes in the absence of metabolic activation, as measured by the alkaline comet assay; studies with metabolic activation were not conducted. Further, a follow-up study concluded that the malaoxon-mediated damage was likely induced by reactive oxygen species. Also, malaoxon is more potent than malathion in inducing intracellular levels of reactive oxygen species and reducing levels of catalase, superoxide dismutase and glutathione in rat PC12 cells treated in vitro, an effect ameliorated by co-treatment with vitamin E. Also, similar to malathion, malaoxon appears to selectively induce markers of oxidative stress in Tox21/ToxCast high-throughput screening assays. When provided in food, malaoxon induced an increase in reciprocal translocations and sex-linked recessive lethals in D. melanogaster, but not for sex-linked recessive lethals when administered by injection. Malaoxon was reported negative for the induction of chromosomal aberrations and sister chromatid exchanges in the bone marrow cells of male mice following a single intraperitoneal injection.

The Meeting concluded that the observed genotoxic effects occur secondary to the formation of reactive oxygen species, which will exhibit a threshold.

The Meeting concluded that malaoxon is unlikely to be genotoxic at anticipated dietary exposures.

## Other metabolites

The oral $\mathrm{LD}_{50}$ in rats was greater than $2000 \mathrm{mg} / \mathrm{kg}$ bw for desmethyl malathion, desmethyl malathion monocarboxylic acid, MMCA, MDCA and desmethyl malaoxon dicarboxylic acid. The oral $\mathrm{LD}_{50}$ in rats for desmethyl malaoxon dicarboxylic acid, trisodium salt, was greater than $2000 \mathrm{mg} / \mathrm{kg}$ bw.

There are a limited number of genotoxicity studies on other metabolites of malathion. MDCA, MMCA, desmethyl malathion monocarboxylic acid, potassium salt, and desmethyl malaoxon dicarboxylic acid, trisodium salt, as well as isomalathion, $O, O, O$-trimethyl phosphorothioate, $O, O, S$ trimethyl phosphorothioate and $O, S, S$-trimethyl phosphorodithioate, were reported negative for bacterial mutagenicity, with and without metabolic activation. Isomalathion induced DNA damage in isolated lymphocytes in the absence of metabolic activation, as measured by the alkaline comet assay; studies with metabolic activation were not conducted. Isomalathion was also reported to induce micronuclei in the human liver-derived HepaRG cell line.

Using quantitative structure-activity relationships, the storage impurity, 2-mercaptosuccinic acid diethyl ester, was determined to have no greater toxicity than malathion.

The potential of malathion metabolites to inhibit acetylcholinesterase activity has been studied in rats. Comparisons of erythrocyte acetylcholinesterase activities indicated that desmethyl malathion, MMCA and MDCA are at least 2.75-, 1.9- and 4.6-fold less potent than malathion.

Based on a comparison of the inhibitions of acetylcholinesterase activities over acute and chronic exposure durations and a comparison of BMDs (see above), the Meeting concluded that malaoxon is approximately 30 -fold more potent than malathion.

## Human data

As in laboratory animals, the inhibition of acetylcholinesterase activity is the most sensitive adverse effect in humans exposed to malathion, mediated through the metabolite malaoxon, which is a more potent inhibitor of acetylcholinesterase activity compared with malathion. A comparative in vitro study indicated that malaoxon was a slightly less potent inhibitor ( $<2$-fold) of human compared with rat acetylcholinesterase activity.

In a study conducted in male and female volunteers, which tested single oral doses of malathion at $0,0.5,1.5,5,10$ and $15 \mathrm{mg} / \mathrm{kg}$ bw, the NOAEL was $15 \mathrm{mg} / \mathrm{kg}$ bw, the highest dose tested, based on the absence of any adverse effects, including the inhibition of erythrocyte acetylcholinesterase activity. In a subsequent study conducted in male and female volunteers, which tested single oral doses of malathion of $0,0.5,1.5,5.0,10.0$ and $15.0 \mathrm{mg} / \mathrm{kg}$ bw, there were no treatment-related adverse events or effects on erythrocyte acetylcholinesterase activity.

In a published study, application of malathion to the forearm of human volunteers increased blood flow, mediated via the inhibition of acetylcholinesterase activity.

In a published non-blinded study, slight inhibition of erythrocyte acetylcholinesterase activity occurred in children following two applications of a $1 \%$ malathion shampoo used to treat head lice.

In a 1994 summary report, there were no poisoning incidents and no inhibition of plasma cholinesterase activity in workers involved in the manufacture of malathion over a 20 -year period. In a subsequent (1999) summary report, biological monitoring of workers employed at dimethoate and malathion manufacturing plants from 1994 to 1999 detected no reduction in plasma cholinesterase activity.

Several epidemiological studies on cancer outcomes in relation to occupational exposure to malathion were available. The evaluation of these studies focused on the occurrence of NHL and prostate cancer, as outlined in section 2.2. One meta-analysis was available, as well as one prospective cohort study, the AHS, with a large sample size and detailed exposure assessment. Cohort studies are considered a powerful design, as recall bias is avoided. All other studies were case-control studies, usually retrospective, which are more prone to recall and selection biases.

The AHS found no evidence of a positive association of NHL with malathion exposure or of an exposure-response relationship. In contrast, various case-control studies reported excess risks of NHL associated with use of malathion. In a large pooled case-control study, the unadjusted estimates showed a significant increased risk of NHL $(R R=1.6 ; 95 \% C I=1.2-2.2)$ associated with ever versus never use of malathion. However, these were attenuated and/or no longer significant when proxy respondents were excluded and analyses were mutually adjusted for other pesticides. Significant elevated risks of NHL were reported from the cross-Canada case-control study of pesticides and health for ever versus never use of malathion ( $\mathrm{OR}=1.96 ; 95 \% \mathrm{CI}=1.42-2.70$ ) and when examining annual days of use, although there was no clear exposure-response relationship across exposure categories. Non-significant increased risks of NHL were reported by two other case-control studies, one of which had limited statistical power based on only five exposed cases. The meta-analysis, which did not include the AHS, found a significant $80 \%$ excess risk ratio for ever versus never use of malathion.

Overall, there is some very weak evidence of a positive association between malathion exposure and NHL from the case-control studies and the overall meta-analysis. However, it is notable that the AHS, which is the only cohort study and is large and of high quality, found no evidence of an association at any exposure level.

There was no evidence of an association with all prostate cancers and malathion exposure in the AHS. However, a significant excess risk of aggressive prostate cancer $(\mathrm{RR}=1.43 ; 95 \% \mathrm{CI}=$ 1.08-1.88) in the highest exposure category (highest quintile of intensity-weighted lifetime days of malathion exposure), along with a significant exposure-response relationship ( $P$ for trend $=0.04$ ), was observed. A significant elevated risk of all prostate cancer was observed in a case-control study
for ever use $(\mathrm{OR}=1.34 ; 95 \% \mathrm{CI}=1.01-1.78)$ and for highest lifetime cumulative exposure versus those unexposed $(\mathrm{OR}=1.49 ; 95 \% \mathrm{CI}=1.02-2.18)$. A significant trend across exposure categories $(P=0.03)$ was also reported. However, interpretation of results from this study is limited by potential for exposure misclassification in the job-exposure matrix used for exposure assessment and by the potential for residual confounding from lack of adjustment for other pesticide exposures. There was no evidence of an association between prostate cancer and malathion exposure in the United Farm Workers of America study, which was limited by the use of ecological rather than individual-level exposure assessment.

Overall, the evidence is suggestive of a positive association between malathion exposure and risk of aggressive prostate cancer; however, the evidence base is limited to the one large AHS cohort study.

Based on a consideration of the results of animal bioassays, genotoxicity assays and epidemiological data from occupational exposures, the Meeting concluded that malathion and its metabolites are unlikely to pose a carcinogenic risk to humans from exposure via the diet.

The Meeting concluded that the existing database on malathion was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The current Meeting reaffirmed the ADI of $0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, based on the NOAEL of 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) in the 2-year study of toxicity and carcinogenicity in rats for the inhibition of brain acetylcholinesterase and using a 100-fold safety factor, established by the 1997 Meeting. The margins of exposure between this ADI and the doses causing liver adenomas in mice and nasal adenomas in rats are 5000 -fold and 1200 -fold, respectively.

The current Meeting reaffirmed the ARfD of $2 \mathrm{mg} / \mathrm{kg}$ bw, based on the NOAEL of $15 \mathrm{mg} / \mathrm{kg}$ bw for the inhibition of erythrocyte acetylcholinesterase activity in a study conducted in male and female volunteers with the application of a 10 -fold safety factor, established by the 2003 Meeting. This ARfD is supported by the NOAEL of $15 \mathrm{mg} / \mathrm{kg}$ bw in a second study conducted in male and female volunteers. The ARfD is considered to be a conservative value, because human acetylcholinesterase is slightly less sensitive ( $<2$-fold) than rat acetylcholinesterase to malaoxon.

The Meeting concluded that the metabolite malaoxon is approximately 30 -fold more toxic than malathion. On this basis, a 30 -fold potency factor should be applied to the residue levels for use in both the acute and chronic dietary exposure estimates for malaoxon, and these should be added to the dietary exposures for malathion and compared with the ARfD and ADI for malathion, respectively.

Both the ADI and ARfD are established for the sum of malathion and malaoxon (corrected for its potency), expressed as parent malathion. The other metabolites of malathion considered by the present Meeting are less potent than the parent compound and therefore would be covered by the ADI and ARfD for malathion. The impurity isomalathion may need to be taken into consideration in the risk assessment depending on its concentration in food commodities.

A toxicological monograph was prepared.

## Levels relevant to risk assessment of malathion

| Species | Study | Effect | NOAEL | LOAEL |
| :--- | :--- | :--- | :--- | :--- |
| Mouse | Two-year study of toxicity | Toxicity | 800 ppm, equal to | 8000 ppm, equal to |
|  | and carcinogenicity $^{\text {a }}$ |  | $143 \mathrm{mg} / \mathrm{kg}$ bw per <br>  |  |
|  |  |  | $1476 \mathrm{mg} / \mathrm{kg}$ bw per |  |


| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Carcinogenicity | 800 ppm , equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day | 8000 ppm , equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rat | Acute neurotoxicity study ${ }^{\text {b }}$ | Toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day | $2000 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | One-month studies of toxicity ${ }^{\mathrm{a}, \mathrm{c}}$ | Toxicity | 500 ppm , equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Thirteen-week studies of toxicity and neurotoxicity ${ }^{\mathrm{a}, \mathrm{c}}$ | Toxicity | 500 ppm , equal to $34 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-year study of toxicity and carcinogenicity ${ }^{\text {a }}$ | Toxicity | 500 ppm , equal to 29 $\mathrm{mg} / \mathrm{kg}$ bw per day | 6000 ppm , equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 500 ppm , equal to 29 $\mathrm{mg} / \mathrm{kg}$ bw per day | 6000 ppm , equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-generation study of reproductive toxicity ${ }^{\text {a,e }}$ | Reproductive toxicity | $\begin{aligned} & 7500 \mathrm{ppm} \text {, equal to } \\ & 595 \mathrm{mg} / \mathrm{kg} \text { bw per } \\ & \text { day }^{\mathrm{d}} \end{aligned}$ | - |
|  |  | Parental toxicity | $\begin{aligned} & 7500 \mathrm{ppm} \text {, equal to } \\ & 595 \mathrm{mg} / \mathrm{kg} \text { bw per } \\ & \text { day }^{\mathrm{d}} \end{aligned}$ | - |
|  |  | Offspring toxicity | 1700 ppm , equal to $130 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $393 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity study ${ }^{\text {b,e }}$ | Maternal toxicity | $400 \mathrm{mg} / \mathrm{kg}$ bw per day | $800 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $800 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {d }}$ | - |
|  | Developmental neurotoxicity study ${ }^{\text {b,e }}$ | Maternal toxicity | $50 \mathrm{mg} / \mathrm{kg}$ bw per day | $150 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | $50 \mathrm{mg} / \mathrm{kg}$ bw per day | $150 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity study ${ }^{\text {be }}$ | Maternal toxicity | $25 \mathrm{mg} / \mathrm{kg}$ bw per day | $50 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $100 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {d }}$ | - |
| Dog | One-year study of toxicity ${ }^{\text {f }}$ | Toxicity | $125 \mathrm{mg} / \mathrm{kg}$ bw per day | $250 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Human | Acute volunteer studies ${ }^{\text {c,f }}$ | Cholinesterase inhibition | $15 \mathrm{mg} / \mathrm{kg} \mathrm{bw}{ }^{\text {d }}$ | - |

${ }^{\text {a }}$ Dietary administration.
${ }^{\text {b }}$ Gavage administration.
c Two or more studies combined.
${ }^{\mathrm{d}}$ Highest dose tested.
${ }^{\text {e }}$ Acetylcholinesterase activity not measured.
${ }^{\mathrm{f}}$ Capsule administration.

## Estimate of acceptable daily intake (ADI)

$0-0.3 \mathrm{mg} / \mathrm{kg}$ bw (for sum of malathion and malaoxon, adjusted for its potency, and expressed as malathion)

## Estimate of acute reference dose (ARfD)

$2 \mathrm{mg} / \mathrm{kg}$ bw (for sum of malathion and malaoxon, adjusted for its potency, and expressed as malathion)

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

Results from in vivo genotoxicity studies investigating oral dosing, because malathion genotoxicity data are highly variable and inconsistent and there is a lack of robust in vivo rodent studies using the oral route of exposure

Critical end-points for setting guidance values for exposure to malathion

| Rate and extent of oral absorption | Rapid; > 77\% |
| :---: | :---: |
| Dermal absorption | Estimates vary (1.44-20.7\% in human skin) |
| Distribution | Rapid tissue distribution |
| Potential for accumulation | No potential for accumulation |
| Rate and extent of excretion | Rapid and complete |
| Metabolism in animals | Extensive; oxidation, hydrolysis, dealkylation and demethylation reactions |
| Toxicologically significant compounds in animals and plants | Malathion, malaoxon, desmethyl malathion, desmethyl malaoxon, MMCA, MDCA, isomalathion |
| Acute toxicity |  |
| Rat, $\mathrm{LD}_{50}$, oral | $>1539$ to <8227 mg/kg bw |
| Rat, $\mathrm{LD}_{50}$, dermal | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| Rat, $\mathrm{LC}_{50}$, inhalation | $>5.2 \mathrm{mg} / \mathrm{L}$ |
| Rabbit, dermal irritation | Slightly irritating |
| Rabbit, ocular irritation | Slightly irritating |
| Guinea-pig, dermal sensitization | Not sensitizing (Buehler assay) |
|  | Sensitizing (maximization assay) |
| Mouse, dermal sensitization | Not sensitizing (local lymph node assay) |
| Short-term studies of toxicity |  |
| Target/critical effect | Acetylcholinesterase inhibition |
| Lowest relevant oral NOAEL | $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day (28 days; rat) |
| Lowest relevant dermal NOAEL | $150 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rabbit) |
| Lowest relevant inhalation NOAEC | $<0.1 \mathrm{mg} / \mathrm{L}$ (13 weeks; rat) |

Long-term studies of toxicity and carcinogenicity
Target/critical effect Acetylcholinesterase inhibition


## Toxicological studies on MDCA

| Rat, $\mathrm{LD}_{50}$, oral | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Genotoxicity | Not mutagenic in prokaryotic assays |

Toxicological studies on desmethyl malaoxon
dicarboxylic acid

| Rat, $\mathrm{LD}_{50}$, oral | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Genotoxicity | Not mutagenic in prokaryotic assays |
| Human data | Acetylcholinesterase inhibition: |
|  | Acute NOAEL: $15 \mathrm{mg} / \mathrm{kg}$ bw, highest dose tested |
|  | No adverse effects in manufacturing personnel |
|  |  |

[^8]
## Summary

|  | Value | Studies | Safety factor |
| :--- | :--- | :--- | :--- |
| ADI | $0-0.3 \mathrm{mg} / \mathrm{kg}$ bw | Two-year chronic toxicity and carcinogenicity study <br> (rat) | 100 |
| ARfD | $2 \mathrm{mg} / \mathrm{kg}$ bw | Single-dose studies (humans) | 10 |

## DIETARY RISK ASSESSMENT

The current residue definition for the estimation of dietary exposure is malathion. The Meeting identified that malaoxon is approximately 30 times more potent than malathion based on the endpoint (acetylcholinesterase inhibition) on which the ADI and ARfD have been established. Malaoxon is generally present in food at concentrations that are approximately $3 \%$ of the malathion concentration. If malaoxon were included in the residue definition for dietary risk assessment, the exposures calculated below for comparison with the health-based guidance values would be approximately double.

## Long-term dietary exposure

The ADI for malathion is $0-0.3 \mathrm{mg} / \mathrm{kg}$ bw. The IEDIs for malathion were estimated for the 17 GEMS/Food cluster diets using the STMR or STMR-P values estimated by JMPR. The results are shown in Annex 3. The IEDI ranged from $0.1 \%$ to $0.5 \%$ of the maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of malathion from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The ARfD for malathion is $2 \mathrm{mg} / \mathrm{kg}$ bw. The IESTI for malathion was calculated for the plant commodities for which STMR and HR levels were estimated by the 1999, 2004 and 2008 JMPRs and for which consumption data were available. The results are shown in Annex 4. The calculated IESTIs were $0-5 \%$ of the ARfD for the general population and $0-9 \%$ of the ARfD for children. The Meeting concluded that the short-term dietary exposure to malathion residues from uses considered by the Meeting was unlikely to present a public health concern.

## 6. RECOMMENDATIONS

The Meeting recommended that a guidance document be developed for the evaluation of genotoxicity studies, taking the experience gained from this meeting into account.

## ANNEX 1: ACCEPTABLE DAILY INTAKES AND ACUTE REFERENCE DOSES RECORDED BY THE MAY 2016 MEETING

| Pesticide (Codex reference <br> number) | Acceptable daily intake (ADI) <br> $(\mathbf{m g} / \mathbf{k g}$ bw) | Acute reference dose (ARfD) <br> $(\mathbf{m g} / \mathbf{k g}$ bw) |
| :--- | :---: | :---: |
| Diazinon (22) | $0-0.003$ | 0.03 |
| Glyphosate (158) | $0-1^{\mathrm{a}}$ | Unnecessary |
| Malathion (49) | $0-0.3^{\mathrm{b}}$ | $2^{\mathrm{b}}$ |
| a |  |  |

[^9]
## ANNEX 2: INDEX OF REPORTS AND EVALUATIONS OF PESTICIDES BY THE JMPR

Numbers in parentheses after the names of pesticides are Codex classification numbers. The abbreviations used are:

T, evaluation of toxicology
R , evaluation of residue and analytical aspects
E, evaluation of effects on the environment

| Abamectin (177) | $\begin{aligned} & 1992 \text { (T,R), } 1994 \text { (T,R), } 1995 \text { (T), } 1997 \text { (T,R), } \\ & 2000 \text { (R), } 2015 \text { (R) } \end{aligned}$ |
| :---: | :---: |
| Acephate (095) | 1976 (T,R), 1979 (R), 1981 (R), 1982 (T), 1984 (T,R), 1987 (T), 1988 (T), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1994 (R), 1996 (R), 2002 (T), 2003 (R), 2004 (corr. to 2003 report), 2005 (T), 2006 (R), 2011 (R) |
| Acetamiprid (246) | 2011 (T,R), 2012 (R), 2015 (R) |
| Acetochlor (280) | 2015 (T,R) |
| Acrylonitrile | 1965 (T,R) |
| Aldicarb (117) | 1979 (T,R), 1982 (T,R), 1985 (R), 1988 (R), 1990 (R), 1991 (corr. to 1990 evaluation), 1992 (T), 1993 (R), 1994 (R), 1996 (R), 2001 (R), 2002 (R), 2006 (R) |
| Aldrin (001) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (R), } 1974 \text { (R), } 1975 \text { (R), } \\ & 1977 \text { (T), } 1990 \text { (R), } 1992 \text { (R) } \end{aligned}$ |
| Allethrin | 1965 (T,R) |
| Ametoctradin (253) | 2012 (T,R) |
| Aminocarb (134) | 1978 (T,R), 1979 (T,R) |
| Aminocyclopyrachlor (272) | 2014 (T,R) |
| Aminomethylphosphonic acid (AMPA, 198) | 1997 (T,R) |
| Aminopyralid (220) | 2006 (T,R), 2007 (T,R) |
| Amitraz (122) | 1980 (T,R), 1983 (R), 1984 (T,R), 1985 (R), 1986 (R), 1989 (R), 1990 (T,R), 1991 (R \& corr. to 1990 R evaluation), 1998 (T) |
| Amitrole (079) | 1974 (T,R), 1977 (T), 1993 (T,R), 1997 (T), 1998 (R) |
| Anilazine (163) | 1989 (T,R), 1992 (R) |
| Atrazine | 2007 (T) |
| Azinphos-ethyl (068) | 1973 (T,R), 1983 (R) |
| Azinphos-methyl (002) | $\begin{aligned} & 1965(\mathrm{~T}), 1968 \text { (T,R), } 1972(\mathrm{R}), 1973(\mathrm{~T}), 1974(\mathrm{R}), \\ & 1991 \text { (T,R), } 1992 \text { (corr. to } 1991 \text { report), } 1993(\mathrm{R}) \text {, } \\ & 1995(\mathrm{R}), 2007(\mathrm{~T}) \end{aligned}$ |

Azocyclotin (129)

Azoxystrobin (229)
Benalaxyl (155)

Bendiocarb (137)
Benomyl (069)

Bentazone (172)

Benzovinfiflupyr (261)
BHC (technical-grade)
Bifenazate (219)
Bifenthrin (178)

Binapacryl (003)

Bioresmethrin (093)
Biphenyl
Bitertanol (144)

Bixafen (262)
Boscalid (221)
Bromide ion (047)

Bromomethane (052)

Bromophos (004)

Bromophos-ethyl (005)
Bromopropylate (070)
Butocarboxim (139)
Buprofezin (173)
sec-Butylamine (089)

Cadusafos (174)
Campheclor (071)

1979 (R), 1981 (T), 1982 (R), 1983 (R), 1985 (R), 1989 (T,R), 1991 (R), 1994 (T), 2005 (T,R)

2008 (T,R), 2011 (R), 2012 (R), 2013 (R)
1986 (R), 1987 (T), 1988 (R), 1992 (R), 1993 (R), 2005 (T), 2009 (R)

1982 (T,R), 1984 (T,R), 1989 (R), 1990 (R)
1973 (T,R), 1975 (T,R), 1978 (T,R), 1983 (T,R), 1988 (R), 1990 (R), 1994 (R), 1995 (T,E), 1998 (R)

1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1995 (R), 1998 (T,R), 1999 (corr. to 1998 report), 2004 (T), 2012 (T), 2013 (R)

2013 (T), 2014 (R)
1965 (T), 1968 (T,R), 1973 (T,R) (see also Lindane)
2006 (T,R), 2008 (R), 2010 (R)
1992 (T,R), 1995 (R), 1996 (R), 1997 (R), 2009 (T), 2010 (R), 2015 (R)

1969 (T,R), 1974 (R), 1982 (T), 1984 (R), 1985 (T,R)

1975 (R), 1976 (T,R), 1991 (T,R)
See Diphenyl
1983 (T), 1984 (R), 1986 (R), 1987 (T), 1988 (R), 1989 (R), 1991 (R), 1998 (T), 1999 (R), 2002 (R)

2013 (T,R)
2006 (T,R), 2008 (R), 2010 (R)
1968 (R), 1969 (T,R), 1971 (R), 1979 (R), 1981 (R), 1983 (R), 1988 (T,R), 1989 (R), 1992 (R)

1965 (T,R), 1966 (T,R), 1967 (R), 1968 (T,R), 1971 (R), 1979 (R), 1985 (R), 1992 (R)

1972 (T,R), 1975 (R), 1977 (T,R), 1982 (R), 1984 (R), 1985 (R)

1972 (T,R), 1975 (T,R), 1977 (R)
1973 (T,R), 1993 (T,R)
1983 (R), 1984 (T), 1985 (T), 1986 (R)
1991 (T,R), 1995 (R), 1996 (corr. to 1995 report.), 1999 (R), 2008 (T,R), 2009 (R), 2012 (R), 2014 (R)

1975 (T,R), 1977 (R), 1978 (T,R), 1979 (R), 1980 (R), 1981 (T), 1984 (T,R: withdrawal of temporary ADI, but no evaluation)

1991 (T,R), 1992 (R), 1992 (R), 2009 (R), 2010 (R)
1968 (T,R), 1973 (T,R)

| Captafol (006) | $\begin{aligned} & 1969 \text { (T,R), } 1973 \text { (T,R), } 1974 \text { (R), } 1976 \text { (R), } \\ & 1977 \text { (T,R), } 1982 \text { (T), } 1985 \text { (T,R), } 1986 \text { (corr. to } \\ & 1985 \text { report), } 1990 \text { (R), } 1999 \text { (ARfD) } \end{aligned}$ |
| :---: | :---: |
| Captan (007) | 1965 (T), 1969 (T,R), 1973 (T), 1974 (R), 1977 (T,R), 1978 (T,R), 1980 (R), 1982 (T), 1984 (T,R), 1986 ( R ), 1987 ( R and corr. to 1986 R evaluation), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1994 (R), 1995 (T), 1997 (R), 2000 (R), 2004 (T), 2007 (T) |
| Carbaryl (008) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (T,R), } 1968 \text { (R), } \\ & 1969(\mathrm{~T}, \mathrm{R}), 1970 \quad(\mathrm{R}), 1973 \\ & 1976(\mathrm{R}), 1977(\mathrm{R}), 1979(\mathrm{R}), 1984(\mathrm{R}), 1996 \text { (T), }, 1975 \\ & 2001(\mathrm{~T}), 2002(\mathrm{R}), 2007(\mathrm{R}) \end{aligned}$ |
| Carbendazim (072) |  |
| Carbofuran (096) | $\begin{aligned} & 1976 \text { (T,R), } 1979 \text { (T,R), } 1980 \text { (T), } 1982 \text { (T), } \\ & 1991(\mathrm{R}), 1993 \text { (R), } 1996 \text { (T), } 1997 \text { (R), } 1999 \text { (corr. } \\ & \text { to } 1997 \text { report), } 2002 \text { (T,R), } 2003 \text { (R) (See also } \\ & \text { carbosulfan), } 2004 \text { (R), } 2008 \text { (T), } 2009 \text { (R) } \end{aligned}$ |
| Carbon disulfide (009) | 1965 (T,R), 1967 (R), 1968 (R), 1971 (R), 1985 (R) |
| Carbon tetrachloride (010) | $\begin{aligned} & 1965 \text { (T,R), } 1967 \text { (R), } 1968 \text { (T,R), } 1971 \text { (R), } \\ & 1979 \text { (R), } 1985(\mathrm{R}) \end{aligned}$ |
| Carbophenothion (011) | $\begin{aligned} & 1972 \text { (T,R), } 1976 \text { (T,R), } 1977 \text { (T,R), } 1979 \text { (T,R), } \\ & 1980 \text { (T,R), } 1983 \text { (R) } \end{aligned}$ |
| Carbosulfan (145) | 1984 (T,R), 1986 (T), 1991 (R), 1992 (corr. to 1991 report), 1993 (R), 1997 (R), 1999 (R), 2002 (R), 2003 (T,R), 2004 ( R , corr. to 2003 report) |
| Cartap (097) | 1976 (T,R), 1978 (T,R), 1995 (T,R) |
| Chinomethionat (080) | ```1968 (T,R) (as oxythioquinox), 1974 (T,R), 1977 (T,R), 1981 (T,R), 1983 (R), }1984\mathrm{ (T,R), 1987 (T)``` |
| Chlorantraniliprole (230) | 2008 (T,R), 2010 (R), 2013 (R), 2014 (R) |
| Chlorbenside | 1965 (T) |
| Chlordane (012) | $\begin{array}{lllll} 1965(\mathrm{~T}), \quad 1967 & (\mathrm{~T}, \mathrm{R}), \quad 1969 & (\mathrm{R}), 1970 & (\mathrm{~T}, \mathrm{R}), \\ 1972(\mathrm{R}), & 1974 & (\mathrm{R}), \quad 1977 & (\mathrm{~T}, \mathrm{R}), & 1982 \\ 1984(\mathrm{~T}, \mathrm{R}), 1986 & (\mathrm{~T}) & & & \end{array}$ |
| Chlordimeform (013) | 1971 (T,R), 1975 (T,R), 1977 (T), 1978 (T,R), 1979 (T), 1980 (T), 1985 (T), 1986 (R), 1987 (T) |
| Chlorfenapyr (254) | 2013 (T) |
| Chlorfenson | 1965 (T) |
| Chlorfenvinphos (014) | 1971 (T,R), 1984 (R), 1994 (T), 1996 (R) |
| Chlormequat (015) | $\begin{aligned} & 1970 \text { (T,R), } 1972 \text { (T,R), } 1976 \text { (R), } 1985 \text { (R), } \\ & 1994 \text { (T,R), } 1997 \text { (T), } 1999 \text { (ARfD), } 2000 \text { (R) } \end{aligned}$ |

Chlorobenzilate (016)

Chloropicrin
Chloropropylate
Chlorothalonil (081)

Chlorpropham (201)
Chlorpyrifos (017)

Chlorpyrifos-methyl (090)

Chlorthion
Clethodim (187)
Clofentezine (156)

Clothianidin (238)
Coumaphos (018)

Crufomate (019)
Cyanophenfos (091)

Cyantraniliprole (263)
Cyazofamid (281)
Cycloxydim (179)
Cyflumetofen (273)
Cyfluthrin (157)

Cyhalothrin (146)

Cyhexatin (067)

Cypermethrin (118)

1965 (T), 1968 (T,R), 1972 (R), 1975 (R), 1977 (R), 1980 (T)

1965 (T,R)
1968 (T,R), 1972 (R)
1974 (T,R), 1977 (T,R), 1978 (R), 1979 (T,R), 1981 (T,R), 1983 (T,R), 1984 (corr. to 1983 report and T evaluation), 1985 (T,R), 1987 (T), 1988 (R), 1990 (T,R), 1991 (corr. to 1990 evaluation), 1992 (T), 1993 (R), 1997 (R), 2009 (T), 2010 (R), 2012 (R), 2015 (R)

1965 (T), 2000 (T), 2001 (R), 2005 (T), 2008 (R)
1972 (T,R), 1974 (R), 1975 (R), 1977 (T,R), 1981 (R), 1982 (T,R), 1983 (R), 1989 (R), 1995 (R), 1999 (T), 2000 (R), 2004 (R), 2006 (R)

1975 (T,R), 1976 (R, Annex I only), 1979 (R), 1990 (R), 1991 (T,R), 1992 (T and corr. to 1991 report), 1993 (R), 1994 (R), 2001 (T), 2009 (R)

1965 (T)
1994 (T,R), 1997 (R), 1999 (R), 2002 (R)
1986 (T,R), 1987 (R), 1989 (R), 1990 (R), 1992 (R), 2005 (T), 2007 (R)

2010 (T,R), 2011 (R), 2014 (R)
1968 (T,R), 1972 (R), 1975 (R), 1978 (R), 1980 (T,R), 1983 (R), 1987 (T), 1990 (T,R)

1968 (T,R), 1972 (R)
1975 (T,R), 1978 (T: ADI extended, but no evaluation), 1980 (T), 1982 (R), 1983 (T)

2013 (T,R), 2015 (R)
2015 (T, R)
1992 (T,R), 1993 (R), 2009 (T), 2012 (R)
2014 (T,R)
1986 (R), 1987 ( T and corr. to 1986 report), 1989 (R), 1990 (R), 1992 (R), 2006 (T), 2007 (R)

1984 (T,R), 1986 (R), 1988 (R), 2007 (T), 2008 (R), 2015 (R)

1970 (T,R), 1973 (T,R), 1974 (R), 1975 (R), 1977 (T), 1978 (T,R), 1980 (T), 1981 (T), 1982 (R), 1983 (R), 1985 (R), 1988 (T), 1989 (T), 1991 (T,R), 1992 (R), 1994 (T), 2005 (T,R)

1979 (T,R), 1981 (T,R), 1982 (R), 1983 (R), 1984 (R), 1985 (R), 1986 (R), 1987 (corr. to 1986 evaluation), 1988 (R), 1990 (R), 2006 (T), 2008 (R), 2009 (R), 2011 (R)

Cyproconazole (239)
Cyprodinil (207)

Cyromazine (169)

2,4-D (020)

Daminozide (104)
DDT (021)

Deltamethrin (135)

Demeton (092)
Demeton-S-methyl (073)

Demeton-S-methylsulfon (164)

Dialifos (098)
Diazinon (022)

1,2-Dibromoethane (023)

Dicamba (240)
Dichlobenil (274)
Dicloran (083)
Dichlorfluanid (082)

1,2-Dichloroethane (024)
Dichlorvos (025)

Dicloran (083)
Dicofol (026)

Dieldrin (001)

2010 (T,R), 2013 (R)
2003 (T,R), 2004 (corr. to 2003 report), 2013 (R), 2015 (R)

1990 (T,R), 1991 (corr. to 1990 R evaluation), 1992 (R), 2006 (T), 2007 (R), 2012 (R)

1970 (T,R), 1971 (T,R), 1974 (T,R), 1975 (T,R), 1980 (R), 1985 (R), 1986 (R), 1987 (corr. to 1986 report, Annex I), 1996 (T), 1997 (E), 1998 (R), 2001 (R)

1977 (T,R), 1983 (T), 1989 (T,R), 1991 (T)
1965 (T), 1966 (T,R), 1967 (T,R), 1968 (T,R), 1969 (T,R), 1978 (R), 1979 (T), 1980 (T), 1983 (T), 1984 (T), 1993 (R), 1994 (R), 1996 (R)

1980 (T,R), 1981 (T,R), 1982 (T,R), 1984 (R), 1985 (R), 1986 (R), 1987 (R), 1988 (R), 1990 (R), 1992 (R), 2000 (T), 2002 (R)

1965 (T), 1967 (R), 1975 (R), 1982 (T)
1973 (T,R), 1979 (R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R), 1998 (R)

1973 (T,R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R)

1976 (T,R), 1982 (T), 1985 (R)
1965 (T), 1966 (T), 1967 (R), 1968 (T,R), 1970 (T,R), 1975 (R), 1979 (R), 1993 (T,R), 1994 (R), 1996 (R), 1999 (R), 2001 (T), 2006 (T,R), 2016 (T)

1965 (T,R), 1966 (T,R), 1967 (R), 1968 (R), 1971 (R), 1979 (R), 1985 (R)

2010 (T,R), 2011 (R), 2012 (R), 2013 (R)
2014 (T,R)
2003 (R)
1969 (T,R), 1974 (T,R), 1977 (T,R), 1979 (T,R), 1981 (R), 1982 (R), 1983 (T,R), 1985 (R)

1965 (T,R), 1967 (R), 1971 (R), 1979 (R), 1985 (R)
1965 (T,R), 1966 (T,R), 1967 (T,R), 1969 (R), 1970 (T,R), 1974 (R), 1977 (T), 1993 (T,R), 2011 (T), 2012 (R)

1974 (T,R), 1977 (T,R), 1998 (T,R)
1968 (T,R), 1970 (R), 1974 (R), 1992 (T,R), 1994 (R), 2011 (T), 2012 (R)

1965 (T), 1966 (T,R), 1967 (T,R), 1968 (R), 1969 (R), 1970 (T,R), 1974 (R), 1975 (R), 1977 (T), 1990 (R), 1992 (R)

Difenoconazole (224)
Diflubenzuron (130)

Dimethenamid-P (214)
Dimethipin (151)

Dimethoate (027)

Dimethomorph (225)
Dimethrin
Dinocap (087)

Dinotefuran (255)
Dioxathion (028)
Diphenyl (029)
Diphenylamine (030)

Diquat (031)

Disulfoton (074)

Dithianon (180)

Dithiocarbamates (105)

4,6-Dinitro-ortho-cresol (DNOC)
Dodine (084)

Edifenphos (099)
Emamectin benzoate (247)
Endosulfan (032)

Endrin (033)

2007 (T,R), 2010 (R), 2013 (R), 2015 (R)
1981 (T,R), 1983 (R), 1984 (T,R), 1985 (T,R), 1988 (R), 2001 (T), 2002 (R), 2011 (R)

2005 (T,R)
1985 (T,R), 1987 (T,R), 1988 (T,R), 1999 (T), 2001 (R), 2004 (T)

1965 (T), 1966 (T), 1967 (T,R), 1970 (R), 1973 (R in evaluation of formothion), 1977 (R), 1978 (R), 1983 (R) 1984 (T,R), 1986 (R), 1987 (T,R), 1988 (R), 1990 (R), 1991 (corr. to 1990 evaluation), 1994 (R), 1996 (T), 1998 (R), 2003 (T,R), 2004 (corr. to 2003 report), 2006 (R), 2008 (R)

2007 (T,R), 2014 (R)
1965 (T)
1969 (T,R), 1974 (T,R), 1989 (T,R), 1992 (R), 1998 (R), 1999 (R), 2000 (T), 2001 (R)

2012 (T,R)
1968 (T,R), 1972 (R)
1966 (T,R), 1967 (T)
1969 (T,R), 1976 (T,R), 1979 (R), 1982 (T), 1984 (T,R), 1998 (T), 2001 (R), 2003 (R), 2008 (R)

1970 (T,R), 1972 (T,R), 1976 (R), 1977 (T,R), 1978 (R), 1994 (R), 2013 (T,R)

1973 (T,R), 1975 (T,R), 1979 ( R ), 1981 (R), 1984 (R), 1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1996 (T), 1998 (R), 2006 (R)

1992 (T,R), 1995 (R), 1996 (corr. to 1995 report), 2010 (T), 2013 (T,R)

1965 (T), 1967 (T,R), 1970 (T,R), 1983 (R propineb, thiram), 1984 ( R propineb), 1985 ( R ), 1987 ( T thiram), 1988 ( R thiram), 1990 (R), 1991 (corr. to 1990 evaluation), 1992 ( T thiram), 1993 (T,R), 1995 (R), 1996 (T,R ferbam, ziram; R thiram), 2004 (R), 2012 (R), 2014 (R)

1965 (T)
1974 (T,R), 1976 (T,R), 1977 (R), 2000 (T), 2003 (R), 2004 (corr. to 2003 report)

1976 (T,R), 1979 (T,R), 1981 (T,R)
2011 (T,R), 2014 (R)
1965 (T), 1967 (T,R), 1968 (T,R), 1971 (R), 1974 (R), 1975 (R), 1982 (T), 1985 (T,R), 1989 (T,R), 1993 (R), 1998 (T), 2006 (R), 2010 (R)

1965 (T), 1970 (T,R), 1974 (R), 1975 (R), 1990 (R), 1992 (R)

Esfenvalerate (204)
Ethephon (106)

Ethiofencarb (107)

Ethion (034)

Ethoprophos (149)
Ethoxyquin (035)
Ethylene dibromide
Ethylene dichloride
Ethylene oxide
Ethylenethiourea (ETU) (108)

Etofenprox (184)
Etoxazole (241)
Etrimfos (123)

Famoxadone (208)
Fenamidone (264)
Fenamiphos (085)

Fenarimol (192)
Fenbuconazole (197)
Fenbutatin oxide (109)
Fenchlorfos (036)
Fenhexamid (215)
Fenitrothion (037)

Fenpropathrin (185)
Fenpropimorph (188)
Fenpyroximate (193)

Fensulfothion (038)
Fenthion (039)

2002 (T,R)
1977 (T,R), 1978 (T,R), 1983 (R), 1985 (R), 1993 (T), 1994 (R), 1995 (T), 1997 (T), 2002 (T), 2015 (T, R)

1977 (T,R), 1978 (R), 1981 (R), 1982 (T,R), 1983 (R)

1968 (T,R), 1969 (R), 1970 (R), 1972 (T,R), 1975 (R), 1982 (T), 1983 (R), 1985 (T), 1986 (T), 1989 (T), 1990 (T), 1994 (R)

1983 (T), 1984 (R), 1987 (T), 1999 (T), 2004 (R)
1969 (T,R), 1998 (T), 1999 (R), 2005 (T), 2008 (R)
See 1,2-Dibromoethane
See 1,2-Dichloroethane
1965 (T,R), 1968 (T,R), 1971 (R)
1974 (R), 1977 (T,R), 1986 (T,R), 1987 (R), 1988 (T,R), 1990 (R), 1993 (T,R)

1993 (T,R), 2011 (T,R)
2010 (T,R), 2011 (R)
1980 (T,R), 1982 (T,R), 1986 (T,R), 1987 (R), 1988 (R), 1989 (R), 1990 (R)

2003 (T,R)
2013 (T), 2014 (T,R)
1974 (T,R), 1977 (R), 1978 (R), 1980 (R), 1985 (T), 1987 (T), 1997 (T), 1999 (R), 2002 (T), 2006 (R)

1995 (T,R,E), 1996 (R and corr. to 1995 report)
1997 (T,R), 2009 (R), 2012 (T), 2013 (R)
1977 (T,R), 1979 (R), 1992 (T), 1993 (R)
1968 (T,R), 1972 (R), 1983 (R)
2005 (T,R)
1969 (T,R), 1974 (T,R), 1976 (R), 1977 (T,R), 1979 (R), 1982 (T), 1983 ( R ), 1984 (T,R), 1986 (T,R), 1987 ( R and corr. to 1986 R evaluation), 1988 (T), 1989 (R), 2000 (T), 2003 (R), 2004 (R, corr. to 2003 report), 2007 (T,R)

1993 (T,R), 2006 (R), 2012 (T), 2014 (R)
1994 (T), 1995 (R), 1999 (R), 2001 (T), 2004 (T)
1995 (T,R), 1996 (corr. to 1995 report), 1999 (R), 2004 (T), 2007 (T), 2010 (R), 2013 (R)

1972 (T,R), 1982 (T), 1983 (R)
1971 (T,R), 1975 (T,R), 1977 (R), 1978 (T,R), 1979 (T), 1980 (T), 1983 (R), 1989 (R),

Fentin compounds (040)

Fenvalerate (119)

Ferbam

Fipronil (202)
Fipronil-desulfinyl
Flonicamid (282)
Flubendiamide (242)
Flucythrinate (152)

Fludioxonil (211)
Fluensulfone (265)
Flufenoxuron (275)
Flumethrin (195)
Fluopicolide (235)
Fluopyram (243)
Flupyradifurone (285)
Flusilazole (165)

Flutolanil (205)
Flutriafol (248)
Fluxapyroxad (256)
Folpet (041)

Formothion (042)

Glufosinate-ammonium (175)

Glyphosate (158)

Guazatine (114)

1995 (T,R,E), 1996 (corr. to 1995 report), 1997 (T), 2000 (R)

1965 (T), 1970 (T,R), 1972 (R), 1986 (R), 1991 (T,R), 1993 (R), 1994 (R)

1979 (T,R), 1981 (T,R), 1982 (T), 1984 (T,R), 1985 (R), 1986 (T,R), 1987 ( R and corr. to 1986 report), 1988 (R), 1990 (R), 1991 (corr. to 1990 R evaluation), 2012 (T,R)

See Dithiocarbamates, 1965 (T), 1967 (T,R), 1996 (T,R)

1997 (T), 2000 (T), 2001 (R)
1997 (T)
2015 (T,R)
2010 (T,R)
1985 (T,R), 1987 (R), 1988 (R), 1989 (R), 1990 (R), 1993 (R)

2004 (T,R), 2006 (R), 2010 (R), 2012 (R), 2013 (R)
2013 (T), 2014 (T,R)
2014 (T,R)
1996 (T,R)
2009 (T,R), 2014 (R)
2010 (T,R), 2012 (R), 2014 (R), 2015 (R)
2015 (T)
1989 (T,R), 1990 (R), 1991 (R), 1993 (R), 1995 (T), 2007 (T,R)

2002 (T,R), 2013 (R)
2011 (T,R), 2015 (R)
2012 (T,R), 2015 (R)
1969 (T,R), 1973 (T), 1974 (R), 1982 (T), 1984 (T,R), 1986 (T), 1987 (R), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1993 (T,R), 1994 (R), 1995 (T), 1997 (R), 1998 (R), 1999 (R) , 2002 (T), 2004 (T), 2007 (T)

1969 (T,R), 1972 (R), 1973 (T,R), 1978 (R), 1998 (R)

1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1998 (R), 1999 (T,R), 2012 (T,R), 2014 (R)

1986 (T,R), 1987 ( R and corr. to 1986 report), 1988 (R), 1994 (R), 1997 (T,R), 2004 (T), 2005 (R), 2011 (T,R), 2013 (R), 2016 (T)

1978 (T,R), 1980 (R), 1997 (T,R)

| Haloxyfop (194) | 1995 (T,R), 1996 ( R and corr. to 1995 report), 2001 (R), 2006 (T), 2009 (R) |
| :---: | :---: |
| Heptachlor (043) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (R), } 1968 \text { (R), } 1969 \text { (R), } \\ & 1970 \text { (T,R), } 1974 \text { (R), } 1975 \text { (R), } 1977 \text { (R), } 1987 \text { (R), } \\ & 1991 \text { (T,R), } 1992 \text { (corr. to } 1991 \text { report, Annex I), } \\ & 1993 \text { (R), } 1994 \text { (R) } \end{aligned}$ |
| Hexachlorobenzene (044) | $\begin{aligned} & 1969 \text { (T,R), } 1973 \text { (T,R), } 1974 \text { (T,R), } 1978 \text { (T), } \\ & 1985 \text { (R) } \end{aligned}$ |
| Hexaconazole (170) | 1990 (T,R), 1991 ( R and corr. to 1990 R evaluation), 1993 (R) |
| Hexythiazox (176) | 1991 (T,R), 1994 (R), 1998 (R), 2008 (T), 2009 (R) |
| Hydrogen cyanide (045) | 1965 (T,R) |
| Hydrogen phosphide (046) | $\begin{aligned} & 1965 \text { (T,R), } 1966 \text { (T,R), } 1967 \text { (R), } 1969 \text { (R), } \\ & 1971 \text { (R) } \end{aligned}$ |
| Imazalil (110) | $\begin{aligned} & 1977 \text { (T,R), } 1980(\mathrm{~T}, \mathrm{R}), 1984(\mathrm{~T}, \mathrm{R}), 1985(\mathrm{~T}, \mathrm{R}), \\ & 1986(\mathrm{~T}), 1988(\mathrm{R}), 1989(\mathrm{R}), 1991(\mathrm{~T}), 1994(\mathrm{R}), \\ & 2000(\mathrm{~T}), 2001(\mathrm{~T}), 2005(\mathrm{~T}) \end{aligned}$ |
| Imazamox (276) | 2014 (T,R) |
| Imazapic (266) | 2013 (T,R), 2015 (R) |
| Imazapyr (267) | 2013 (T,R), 2015 (R) |
| Imidacloprid (206) | $\begin{aligned} & 2001 \text { (T), } 2002 \text { (R), } 2006 \text { (R), } 2008 \text { (R), } 2012 \text { (R), } \\ & 2015 \text { (R) } \end{aligned}$ |
| Indoxacarb (216) | 2005 (T,R), 2007 (R), 2009 (R), 2012 (R), 2013 (R) |
| Iprodione (111) | $\begin{aligned} & 1977 \text { (T,R), } 1980 \text { (R), } 1992 \text { (T), } 1994 \text { (R), } 1995 \text { (T), } \\ & 2001 \text { (R) } \end{aligned}$ |
| Isofenphos (131) | $\begin{aligned} & 1981 \text { (T,R), } 1982 \text { (T,R), } 1984 \text { (R), } 1985 \text { (R), } \\ & 1986(T, R), 1988(R), 1992(R) \end{aligned}$ |
| Isopyrazam (249) | 2011 (T,R) |
| Isoxaflutole (268) | 2013 (T,R) |
| Kresoxim-methyl (199) | 1998 (T,R), 2001 (R) |
| Lead arsenate | 1965 (T), 1968 (T,R) |
| Leptophos (088) | 1974 (T,R), 1975 (T,R), 1978 (T,R) |
| Lindane (048) | 1965 (T), 1966 (T,R), 1967 (R), 1968 (R), 1969 (R), 1970 (T,R, published as Annex VI to 1971 evaluations), 1973 (T,R), 1974 (R), 1975 (R), 1977 (T,R), 1978 (R), 1979 (R), 1989 (T,R), 1997 (T), 2002 (T), 2003 (R), 2004 (corr. to 2003 report), 2015 (R) |
| Lufenuron (286) | 2015 (T, R) |
| Malathion (049) | 1965 (T), 1966 (T,R), 1967 (corr. to 1966 R evaluation), 1968 (R), 1969 (R), 1970 (R), 1973 (R), 1975 (R), 1977 (R), 1984 (R), 1997 (T), 1999 (R), |

Maleic hydrazide (102)

Mancozeb (050)

Mandipropamid (231)
Maneb

MCPA (257)
Mecarbam (124)

Meptyldinocap (244)
Mesotrione (277)
Metaflumizone (236)
Metalaxyl (138)

Metalaxyl -M (212)
Methacrifos (125)

Methamidophos (100)

Methidathion (051)

Methiocarb (132)

Methomyl (094)

Methoprene (147)

Methoxychlor
Methoxyfenozide (209)

Methyl bromide (052)
Metrafenone (278)
Metiram (186)
Mevinphos (053)
MGK 264

2000 (R), 2003 (T), 2004 (R), 2005 (R), 2008 (R), 2013 (R), 2016 (T)

1976 (T,R), 1977 (T,R), 1980 (T), 1984 (T,R), 1996 (T), 1998 (R)

1967 (T,R), 1970 (T,R), 1974 (R), 1977 (R), 1980 (T,R), 1993 (T,R)

2008 (T,R), 2013 (R)
See Dithiocarbamates, 1965 (T), 1967 (T,R), 1987 (T), 1993 (T,R)

2012 (T,R)
1980 (T,R), 1983 (T,R), 1985 (T,R), 1986 (T,R), 1987 (R)

2010 (T,R)
2014 (T,R)
2009 (T,R)
1982 (T,R), 1984 (R), 1985 (R), 1986 (R), 1987 (R), 1989 (R), 1990 (R), 1992 (R), 1995 (R)

2002 (T), 2004 (R)
1980 (T,R), 1982 (T), 1986 (T), 1988 (T), 1990 (T,R), 1992 (R)

1976 (T,R), 1979 (R), 1981 (R), 1982 (T,R), 1984 (R), 1985 (T), 1989 (R), 1990 (T,R), 1994 (R), 1996 (R), 1997 (R), 2002 (T), 2003 (R), 2004 (R, corr. to 2003 report)

1972 (T,R), 1975 (T,R), 1979 (R), 1992 (T,R), 1994 (R), 1997 (T)

1981 (T,R), 1983 (T,R), 1984 (T), 1985 (T), 1986 (R), 1987 (T,R), 1988 (R), 1998 (T), 1999 (R), 2005 (R)

1975 (R), 1976 (R), 1977 (R), 1978 (R), 1986 (T,R), 1987 (R), 1988 (R), 1989 (T,R), 1990 (R), 1991 (R), 2001 (T,R), 2004 (R), 2008 (R)

1984 (T,R), 1986 (R), 1987 (T and corr. to 1986 report), 1988 (R), 1989 (R), 2001 (T), 2005 (R)

1965 (T), 1977 (T)
2003 (T,R), 2004 (corr. to 2003 report), 2006 (R), 2009 (R), 2012 (R)

See Bromomethane
2014 (T,R)
1993 (T), 1995 (R)
1965 (T), 1972 (T,R), 1996 (T), 1997 (E,R), 2000 (R) 1967 (T,R)

| Monocrotophos (054) | $\begin{aligned} & 1972 \text { (T,R), } 1975 \text { (T,R), } 1991 \text { (T,R), } 1993 \text { (T), } \\ & 1994 \text { (R) } \end{aligned}$ |
| :---: | :---: |
| Myclobutanil (181) | $\begin{aligned} & 1992 \text { (T,R), } 1997 \text { (R), } 1998 \text { (R), (2001 (R)), } 2014 \\ & (T, R) \end{aligned}$ |
| Nabam | See Dithiocarbamates, 1965 (T), 1976 (T,R) |
| Nitrofen (140) | 1983 (T,R) |
| Novaluron (217) | 2005 (T,R), 2010 (R) |
| Omethoate (055) | $\begin{aligned} & 1971 \text { (T,R), } 1975 \text { (T,R), } 1978 \text { (T,R), } 1979 \text { (T), } \\ & 1981 \text { (T,R), } 1984(\mathrm{R}), 1985 \text { (T), } 1986 \text { (R), } 1987 \text { (R), } \\ & 1988 \text { (R), } 1990 \text { (R), } 1998 \text { (R) } \end{aligned}$ |
| Organomercury compounds | 1965 (T), 1966 (T,R), 1967 (T,R) |
| Oxamyl (126) | $\begin{aligned} & 1980 \text { (T,R), } 1983 \text { (R), } 1984 \text { (T), } 1985 \text { (T,R), } \\ & 1986 \text { (R), } 2002(\mathrm{~T}, \mathrm{R}) \end{aligned}$ |
| Oxydemeton-methyl (166) | 1965 (T, as demeton-S-methyl sulfoxide), 1967 (T), 1968 (R), 1973 (T,R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R), 1998 (R), 1999 (corr. to 1992 report), 2002 (T), 2004 (R) |
| Oxythioquinox | See Chinomethionat |
| Paclobutrazol (161) | 1988 (T,R), 1989 (R) |
| Paraquat (057) | $\begin{aligned} & 1970 \text { (T,R), } 1972 \text { (T,R), } 1976 \text { (T,R), } 1978 \text { (R), } \\ & 1981 \text { (R), } 1982 \text { (T), } 1985 \text { (T), } 1986 \text { (T), } 2003 \text { (T), } \\ & 2004 \text { (R), } 2009 \text { (R) } \end{aligned}$ |
| Parathion (058) | $\begin{aligned} & 1965 \text { (T), } 1967 \text { (T,R), } 1969 \text { (R), } 1970 \text { (R), } 1984 \text { (R), } \\ & 1991 \text { (R), } 1995 \text { (T,R), } 1997 \text { (R), } 2000 \text { (R) } \end{aligned}$ |
| Parathion-methyl (059) | $\begin{aligned} & 1965(\mathrm{~T}), 1968(\mathrm{~T}, \mathrm{R}), 1972(\mathrm{R}), 1975 \\ & 1978(\mathrm{~T}, \mathrm{R}), \quad 1979 \quad(\mathrm{~T}), \mathrm{R}), 1980(\mathrm{~T}), 1982(\mathrm{~T}), \\ & 1984(\mathrm{~T}, \mathrm{R}), 1991(\mathrm{R}), 1992(\mathrm{R}), 1994(\mathrm{R}), 1995(\mathrm{~T}), \\ & 2000(\mathrm{R}), 2003(\mathrm{R}) \end{aligned}$ |
| Penconazole (182) | 1992 (T,R), 1995 (R), 2015 (T) |
| Penthiopyrad (253) | 2011 (T), 2012 (R), 2013 (R) |
| Permethrin (120) | 1979 (T,R), 1980 (R), 1981 (T,R), 1982 (R), 1983 (R), 1984 (R), 1985 (R), 1986 (T,R), 1987 (T), 1988 (R), 1989 (R), 1991 (R), 1992 (corr. to 1991 report), 1999 (T) |
| 2-Phenylphenol (056) | $\begin{aligned} & 1969(\mathrm{~T}, \mathrm{R}), \quad 1975 \quad(\mathrm{R}), 1983 \quad(\mathrm{~T}), 1985 \quad(\mathrm{~T}, \mathrm{R}), \\ & 1989(\mathrm{~T}), 1990(\mathrm{~T}, \mathrm{R}), 1999(\mathrm{~T}, \mathrm{R}), 2002(\mathrm{R}) \end{aligned}$ |
| Phenothrin (127) | $\begin{aligned} & 1979 \text { (R), } 1980 \text { (T,R), } 1982 \text { (T), } 1984 \text { (T), } 1987 \text { (R), } \\ & 1988 \text { (T,R) } \end{aligned}$ |
| Phenthoate (128) | 1980 (T,R), 1981 (R), 1984 (T) |
| Phorate (112) | $\begin{aligned} & 1977 \text { (T,R), } 1982 \text { (T), } 1983 \text { (T), } 1984 \text { (R), } 1985 \text { (T), } \\ & 1990 \text { (R), } 1991 \text { (R), } 1992 \text { (R), } 1993 \text { (T), } 1994 \text { (T), } \\ & 1996 \text { (T), } 2004 \text { (T), } 2005 \text { (R), } 2012 \text { (R), } 2014 \text { (R) } \end{aligned}$ |
| Phosalone (060) | $\begin{aligned} & 1972 \text { (T,R), } 1975 \text { (R), } 1976 \text { (R), } 1993 \text { (T), } 1994 \text { (R), } \\ & 1997 \text { (T), } 1999 \text { (R), } 2001 \text { (T) } \end{aligned}$ |


| Phosmet (103) | 1976 (R), 1977 (corr. to 1976 R evaluation), 1978 (T,R), 1979 (T,R), 1981 (R), 1984 (R), 1985 (R), 1986 (R), 1987 ( R and corr. to 1986 R evaluation), 1988 (R), 1994 (T), 1997 (R), 1998 (T), 2002 (R), 2003 (R), 2007 (R) |
| :---: | :---: |
| Phosphine | See Hydrogen phosphide |
| Phosphamidon (061) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T), } 1968 \text { (T,R), } 1969 \text { (R), } 1972 \text { (R), } \\ & 1974 \text { (R), } 1982 \text { (T), } 1985 \text { (T), } 1986 \text { (T) } \end{aligned}$ |
| Phoxim (141) | $\begin{aligned} & 1982 \text { (T), } 1983 \text { (R), } 1984 \text { (T,R), } 1986 \text { (R), } 1987 \text { (R), } \\ & 1988 \text { (R) } \end{aligned}$ |
| Picoxystrobin (258) | 2012 (T,R), 2013 (R) |
| Piperonyl butoxide (062) | 1965 (T,R), 1966 (T,R), 1967 (R), 1969 (R), 1972 (T,R), 1992 (T,R), 1995 (T), 2001 (R), 2002 (R) |
| Pirimicarb (101) | $\begin{aligned} & 1976 \text { (T,R), } 1978 \text { (T,R), } 1979 \text { (R), } 1981 \text { (T,R), } \\ & 1982 \text { (T), } 1985 \text { (R), } 2004 \text { (T), } 2006 \text { (R) } \end{aligned}$ |
| Pirimiphos-methyl (086) | 1974 (T,R), 1976 (T,R), 1977 (R), 1979 (R), 1983 (R), 1985 (R), 1992 (T), 1994 (R), 2003 (R), 2004 ( R , corr. to 2003 report), 2006 (T) |
| Prochloraz (142) | 1983 (T,R), 1985 (R), 1987 (R), 1988 (R), 1989 (R), $1990(\mathrm{R}), 1991$ (corr. to 1990 report, Annex I, and R evaluation), 1992 (R), 2001 (T), 2004 (R), 2009 (R) |
| Procymidone(136) | 1981 (R), 1982 (T), 1989 (T,R), 1990 (R), 1991 (corr. to 1990 Annex I), 1993 (R), 1998 (R), 2007 (T) |
| Profenofos (171) | $\begin{aligned} & 1990 \text { (T,R), } 1992 \text { (R), } 1994 \text { (R), } 1995 \text { (R), } 2007 \text { (T), } \\ & 2008 \text { (R), } 2011 \text { (R) } \end{aligned}$ |
| Propamocarb (148) | $\begin{aligned} & 1984 \text { (T,R), } 1986 \text { (T,R), } 1987 \text { (R), } 2005 \text { (T), } \\ & 2006 \text { (R), } 2014 \text { (R) } \end{aligned}$ |
| Propargite (113) | $\begin{aligned} & 1977 \text { (T,R), } 1978 \text { (R), } 1979 \text { (R), } 1980 \text { (T,R), } \\ & 1982 \text { (T,R), } 1999 \text { (T), } 2002 \text { (R), } 2006 \text { (R) } \end{aligned}$ |
| Propham (183) | 1965 (T), 1992 (T,R) |
| Propiconazole (160) | $\begin{aligned} & 1987 \text { (T,R), } 1991 \text { (R), } 1994 \text { (R), } 2004 \text { (T), } 2006 \text { (R), } \\ & 2007 \text { (R), } 2013 \text { (R), } 2014 \text { (R), } 2015 \text { (R) } \end{aligned}$ |
| Propineb | $\begin{array}{lllll} 1977 \text { (T,R), } 1980 \text { (T), } 1983 \text { (T), } 1984 \text { (R), } \\ 1985(T, R), 1993(T, R), 2004(R) \end{array}$ |
| Propoxur (075) | $\begin{aligned} & 1973 \text { (T,R), } 1977 \text { (R), } 1981 \text { (R), } 1983 \text { (R), } 1989 \text { (T), } \\ & 1991 \text { (R), } 1996 \text { (R) } \end{aligned}$ |
| Propylene oxide (250) | 2011 (T,R) |
| Propylenethiourea (PTU, 150) | 1993 (T,R), 1994 (R), 1999 (T) |
| Prothioconazole (232) | 2008 (T,R), 2009 (R), 2014 (R) |
| Pymetrozine (279) | 2014 (T,R) |
| Pyraclostrobin (210) | $\begin{aligned} & 2003 \text { (T), } 2004 \text { (R), } 2006 \text { (R), } 2011 \text { (R), } 2012 \text { (R), } \\ & 2014 \text { (R) } \end{aligned}$ |
| Pyrazophos (153) | 1985 (T,R), 1987 (R), 1992 (T,R), 1993 (R) |

Pyrethrins (063)

Pyrimethanil (226)
Pyriproxyfen (200)
Quinclorac (287)
Quinoxyfen (223)
Quintozene (064)

Saflufenacil (251)
Sedaxane (259)
Spices
Spinetoram (233)
Spinosad (203)
Spirodiclofen (237)
Spirotetramat (234)
Sulfoxaflor (252)
Sulfuryl fluoride (218)
2,4,5-T (121)
Tebuconazole (189)

Tebufenozide (196)

Tecnazine (115)

Teflubenzuron (190)
Temephos
Terbufos (167)
Thiabendazole (065)

Thiacloprid (223)
Thiamethoxam (245)
Thiodicarb (154)

Thiometon (076)

1965 (T), 1966 (T,R), 1967 (R), 1968 (R), 1969 (R), 1970 (T), 1972 (T,R), 1974 (R), 1999 (T), 2000 (R), 2003 (T,R), 2005 (R)

2007 (T,R), 2013 (R)
1999 (R,T), 2000 (R), 2001 (T)
2015 (T, R)
2006 (T,R)
1969 (T,R), 1973 (T,R), 1974 (R), 1975 (T,R), 1976 (Annex I, corr. to 1975 R evaluation), 1977 (T,R), 1995 (T,R), 1998 (R)

2011 (T,R)
2012 (T,R), 2014 (R)
2004 (R), 2005 (R), 2007 (R), 2010 (R), 2015 (R)
2008 (T,R), 2012 (R)
2001 (T,R), 2004 (R), 2008 (R), 2011 (R)
2009 (T,R)
2008 (T,R), 2011 (R), 2012 (R), 2013 (R), 2015 (R)
2011 (T,R), 2013 (R), 2014 (R)
2005 (T,R)
1970 (T,R), 1979 (T,R), 1981 (T)
1994 (T,R), 1996 (corr. to Annex II of 1995 report), 1997 (R), 2008 (R), 2010 (T), 2011 (R), 2015 (R)

1996 (T,R), 1997 (R), 1999 (R), 2001 (T,R), 2003 (T)

1974 (T,R), 1978 (T,R), 1981 (R), 1983 (T), 1987 (R), 1989 (R), 1994 (T,R)

1994 (T), 1996 (R)
2006 (T)
1989 (T,R), 1990 (T,R), 2003 (T), 2005 (R)
1970 (T,R), 1971 (R), 1972 (R), 1975 (R), 1977 (T,R), 1979 (R), 1981 (R), 1997 (R), 2000 (R), 2006 (T,R)

2006 (T,R)
2010 (T,R), 2011 (R), 2012 (R), 2014 (R)
1985 (T,R), 1986 (T), 1987 (R), 1988 (R), 2000 (T),
2001 (R)
1969 (T,R), 1973 (T,R), 1976 (R), 1979 (T,R), 1988 (R)

| Thiophanate-methyl (077) | $\begin{aligned} & 1973 \text { (T,R), } 1975 \text { (T,R), } 1977 \text { (T), } 1978 \text { (R), } \\ & 1988 \text { (R), } 2002(\mathrm{R}), 1990(\mathrm{R}), 1994 \text { (R), } 1995(\mathrm{~T}, \mathrm{E}), \\ & 1998(\mathrm{~T}, \mathrm{R}), 2006(\mathrm{~T}) \end{aligned}$ |
| :---: | :---: |
| Thiram (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1970 (T,R), 1974 (T), 1977 (T), 1983 (R), 1984 (R), 1985 (T,R), 1987 (T), 1988 (R), 1989 (R), 1992 (T), 1996 (R) |
| Tolclofos-methyl (191) | 1994 (T,R), 1996 (corr. to Annex II of 1995 report) |
| Tolfenpyrad (269) | 2013 (T) |
| Tolylfluanid (162) | $\begin{aligned} & 1988 \text { (T,R), } 1990 \text { (R), } 1991 \text { (corr. to } 1990 \text { report), } \\ & 2002(\mathrm{~T}, \mathrm{R}), 2003 \text { (R) } \end{aligned}$ |
| Toxaphene | See Camphechlor |
| Triadimefon (133) | 1979 (R), 1981 (T,R), 1983 (T,R), 1984 (R), 1985 (T,R), 1986 (R), 1987 ( R and corr. to 1986 R evaluation), 1988 (R), 1989 (R), 1992 (R), 1995 (R), 2004 (T), 2007 (R) |
| Triadimenol (168) | $\begin{aligned} & 1989 \text { (T,R), } 1992 \text { (R), } 1995 \text { (R), } 2004 \text { (T), } 2007 \text { (R), } \\ & 2014 \text { (R) } \end{aligned}$ |
| Triazolylalanine | 1989 (T,R) |
| Triazophos (143) | 1982 (T), 1983 (R), 1984 (corr. to 1983 report, Annex I), 1986 (T,R), 1990 (R), 1991 (T and corr. to 1990 R evaluation), 1992 (R), 1993 (T,R), 2002 (T), 2007 (R), 2010 (R), 2013 (R) |
| Trichlorfon (066) | 1971 (T,R), 1975 (T,R), 1978 (T,R), 1987 (R) |
| Trichloronat | 1971 (T,R) |
| Trichloroethylene | 1968 (R) |
| Tricyclohexyltin hydroxide | See Cyhexatin |
| Trifloxystrobin (213) | 2004 (T,R), 2012 (R), 2015 (R) |
| Triflumizole (270) | 2013 (T,R) |
| Triforine (116) | 1977 (T), 1978 (T,R), 1997 (T), 2004 (R), 2014 (T,R) |
| Trinexapac-ethyl (271) | 2013 (T,R) |
| Triphenyltin compounds | See Fentin compounds |
| Vamidothion (078) | $\begin{aligned} & 1973 \text { (T,R), } 1982 \text { (T), } 1985 \text { (T,R), } 1987 \text { (R), } \\ & 1988 \text { (T), } 1990(\mathrm{R}), 1992(\mathrm{R}) \end{aligned}$ |
| Vinclozolin (159) | 1986 (T,R), 1987 ( R and corr. to 1986 report and R evaluation), 1988 (T,R), 1989 (R), 1990 (R), 1992 (R), 1995 (T) |
| Zineb (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1993 (T) |
| Ziram (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1996 (T,R) |
| Zoxamide (227) | 2007 (T,R), 2009 (R) |

DIAZINON
ANNEX 3: INTERNATIONAL ESTIMATED DAILY INTAKES OF PESTICIDE RESIDUES
International Estimated Daily Intake (IEDI) ADI $=0-0.003 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | Expr <br> as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} /$ day |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G01 <br> diet | $\begin{aligned} & \text { G01 } \\ & \text { intake } \end{aligned}$ | $\begin{gathered} \text { G02 } \\ \text { diet } \end{gathered}$ | $\begin{aligned} & \text { G02 } \\ & \text { intake } \end{aligned}$ | $\begin{gathered} \text { G03 } \\ \text { diet } \end{gathered}$ | $\begin{aligned} & \text { G03 } \\ & \text { intake } \end{aligned}$ | G04 <br> diet | G04 <br> intake | $\begin{gathered} \text { G05 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G06 } \\ \text { diet } \end{gathered}$ | $\begin{aligned} & \text { G06 } \\ & \text { intake } \end{aligned}$ |
| $\begin{aligned} & \text { FP } \\ & 0009 \end{aligned}$ | Pome fruit, raw (incl apple juice, excl cider) | RAC | 0.04 | 19.69 | 0.79 | 38.08 | 1.52 | 3.43 | 0.14 | 32.35 | 1.29 | 7.98 | 0.32 | 64.35 | 2.57 |
| JF $0226$ | Apple juice, single strength (incl concentrated) | PP | 0.0004 | 0.32 | 0.00 | 3.07 | 0.00 | 0.10 | 0.00 | 5.00 | 0.00 | 0.29 | 0.00 | 5.57 | 0.00 |
| $\begin{aligned} & \text { FS } \\ & 0013 \end{aligned}$ | Cherries, raw | RAC | 1 | 0.92 | 0.92 | 9.15 | 9.15 | 0.10 | 0.10 | 0.61 | 0.61 | 0.10 | 0.10 | 6.64 | 6.64 |
| $\begin{aligned} & \text { FS } \\ & 0014 \end{aligned}$ | Plums, raw (incl dried plums, incl Chinese jujube) | RAC | 1 | 2.67 | 2.67 | 8.77 | 8.77 | 0.10 | 0.10 | 3.03 | 3.03 | 0.70 | 0.70 | 4.34 | 4.34 |
| $\begin{aligned} & \text { DF } \\ & 0014 \end{aligned}$ | Plum, dried (prunes) | PP | 2 | 0.10 | 0.20 | 0.10 | 0.20 | 0.10 | 0.20 | 0.18 | 0.36 | 0.10 | 0.20 | 0.10 | 0.20 |
| - | Peaches and nectarines, raw | RAC | 0.2 | 2.87 | 0.57 | 2.21 | 0.44 | 0.15 | 0.03 | 5.94 | 1.19 | 1.47 | 0.29 | 15.66 | 3.13 |
| $\begin{aligned} & \text { FB } \\ & 0264 \end{aligned}$ | Blackberries, raw | RAC | 0.1 | 0.35 | 0.04 | 0.11 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 1.23 | 0.12 |
| $\begin{aligned} & \text { FB } \\ & 0266 \end{aligned}$ | Dewberries, incl boysen- \& loganberry, raw | RAC | 0.1 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 |
| $\begin{aligned} & \text { FB } \\ & 0272 \end{aligned}$ | Raspberries, red, black, raw | RAC | 0.2 | 0.10 | 0.02 | 0.93 | 0.19 | 0.10 | 0.02 | 0.10 | 0.02 | 0.10 | 0.02 | 0.10 | 0.02 |
| $\begin{aligned} & \text { FB } \\ & 0021 \end{aligned}$ | Currants, red, black, white, raw | RAC | 0.2 | 0.10 | 0.02 | 0.74 | 0.15 | 0.10 | 0.02 | 0.10 | 0.02 | 0.10 | 0.02 | 0.10 | 0.02 |
| $\begin{aligned} & \text { FB } \\ & 0265 \end{aligned}$ | Cranberries, raw | RAC | 0.05 | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 |
| FB | Strawberry, raw | RAC | 0.1 | 0.70 | 0.07 | 2.01 | 0.20 | 0.10 | 0.01 | 1.36 | 0.14 | 0.37 | 0.04 | 2.53 | 0.25 |


Annex 3

| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | G01 <br> diet | $\begin{aligned} & \text { G01 } \\ & \text { intake } \end{aligned}$ | $\begin{gathered} \text { G02 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G02 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G03 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G03 } \\ \text { intake } \end{gathered}$ | G04 <br> diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | G05 <br> diet | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | G06 <br> diet | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
| $\begin{aligned} & \text { VL } \\ & 0482 \end{aligned}$ | Lettuce, head, raw | RAC | 0.5 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf, raw | RAC | 0.5 | 0.53 | 0.27 | 0.36 | 0.18 | 0.16 | 0.08 | 6.21 | 3.11 | 1.90 | 0.95 | 6.05 | 3.03 |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach, raw | RAC | 0.5 | 0.74 | 0.37 | 0.22 | 0.11 | 0.10 | 0.05 | 0.91 | 0.46 | 0.10 | 0.05 | 2.92 | 1.46 |
| - | Chinese cabbage flowering stalk, raw | RAC | 0.05 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { VP } \\ & 0062 \end{aligned}$ | Beans, green, without pods, raw: beans except broad bean \& soya bean (i.e. immature seeds only) (Phaseolus spp) | RAC | 0.2 | 1.56 | 0.31 | 0.60 | 0.12 | 0.49 | 0.10 | 1.18 | 0.24 | 0.90 | 0.18 | 7.79 | 1.56 |
| $\begin{aligned} & \text { VP } \\ & 0064 \end{aligned}$ | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) | RAC | 0.2 | 1.97 | 0.39 | 0.51 | 0.10 | 0.10 | 0.02 | 0.79 | 0.16 | 3.68 | 0.74 | 3.80 | 0.76 |
| $\begin{aligned} & \text { VR } \\ & 0494 \end{aligned}$ | Radish roots, raw | RAC | 0.1 | 2.31 | 0.23 | 4.09 | 0.41 | 2.53 | 0.25 | 6.15 | 0.62 | 5.88 | 0.59 | 2.97 | 0.30 |
| $\begin{aligned} & \text { VR } \\ & 0577 \end{aligned}$ | Carrots, raw | RAC | 0.5 | 9.51 | 4.76 | 30.78 | 15.39 | 0.37 | 0.19 | 8.75 | 4.38 | 2.80 | 1.40 | 6.10 | 3.05 |
| $\begin{aligned} & \text { VR } \\ & 0589 \end{aligned}$ | Potato, raw (incl flour, incl frozen, incl starch, incl tapioca) | RAC | 0.01 | 59.74 | 0.60 | 316.14 | 3.16 | 9.78 | 0.10 | 60.26 | 0.60 | 54.12 | 0.54 | 119.82 | 1.20 |
| $\begin{aligned} & \text { VR } \\ & 0589 \end{aligned}$ | Potato, raw (incl flour, incl frozen, incl tapioca, excl starch) | RAC | 0 | 59.60 | 0.00 | 316.10 | 0.00 | 9.77 | 0.00 | 59.59 | 0.00 | 54.12 | 0.00 | 119.82 | 0.00 |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet, raw (incl sugar) | RAC | 0.1 | 0.13 | 0.01 | NC | - | 0.10 | 0.01 | 0.66 | 0.07 | 0.47 | 0.05 | 88.94 | 8.89 |
| $\begin{aligned} & \text { GC } \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0 | 29.81 | 0.00 | 44.77 | 0.00 | 108.95 | 0.00 | 52.37 | 0.00 | 60.28 | 0.00 | 75.69 | 0.00 |
| $\begin{aligned} & \mathrm{TN} \\ & 0660 \\ & \hline \end{aligned}$ | Almonds, nutmeat | RAC | 0.05 | 1.38 | 0.07 | 0.10 | 0.01 | 0.10 | 0.01 | 1.00 | 0.05 | 0.10 | 0.01 | 0.81 | 0.04 |
| $\begin{aligned} & \mathrm{TN} \\ & 0678 \end{aligned}$ | Walnuts, nutmeat | RAC | 0 | 0.23 | 0.00 | 1.49 | 0.00 | 0.10 | 0.00 | 0.33 | 0.00 | 0.10 | 0.00 | 2.06 | 0.00 |


Annex 3
International Estimated Daily Intake (IEDI)
$\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg}$ bw


| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | G07 <br> intake | G08 <br> diet | G08 intake | G09 <br> diet | G09 <br> intake | G10 <br> diet | G10 intake | G11 diet | G11 intake | G12 <br> diet | G12 intake |
| $\begin{aligned} & \text { VB } \\ & 0041 \end{aligned}$ | Cabbages, head, raw | RAC | 0.01 | 8.97 | 0.09 | 27.12 | 0.27 | 1.44 | 0.01 | 24.96 | 0.25 | 4.55 | 0.05 | 11.23 | 0.11 |
| $\begin{aligned} & \text { VB } \\ & 0400 \end{aligned}$ | Broccoli, raw | RAC | 0.5 | 4.24 | 2.12 | 1.76 | 0.88 | NC | - | 0.51 | 0.26 | 3.79 | 1.90 | 0.26 | 0.13 |
| VB <br> 0405 | Kohlrabi, raw | RAC | 0.2 | NC | - | 3.25 | 0.65 | NC | - | NC | - | 0.10 | 0.02 | 0.36 | 0.07 |
| $\begin{aligned} & \text { VC } \\ & 0046 \end{aligned}$ | Melons, raw (excl watermelons) | RAC | 0.2 | 9.20 | 1.84 | 11.95 | 2.39 | 14.63 | 2.93 | 8.99 | 1.80 | 7.86 | 1.57 | 2.46 | 0.49 |
| $\begin{aligned} & \mathrm{VC} \\ & 0424 \end{aligned}$ | Cucumber, raw | RAC | 0.1 | 6.72 | 0.67 | 11.03 | 1.10 | 32.10 | 3.21 | 15.10 | 1.51 | 4.05 | 0.41 | 9.57 | 0.96 |
| $\begin{aligned} & \text { VC } \\ & 0431 \end{aligned}$ | Squash, summer, raw (= courgette, zucchini) | RAC | 0.05 | NC | - | NC | - | 5.48 | 0.27 | NC | - | NC | - | 1.03 | 0.05 |
| $\begin{aligned} & \text { VO } \\ & 0445 \end{aligned}$ | Peppers, sweet, raw (incl dried) | RAC | 0.05 | 0.82 | 0.04 | 1.53 | 0.08 | 10.85 | 0.54 | 4.59 | 0.23 | 1.84 | 0.09 | 2.00 | 0.10 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.02 | 11.43 | 0.23 | 3.71 | 0.07 | 0.74 | 0.01 | 13.63 | 0.27 | 3.07 | 0.06 | 1.50 | 0.03 |
| $\begin{aligned} & \text { VO } \\ & 0448 \end{aligned}$ | Tomato, raw (incl juice, incl paste, incl canned) | RAC | 0.12 | 64.74 | 7.77 | 68.31 | 8.20 | 36.05 | 4.33 | 82.09 | 9.85 | 54.50 | 6.54 | 11.69 | 1.40 |
| $\begin{aligned} & \text { VL } \\ & 0480 \end{aligned}$ | Kale, raw (i.e. collards) (i.e. Brassica) | RAC | 0.05 | NC | - | NC | - | 14.54 | 0.73 | NC | - | NC | - | 2.32 | 0.12 |
| $\begin{aligned} & \text { VL } \\ & 0482 \end{aligned}$ | Lettuce, head, raw | RAC | 0.5 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf, raw | RAC | 0.5 | 14.50 | 7.25 | 11.76 | 5.88 | 13.14 | 6.57 | 19.50 | 9.75 | 4.81 | 2.41 | 2.23 | 1.12 |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach, raw | RAC | 0.5 | 2.20 | 1.10 | 1.76 | 0.88 | 13.38 | 6.69 | 2.94 | 1.47 | 5.53 | 2.77 | 0.10 | 0.05 |
| - | Chinese cabbage flowering stalk, raw | RAC | 0.05 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP | Beans, green, without pods, raw: beans | RAC | 0.2 | 2.21 | 0.44 | 5.25 | 1.05 | 4.17 | 0.83 | 1.61 | 0.32 | 16.95 | 3.39 | 0.17 | 0.03 |

Annex 3

| DIAZINON (22) International Estimated Daily Intake (IEDI) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G10 diet | $\begin{aligned} & \text { G10 } \\ & \text { intake } \end{aligned}$ | G11 <br> diet | $\begin{aligned} & \text { G11 } \\ & \text { intake } \end{aligned}$ | G12 diet | $\begin{gathered} \text { G12 } \\ \text { intake } \end{gathered}$ |
|  |  |  |  | G07 <br> diet | G07 <br> intake | G08 <br> diet | G08 <br> intake | G09 <br> diet | G09 <br> intake |  |  |  |  |  |  |
| 0062 | except broad bean $\&$ soya bean (i.e. immature seeds only) (Phaseolus spp) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { VP } \\ & 0064 \end{aligned}$ | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) | RAC | 0.2 | 10.72 | 2.14 | 1.99 | 0.40 | 2.72 | 0.54 | 4.26 | 0.85 | 4.23 | 0.85 | NC | - |
| $\begin{aligned} & \text { VR } \\ & 0494 \end{aligned}$ | Radish roots, raw | RAC | 0.1 | 3.83 | 0.38 | 11.99 | 1.20 | NC | - | 5.26 | 0.53 | 2.19 | 0.22 | 4.37 | 0.44 |
| $\begin{aligned} & \text { VR } \\ & 0577 \end{aligned}$ | Carrots, raw | RAC | 0.5 | 26.26 | 13.13 | 27.13 | 13.57 | 10.07 | 5.04 | 16.49 | 8.25 | 44.69 | 22.35 | 8.75 | 4.38 |
| $\begin{aligned} & \text { VR } \\ & 0589 \end{aligned}$ | Potato, raw (incl flour, incl frozen, incl starch, incl tapioca) | RAC | 0.01 | 225.03 | 2.25 | 234.24 | 2.34 | 71.48 | 0.71 | 177.55 | 1.78 | 234.55 | 2.35 | 37.71 | 0.38 |
| $\begin{aligned} & \text { VR } \\ & 0589 \end{aligned}$ | Potato, raw (incl flour, incl frozen, incl tapioca, excl starch) | RAC | 0 | 225.03 | 0.00 | 226.35 | 0.00 | 71.26 | 0.00 | 173.36 | 0.00 | 234.55 | 0.00 | 37.71 | 0.00 |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet, raw (incl sugar) | RAC | 0.1 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | NC | - |
| $\begin{aligned} & \text { GC } \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0 | 18.51 | 0.00 | 26.18 | 0.00 | 26.04 | 0.00 | 39.99 | 0.00 | 7.36 | 0.00 | 64.58 | 0.00 |
| $\begin{aligned} & \text { TN } \\ & 0660 \end{aligned}$ | Almonds, nutmeat | RAC | 0.05 | 0.81 | 0.04 | 2.21 | 0.11 | 0.10 | 0.01 | 1.02 | 0.05 | 1.47 | 0.07 | NC | - |
| $\begin{aligned} & \text { TN } \\ & 0678 \end{aligned}$ | Walnuts, nutmeat | RAC | 0 | 0.34 | 0.00 | 0.84 | 0.00 | 0.28 | 0.00 | 0.39 | 0.00 | 0.45 | 0.00 | NC | - |
| DH <br> 1100 | Hops, dry | RAC | 0.5 | NC | - | NC | - | 0.10 | 0.05 | 0.10 | 0.05 | NC | - | NC | - |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $-80 \%$ as muscle | RAC | 0.02 | 112.02 | 2.24 | 120.71 | 2.41 | 63.46 | 1.27 | 88.99 | 1.78 | 96.24 | 1.92 | 41.02 | 0.82 |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $-20 \%$ as fat | RAC | 0.3 | 28.01 | 8.40 | 30.18 | 9.05 | 15.86 | 4.76 | 22.25 | 6.67 | 24.06 | 7.22 | 10.25 | 3.08 |
| MO | Edible offal (mammalian), raw | RAC | 0.01 | 15.17 | 0.15 | 5.19 | 0.05 | 6.30 | 0.06 | 6.78 | 0.07 | 3.32 | 0.03 | 3.17 | 0.03 |


| DIAZINON (22) International Estimated Daily Intake (IEDI) ADI = 0-0.003 mg/k |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | $\begin{aligned} & \mathrm{STMR} \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} /$ day |  |  | $\begin{gathered} \text { G10 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G10 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G11 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G11 } \\ \text { intake } \end{gathered}$ | G12diet | $\begin{aligned} & \text { G12 } \\ & \text { intake } \end{aligned}$ |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G07 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G08 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G08 } \\ \text { intake } \end{gathered}$ | G09 | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ |  |  |  |  |  |  |
| 0105 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { ML } \\ & 0106 \end{aligned}$ | Milks, raw or skimmed (incl dairy products) | RAC | 0.02 | 388.92 | 7.78 | 335.88 | 6.72 | 49.15 | 0.98 | 331.25 | 6.63 | 468.56 | 9.37 | 245.45 | 4.91 |
| $\begin{aligned} & \text { PM } \\ & 0110 \end{aligned}$ | Poultry meat, raw (incl prepared) | RAC | 0.02 | 73.76 | 1.48 | 53.86 | 1.08 | 23.98 | 0.48 | 87.12 | 1.74 | 53.38 | 1.07 | 84.45 | 1.69 |
| $\begin{aligned} & \text { PO } \\ & 0111 \end{aligned}$ | Poultry edible offal, raw (incl prepared) | RAC | 0.02 | 0.33 | 0.01 | 0.72 | 0.01 | 0.27 | 0.01 | 0.35 | 0.01 | 0.80 | 0.02 | NC | - |
| $\begin{aligned} & \text { PE } \\ & 0840 \end{aligned}$ | Chicken eggs, raw (incl dried) | RAC | 0.02 | 25.49 | 0.51 | 29.46 | 0.59 | 23.08 | 0.46 | 33.03 | 0.66 | 36.39 | 0.73 | 8.89 | 0.18 |
| Total intake ( $\mu \mathrm{g} /$ person $)=$ |  |  |  |  | 77.3 |  | 79.6 |  | 53.1 |  | 74.7 |  | 84.2 |  | 30.8 |
|  | Body weight per region (kg bw) = |  |  |  | 60 |  | 60 |  | 55 |  | 60 |  | 60 |  | 60 |
|  | $\operatorname{ADI}(\mu \mathrm{g} /$ person $)=$ |  |  |  | 180 |  | 180 |  | 165 |  | 180 |  | 180 |  | 180 |
|  | \%ADI $=$ |  |  |  | $43.0 \%$ |  | 44.2\% |  | 32.2\% |  | $41.5 \%$ |  | $46.8 \%$ |  | 17.1\% |
|  | Rounded \%ADI = |  |  |  | 40\% |  | 40\% |  | 30\% |  | 40\% |  | 50\% |  | 20\% |

DIAZINON (22)
International Estimated Daily Intake (IEDI)
$\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg}$ bw


| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} / \mathrm{day}$ |  |  |  |  |  |  |
|  |  |  |  | G13 <br> diet | G13 <br> intake | G14 <br> diet | G14 <br> intake | G15 <br> diet | G15 <br> intake | G16 <br> diet | G16 <br> intake | G17 <br> diet | G17 <br> intake |
| VC 0431 | Squash, summer, raw (= courgette, zucchini) | RAC | 0.05 | 0.10 | 0.01 | 1.01 | 0.05 | NC | - | 1.91 | 0.10 | NC | - |
| VO 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.05 | 5.49 | 0.27 | 10.57 | 0.53 | 8.84 | 0.44 | 0.91 | 0.05 | NC | - |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.02 | 3.63 | 0.07 | 20.50 | 0.41 | 8.78 | 0.18 | 0.10 | 0.00 | 0.17 | 0.00 |
| VO 0448 | Tomato, raw (incl juice, incl paste, incl canned) | RAC | 0.12 | 15.50 | 1.86 | 5.78 | 0.69 | 71.52 | 8.58 | 2.00 | 0.24 | 12.50 | 1.50 |
| VL 0480 | Kale, raw (i.e. collards) (i.e. Brassica) | RAC | 0.05 | 0.79 | 0.04 | 0.62 | 0.03 | NC | - | 0.10 | 0.01 | NC | - |
| VL 0482 | Lettuce, head, raw | RAC | 0.5 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VL 0483 | Lettuce, leaf, raw | RAC | 0.5 | 0.29 | 0.15 | 0.10 | 0.05 | 6.71 | 3.36 | 0.10 | 0.05 | NC | - |
| VL 0502 | Spinach, raw | RAC | 0.5 | 0.17 | 0.09 | 0.10 | 0.05 | 0.81 | 0.41 | 0.10 | 0.05 | NC | - |
| - | Chinese cabbage flowering stalk, raw | RAC | 0.05 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP 0062 | Beans, green, without pods, raw: beans except broad bean \& soya bean (i.e. immature seeds only) (Phaseolus spp) | RAC | 0.2 | 0.30 | 0.06 | 3.13 | 0.63 | 4.11 | 0.82 | 0.10 | 0.02 | NC | - |
| VP 0064 | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) | RAC | 0.2 | 0.21 | 0.04 | 0.10 | 0.02 | 5.51 | 1.10 | 0.10 | 0.02 | NC | - |
| VR 0494 | Radish roots, raw | RAC | 0.1 | 3.96 | 0.40 | 2.86 | 0.29 | 3.30 | 0.33 | 2.67 | 0.27 | 5.34 | 0.53 |
| VR 0577 | Carrots, raw | RAC | 0.5 | 2.07 | 1.04 | 3.00 | 1.50 | 25.29 | 12.65 | 0.10 | 0.05 | NC | - |
| VR 0589 | Potato, raw (incl flour, incl frozen, incl starch, incl tapioca) | RAC | 0.01 | 23.96 | 0.24 | 13.56 | 0.14 | 213.41 | 2.13 | 104.35 | 1.04 | 8.56 | 0.09 |
| VR 0589 | Potato, raw (incl flour, incl frozen, incl tapioca, excl starch) | RAC | 0 | 23.96 | 0.00 | 13.54 | 0.00 | 213.41 | 0.00 | 104.35 | 0.00 | 8.56 | 0.00 |
| VR 0596 | Sugar beet, raw (incl sugar) | RAC | 0.1 | 3.93 | 0.39 | 1.68 | 0.17 | NC | - | NC | - | 36.12 | 3.61 |
| GC 0645 | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0 | 116.66 | 0.00 | 10.52 | 0.00 | 38.46 | 0.00 | 76.60 | 0.00 | 34.44 | 0.00 |
| TN 0660 | Almonds, nutmeat | RAC | 0.05 | 0.10 | 0.01 | 0.10 | 0.01 | 0.61 | 0.03 | 0.10 | 0.01 | NC | - |

Annex 3

| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | $\begin{aligned} & \text { G16 } \\ & \text { diet } \end{aligned}$ | G16 <br> intake | G17 <br> diet | $\begin{gathered} \text { G17 } \\ \text { intake } \end{gathered}$ |
|  |  |  |  | G13 <br> diet | $\begin{aligned} & \text { G13 } \\ & \text { intake } \end{aligned}$ | G14 <br> diet | G14 intake | G15 diet | $\begin{gathered} \text { G15 } \\ \text { intake } \end{gathered}$ |  |  |  |  |
| TN 0678 | Walnuts, nutmeat | RAC | 0 | 0.10 | 0.00 | 0.10 | 0.00 | 0.81 | 0.00 | 0.10 | 0.00 | NC | - |
| DH 1100 | Hops, dry | RAC | 0.5 | NC | - | NC | - | 0.10 | 0.05 | NC | - | NC | - |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $80 \%$ as muscle | RAC | 0.02 | 23.34 | 0.47 | 40.71 | 0.81 | 97.15 | 1.94 | 18.06 | 0.36 | 57.71 | 1.15 |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $20 \%$ as fat | RAC | 0.3 | 5.84 | 1.75 | 10.18 | 3.05 | 24.29 | 7.29 | 4.52 | 1.35 | 14.43 | 4.33 |
| $\begin{aligned} & \text { MO } \\ & 0105 \end{aligned}$ | Edible offal (mammalian), raw | RAC | 0.01 | 4.64 | 0.05 | 1.97 | 0.02 | 10.01 | 0.10 | 3.27 | 0.03 | 3.98 | 0.04 |
| ML 0106 | Milks, raw or skimmed (incl dairy products) | RAC | 0.02 | 108.75 | 2.18 | 70.31 | 1.41 | 436.11 | 8.72 | 61.55 | 1.23 | 79.09 | 1.58 |
| PM 0110 | Poultry meat, raw (incl prepared) | RAC | 0.02 | 3.92 | 0.08 | 12.03 | 0.24 | 57.07 | 1.14 | 5.03 | 0.10 | 55.56 | 1.11 |
| PO 0111 | Poultry edible offal, raw (incl prepared) | RAC | 0.02 | 0.10 | 0.00 | 0.70 | 0.01 | 0.97 | 0.02 | 0.10 | 0.00 | NC | - |
| PE 0840 | Chicken eggs, raw (incl dried) | RAC | 0.02 | 3.83 | 0.08 | 4.27 | 0.09 | 26.38 | 0.53 | 1.13 | 0.02 | 7.39 | 0.15 |
| Total intake ( $\mu \mathrm{g} /$ person $)=$ |  |  |  |  | 12.8 |  | 13.9 |  | 85.5 |  | 6.2 |  | 23.8 |
|  | Body weight per region (kg bw) = |  |  |  | 60 |  | 60 |  | 60 |  | 60 |  | 60 |
|  | $\operatorname{ADI}(\mu \mathrm{g} /$ person $)=$ |  |  |  | 180 |  | 180 |  | 180 |  | 180 |  | 180 |
|  | $\% \mathrm{ADI}=$ |  |  |  | 7.1\% |  | 7.7\% |  | 47.5\% |  | 3.4\% |  | 13.2\% |
|  | Rounded \%ADI = |  |  |  | 7\% |  | 8\% |  | 50\% |  | 3\% |  | 10\% |

GLYPHOSATE

Annex 3
International Estimated Daily Intake (IEDI)
ADI $=0-1 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | $\begin{aligned} & \text { Expr } \\ & \text { as } \end{aligned}$ | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\begin{gathered} \text { G01 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G01 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G02 } \\ \text { diet } \end{gathered}$ | G02 <br> intake | $\begin{aligned} & \text { G03 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G03 } \\ \text { intake } \end{gathered}$ | G04 <br> diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G05 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G06 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
|  | flour) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| - | Maize starch | PP | 0.04 | 0.10 | 0.00 | NC | - | 0.10 | 0.00 | 2.29 | 0.09 | 0.10 | 0.00 | 0.11 | 0.00 |
| $\begin{aligned} & \text { OR } \\ & 0645 \end{aligned}$ | Maize oil | PP | 0.04 | 0.96 | 0.04 | 0.85 | 0.03 | 0.29 | 0.01 | 5.42 | 0.22 | 0.42 | 0.02 | 2.10 | 0.08 |
| $\begin{aligned} & \mathrm{GC} \\ & 0646 \end{aligned}$ | Millet, raw (incl flour, incl beer) | RAC | 3.7 | 1.46 | 5.40 | 2.32 | 8.58 | 5.84 | 21.61 | 0.89 | 3.29 | 16.17 | 59.83 | 0.10 | 0.37 |
| $\begin{aligned} & \mathrm{GC} \\ & 0647 \end{aligned}$ | Oats, raw (incl rolled) | RAC | 3.7 | 0.10 | 0.37 | 7.05 | 26.09 | 0.10 | 0.37 | 1.71 | 6.33 | 0.96 | 3.55 | 0.10 | 0.37 |
| $\begin{aligned} & \mathrm{GC} \\ & 0648 \end{aligned}$ | Quinoa, raw | RAC | 3.7 | NC | - | NC | - | NC | - | NC | - | 0.10 | 0.37 | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0650 \end{aligned}$ | Rye, raw (incl flour) | RAC | 3.7 | 0.13 | 0.48 | 19.38 | 71.71 | 0.10 | 0.37 | 0.12 | 0.44 | 0.10 | 0.37 | 2.15 | 7.96 |
| $\begin{aligned} & \mathrm{GC} \\ & 0651 \end{aligned}$ | Sorghum, raw (incl beer, excl flour) | RAC | 3.7 | NC | - | 0.10 | 0.37 | 3.34 | 12.36 | 0.10 | 0.37 | NC | - | NC | - |
| - | Sorghum, flour (white flour and wholemeal flour) | PP | 1.5 | 3.91 | 5.87 | NC | - | 11.62 | 17.43 | 14.24 | 21.36 | 9.87 | 14.81 | 2.62 | 3.93 |
| $\begin{aligned} & \text { GC } \\ & 0653 \end{aligned}$ | Triticale, raw (incl flour) | RAC | 3.7 | NC | - | NC | - | NC | - | 0.10 | 0.37 | 0.39 | 1.44 | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0654 \end{aligned}$ | Wheat, raw (incl bulgur, incl fermented beverages, incl germ, incl wholemeal bread, excl white flour products, excl white bread) | RAC | 3.7 | 0.10 | 0.37 | 1.13 | 4.18 | 0.10 | 0.37 | 0.10 | 0.37 | 0.74 | 2.74 | 0.10 | 0.37 |
| CF 0654 | Wheat, bran | PP | 1.8 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| CP 1211 | Wheat, white bread | PP | 0.11 | 0.25 | 0.03 | 0.63 | 0.07 | 0.12 | 0.01 | 0.43 | 0.05 | 1.39 | 0.15 | 0.22 | 0.02 |
| CF 1211 | Wheat, white flour (incl white flour products: starch, gluten, macaroni, pastry) | PP | 0.11 | 301.49 | 33.16 | 269.27 | 29.62 | 30.33 | 3.34 | 222.94 | 24.52 | 136.12 | 14.97 | 343.34 | 37.77 |
| - | Wheat, macaroni, dry | PP | 0.11 | 0.72 | 0.08 | 2.20 | 0.24 | 1.22 | 0.13 | 3.99 | 0.44 | 0.53 | 0.06 | 1.66 | 0.18 |
| - | Wheat, pastry, baked | PP | 0.11 | 1.21 | 0.13 | 3.13 | 0.34 | 1.05 | 0.12 | 4.02 | 0.44 | 0.60 | 0.07 | 1.40 | 0.15 |




| GLYPHOSATE (158) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \\ \hline \end{gathered}$ | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} /$ day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | G07 <br> intake | G08 <br> diet | G08 <br> intake | $\begin{gathered} \text { G09 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ | G10 <br> diet | G10 intake | G11 <br> diet | G1 1 <br> intake | G12 <br> diet | G12 <br> intake |
| $\begin{aligned} & \text { FI } \\ & 0327 \end{aligned}$ | Banana, raw (incl plantains) (incl dried) | RAC | 0.05 | 25.14 | 1.26 | 23.37 | 1.17 | 23.06 | 1.15 | 23.40 | 1.17 | 18.44 | 0.92 | 39.29 | 1.96 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.325 | 11.43 | 3.71 | 3.71 | 1.21 | 0.74 | 0.24 | 13.63 | 4.43 | 3.07 | 1.00 | 1.50 | 0.49 |
| $\begin{aligned} & \text { VD } \\ & 0071 \end{aligned}$ | Beans, dry, raw (Phaseolus spp) | RAC | 0.17 | 1.51 | 0.26 | 1.50 | 0.26 | 1.90 | 0.32 | 5.11 | 0.87 | 1.36 | 0.23 | 23.43 | 3.98 |
| $\begin{aligned} & \text { VD } \\ & 0072 \end{aligned}$ | Peas, dry, raw (Pisum spp, Vigna spp): garden peas \& field peas \& cow peas | RAC | 0.5 | 3.80 | 1.90 | 1.25 | 0.63 | 1.06 | 0.53 | 2.33 | 1.17 | 2.70 | 1.35 | 3.83 | 1.92 |
| $\begin{aligned} & \text { VD } \\ & 0533 \end{aligned}$ | Lentil, dry, raw (Ervum lens) | RAC | 0.5 | 0.95 | 0.48 | 1.18 | 0.59 | 0.40 | 0.20 | 0.96 | 0.48 | 0.71 | 0.36 | 1.28 | 0.64 |
| $\begin{aligned} & \text { VD } \\ & 0541 \end{aligned}$ | Soya bean, dry, raw (incl flour, incl paste, incl curd, incl sauce, excl oil) | RAC | 5 | 0.47 | 2.35 | 0.77 | 3.85 | 9.12 | 45.60 | 8.05 | 40.25 | 0.10 | 0.50 | 6.06 | 30.30 |
| $\begin{aligned} & \text { OR } \\ & 0541 \end{aligned}$ | Soya oil, refined | PP | 0.1 | 19.06 | 1.91 | 21.06 | 2.11 | 5.94 | 0.59 | 33.78 | 3.38 | 40.05 | 4.01 | 13.39 | 1.34 |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet, raw | RAC | 3.4 | 0.10 | 0.34 | NC | - | 0.10 | 0.34 | 0.10 | 0.34 | NC | - | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0640 \end{aligned}$ | Barley, raw (incl malt extract, incl pot \& pearled, incl flour \& grits, incl beer, incl malt) | RAC | 3.7 | 36.18 | 133.87 | 53.45 | 197.77 | 9.39 | 34.74 | 35.25 | 130.43 | 46.68 | 172.72 | 15.92 | 58.90 |
| $\begin{aligned} & \text { GC } \\ & 0641 \end{aligned}$ | Buckwheat, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | 0.79 | 2.92 | 0.18 | 0.67 | 0.35 | 1.30 | NC | - | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl beer, incl germ, excl flour, excl oil, excl starch) | RAC | 0.12 | 0.10 | 0.01 | 9.93 | 1.19 | 1.40 | 0.17 | 10.26 | 1.23 | 0.33 | 0.04 | 0.10 | 0.01 |
| $\begin{aligned} & \mathrm{GC} \\ & 0656 \end{aligned}$ | Popcorn (i.e. maize used for preparation of popcorn) | RAC | 0.12 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { CF } \\ & 1255 \end{aligned}$ | Maize, flour (white flour and wholemeal flour) | PP | 0.13 | 14.27 | 1.86 | 12.86 | 1.67 | 19.71 | 2.56 | 12.55 | 1.63 | 4.21 | 0.55 | 52.30 | 6.80 |

Annex 3

| GLYPHOSATE (158) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} /$ day |  |  | G10 <br> diet | G10 intake | G11 <br> diet | G11 intake | G12 <br> diet | $\begin{gathered} \text { G12 } \\ \text { intake } \end{gathered}$ |
|  |  |  |  | G07 <br> diet | G07 <br> intake | G08 <br> diet | G08 <br> intake | G09 <br> diet | G09 <br> intake |  |  |  |  |  |  |
| - | Maize starch | PP | 0.04 | NC | - | NC | - | 0.19 | 0.01 | 7.13 | 0.29 | NC | - | NC | - |
| $\begin{aligned} & \text { OR } \\ & 0645 \end{aligned}$ | Maize oil | PP | 0.04 | 0.90 | 0.04 | 0.47 | 0.02 | 0.15 | 0.01 | 3.01 | 0.12 | 1.86 | 0.07 | 0.36 | 0.01 |
| $\begin{aligned} & \text { GC } \\ & 0646 \end{aligned}$ | Millet, raw (incl flour, incl beer) | RAC | 3.7 | 0.10 | 0.37 | 0.16 | 0.59 | 1.75 | 6.48 | 0.69 | 2.55 | NC | - | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0647 \end{aligned}$ | Oats, raw (incl rolled) | RAC | 3.7 | 7.50 | 27.75 | 6.26 | 23.16 | 0.15 | 0.56 | 4.87 | 18.02 | 3.16 | 11.69 | 2.98 | 11.03 |
| $\begin{aligned} & \mathrm{GC} \\ & 0648 \end{aligned}$ | Quinoa, raw | RAC | 3.7 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { GC } \\ & 0650 \end{aligned}$ | Rye, raw (incl flour) | RAC | 3.7 | 3.21 | 11.88 | 35.38 | 130.91 | 0.21 | 0.78 | 6.50 | 24.05 | 1.49 | 5.51 | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0651 \end{aligned}$ | Sorghum, raw (incl beer, excl flour) | RAC | 3.7 | NC | - | NC | - | 0.10 | 0.37 | 1.15 | 4.26 | NC | - | 7.12 | 26.34 |
| - | Sorghum, flour (white flour and wholemeal flour) | PP | 1.5 | NC | - | NC | - | 1.29 | 1.94 | 0.10 | 0.15 | NC | - | NC | - |
| $\begin{aligned} & \text { GC } \\ & 0653 \end{aligned}$ | Triticale, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | 0.17 | 0.63 | 0.29 | 1.07 | 0.10 | 0.37 | NC | - | NC | - |
| $\begin{aligned} & \text { GC } \\ & 0654 \end{aligned}$ | Wheat, raw (incl bulgur, incl fermented beverages, incl germ, incl wholemeal bread, excl white flour products, excl white bread) | RAC | 3.7 | 1.00 | 3.70 | 0.11 | 0.41 | 0.10 | 0.37 | 0.84 | 3.11 | 0.10 | 0.37 | 0.10 | 0.37 |
| $\begin{aligned} & \text { CF } \\ & 0654 \\ & \hline \end{aligned}$ | Wheat, bran | PP | 1.8 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { CP } \\ & 1211 \end{aligned}$ | Wheat, white bread | PP | 0.11 | 1.30 | 0.14 | 0.46 | 0.05 | 0.10 | 0.01 | 0.22 | 0.02 | 2.44 | 0.27 | 0.77 | 0.08 |
| $\begin{aligned} & \text { CF } \\ & 1211 \end{aligned}$ | Wheat, white flour (incl white flour products: starch, gluten, macaroni, pastry) | PP | 0.11 | 199.38 | 21.93 | 193.50 | 21.29 | 106.30 | 11.69 | 185.31 | 20.38 | 171.11 | 18.82 | 132.37 | 14.56 |
| - | Wheat, macaroni, dry | PP | 0.11 | 6.71 | 0.74 | 4.98 | 0.55 | 2.12 | 0.23 | 1.90 | 0.21 | 2.89 | 0.32 | 4.12 | 0.45 |
| - | Wheat, pastry, baked | PP | 0.11 | 7.93 | 0.87 | 0.51 | 0.06 | 0.29 | 0.03 | 2.44 | 0.27 | 1.78 | 0.20 | 8.64 | 0.95 |


| GLYPHOSATE (158) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR$\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G07 } \\ \text { intake } \end{gathered}$ | G08 <br> diet | G08 <br> intake | G09 <br> diet | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ | G10 <br> diet | $\begin{gathered} \text { G10 } \\ \text { intake } \end{gathered}$ | G11 <br> diet | G11 <br> intake | G12 <br> diet | G12 <br> intake |
| - | Fonio, raw (incl flour) | RAC | 3.7 | NC | - | NC | - | 0.10 | 0.37 | NC | - | NC | - | NC | - |
| - | Cereals, NES, raw (including processed): canagua, quihuicha, Job's tears and wild rice | RAC | 3.7 | 6.17 | 22.83 | 3.01 | 11.14 | 0.76 | 2.81 | 3.30 | 12.21 | 3.38 | 12.51 | 15.84 | 58.61 |
| $\begin{aligned} & \text { GS } \\ & 0659 \end{aligned}$ | Sugar cane, raw | RAC | 0.27 | NC | - | NC | - | 4.27 | 1.15 | 0.10 | 0.03 | NC | - | 3.24 | 0.87 |
| - | Sugar cane, molasses | PP | 2.3 | NC | - | NC | - | 0.10 | 0.23 | NC | - | NC | - | NC | - |
| - | Sugar cane, sugar (incl non-centrifugal sugar, incl refined sugar and maltose) | PP | 0.065 | 92.24 | 6.00 | 95.72 | 6.22 | 24.12 | 1.57 | 77.39 | 5.03 | 117.73 | 7.65 | 100.67 | 6.54 |
| $\begin{aligned} & \text { SO } \\ & 0495 \end{aligned}$ | Rape seed, raw | RAC | 3 | NC | - | NC | - | 0.10 | 0.30 | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { OR } \\ & 0495 \end{aligned}$ | Rape seed oil, edible | PP | 0.009 | 12.52 | 0.11 | 7.63 | 0.07 | 3.00 | 0.03 | 6.01 | 0.05 | NC | - | NC | - |
| $\begin{aligned} & \text { SO } \\ & 0691 \end{aligned}$ | Cotton seed, raw | RAC | 5.2 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { OR } \\ & 0691 \end{aligned}$ | Cotton seed oil, edible | PP | 0.52 | 1.68 | 0.87 | 0.66 | 0.34 | 1.13 | 0.59 | 1.18 | 0.61 | 0.89 | 0.46 | 0.37 | 0.19 |
| $\begin{aligned} & \text { SO } \\ & 0702 \end{aligned}$ | Sunflower seed, raw (incl oil) | RAC | 0.395 | 23.40 | 9.24 | 29.33 | 11.59 | 1.24 | 0.49 | 13.85 | 5.47 | 6.48 | 2.56 | 6.91 | 2.73 |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) - $80 \%$ as muscle | RAC | 0.05 | 112.02 | 5.60 | 120.71 | 6.04 | 63.46 | 3.17 | 88.99 | 4.45 | 96.24 | 4.81 | 41.02 | 2.05 |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) - $20 \%$ as fat | RAC | 0.05 | 28.01 | 1.40 | 30.18 | 1.51 | 15.86 | 0.79 | 22.25 | 1.11 | 24.06 | 1.20 | 10.25 | 0.51 |
| $\begin{aligned} & \text { MO } \\ & 0105 \end{aligned}$ | Edible offal (mammalian), raw | RAC | 2.9 | 15.17 | 43.99 | 5.19 | 15.05 | 6.30 | 18.27 | 6.78 | 19.66 | 3.32 | 9.63 | 3.17 | 9.19 |
| $\begin{aligned} & \text { ML } \\ & 0106 \end{aligned}$ | Milks, raw or skimmed (incl dairy products) | RAC | 0 | 388.92 | 0.00 | 335.88 | 0.00 | 49.15 | 0.00 | 331.25 | 0.00 | 468.56 | 0.00 | 245.45 | 0.00 |


GLYPHOSATE (158)
International Estimated Daily Intake (IEDI)
$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw


| Codex code | Commodity description | Expr as | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G16 diet | G16 <br> intake | $\begin{aligned} & \text { G17 } \\ & \text { diet } \end{aligned}$ | $\begin{aligned} & \text { G17 } \\ & \text { intake } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G13 <br> diet | G13 <br> intake | G14 <br> diet | G14 <br> intake | G15 <br> diet | G15 <br> intake |  |  |  |  |
| GC 0650 | Rye, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | 0.10 | 0.37 | 13.95 | 51.62 | 0.10 | 0.37 | 0.88 | 3.26 |
| GC 0651 | Sorghum, raw (incl beer, excl flour) | RAC | 3.7 | 4.73 | 17.50 | NC | - | NC | - | 13.36 | 49.43 | NC | - |
| - | Sorghum, flour (white flour and wholemeal flour) | PP | 1.5 | 75.99 | 113.99 | 1.82 | 2.73 | NC | - | 19.82 | 29.73 | NC | - |
| GC 0653 | Triticale, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | NC | - | NC | - | NC | - | NC | - |
| GC 0654 | Wheat, raw (incl bulgur, incl fermented beverages, incl germ, incl wholemeal bread, excl white flour products, excl white bread) | RAC | 3.7 | 0.10 | 0.37 | 0.10 | 0.37 | 0.10 | 0.37 | 0.10 | 0.37 | 0.97 | 3.59 |
| CF 0654 | Wheat, bran | PP | 1.8 | NC | - | NC | - | NC | - | NC | - | NC | - |
| CP 1211 | Wheat, white bread | PP | 0.11 | 0.43 | 0.05 | 0.41 | 0.05 | 1.56 | 0.17 | 0.11 | 0.01 | 0.10 | 0.01 |
| CF 1211 | Wheat, white flour (incl white flour products: starch, gluten, macaroni, pastry) | PP | 0.11 | 45.21 | 4.97 | 87.37 | 9.61 | 215.61 | 23.72 | 20.42 | 2.25 | 103.67 | 11.40 |
| - | Wheat, macaroni, dry | PP | 0.11 | 0.52 | 0.06 | 0.63 | 0.07 | 2.99 | 0.33 | 0.26 | 0.03 | 5.18 | 0.57 |
| - | Wheat, pastry, baked | PP | 0.11 | 0.51 | 0.06 | 0.51 | 0.06 | 4.36 | 0.48 | 0.67 | 0.07 | 5.32 | 0.59 |
| - | Fonio, raw (incl flour) | RAC | 3.7 | 0.61 | 2.26 | NC | - | NC | - | NC | - | NC | - |
| - | Cereals, NES, raw (including processed): canagua, quihuicha, Job's tears and wild rice | RAC | 3.7 | 17.71 | 65.53 | 2.00 | 7.40 | 9.61 | 35.56 | 0.45 | 1.67 | 4.55 | 16.84 |
| GS 0659 | Sugar cane, raw | RAC | 0.27 | 5.62 | 1.52 | 50.91 | 13.75 | NC | - | 11.04 | 2.98 | 0.10 | 0.03 |
| - | Sugar cane, molasses | PP | 2.3 | NC | - | NC | - | NC | - | NC | - | NC | - |
| - | Sugar cane, sugar (incl non-centrifugal sugar, incl refined sugar and maltose) | PP | 0.065 | 28.13 | 1.83 | 55.38 | 3.60 | 78.09 | 5.08 | 18.04 | 1.17 | 45.60 | 2.96 |
| SO 0495 | Rape seed, raw | RAC | 3 | NC | - | 0.10 | 0.30 | NC | - | NC | - | NC | - |
| OR 0495 | Rape seed oil, edible | PP | 0.009 | 0.10 | 0.00 | 0.10 | 0.00 | 4.62 | 0.04 | 0.10 | 0.00 | NC | - |
| SO 0691 | Cotton seed, raw | RAC | 5.2 | NC | - | NC | - | NC | - | NC | - | NC | - |
| OR 0691 | Cotton seed oil, edible | PP | 0.52 | 1.28 | 0.67 | 0.10 | 0.05 | 0.45 | 0.23 | 0.42 | 0.22 | 0.15 | 0.08 |
| SO 0702 | Sunflower seed, raw (incl oil) | RAC | 0.395 | 0.94 | 0.37 | 0.22 | 0.09 | 32.01 | 12.64 | 12.12 | 4.79 | 0.48 | 0.19 |



| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G04 diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | $\begin{aligned} & \text { G05 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G06 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G01 diet | $\begin{gathered} \text { G01 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G02 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G02 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G03 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G03 } \\ \text { intake } \end{gathered}$ |  |  |  |  |  |  |
| FC 0001 | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 34.91 | 0.70 | 16.51 | 0.33 | 17.23 | 0.34 | 104.48 | 2.09 | 35.57 | 0.71 | 98.49 | 1.97 |
| FP 0226 | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 13.94 | 1.53 | 30.81 | 3.39 | 15.14 | 1.67 | 23.10 | 2.54 | 6.86 | 0.75 | 55.48 | 6.10 |
| FB 0020 | Blueberries, raw | RAC | 2.27 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 |
| FB 0269 | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 16.25 | 2.60 | 28.96 | 4.63 | 2.87 | 0.46 | 24.22 | 3.88 | 9.33 | 1.49 | 68.64 | 10.98 |
| FB 0275 | Strawberry, raw | RAC | 0.25 | 0.70 | 0.18 | 2.01 | 0.50 | 0.10 | 0.03 | 1.36 | 0.34 | 0.37 | 0.09 | 2.53 | 0.63 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 29.36 | 6.75 | 37.50 | 8.63 | 3.56 | 0.82 | 34.78 | 8.00 | 18.81 | 4.33 | 43.38 | 9.98 |
| - | Onions, green, raw | RAC | 0.52 | 2.45 | 1.27 | 1.49 | 0.77 | 1.02 | 0.53 | 2.60 | 1.35 | 0.60 | 0.31 | 2.03 | 1.06 |
| VC 0424 | Cucumber, raw | RAC | 0.02 | 8.01 | 0.16 | 30.66 | 0.61 | 1.45 | 0.03 | 19.84 | 0.40 | 0.27 | 0.01 | 34.92 | 0.70 |
| Vo 0444 | Peppers, chili, raw (incl dried) | RAC | 0.01 | 6.93 | 0.07 | 10.97 | 0.11 | 8.83 | 0.09 | 9.13 | 0.09 | 6.65 | 0.07 | 20.01 | 0.20 |
| Vo 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 4.49 | 0.04 | 6.44 | 0.06 | 7.21 | 0.07 | 5.68 | 0.06 | 9.52 | 0.10 | 8.92 | 0.09 |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 0.14 | 0.00 | 0.94 | 0.01 | 5.70 | 0.06 | 2.61 | 0.03 | 1.94 | 0.02 | 0.22 | 0.00 |
| VO 0448 | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 42.04 | 8.83 | 76.13 | 15.99 | 10.69 | 2.24 | 84.59 | 17.76 | 24.92 | 5.23 | 203.27 | 42.69 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 2.34 | 0.16 | 1.33 | 0.09 | 1.57 | 0.11 | 4.24 | 0.30 | 0.34 | 0.02 | 2.83 | 0.20 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.29 | 0.00 | 0.29 | 0.00 | 0.10 | 0.00 | 0.38 | 0.00 | 0.10 | 0.00 | 0.14 | 0.00 |
| VL 0485 | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | 0.10 | 0.01 | 0.31 | 0.02 | 0.10 | 0.01 | 0.10 | 0.01 | 0.47 | 0.03 | 0.11 | 0.01 |
| VL 0502 | Spinach, raw | RAC | 0.35 | 0.74 | 0.26 | 0.22 | 0.08 | 0.10 | 0.04 | 0.91 | 0.32 | 0.10 | 0.04 | 2.92 | 1.02 |
| VL 0506 | Turnip greens, raw (i.e. Namenia, | RAC | 1.2 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |


Annex 3

| International Estimated Daily Intake (IEDI) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Diets as | rson/day |  | Intake | $\mathrm{g} / \mathrm{perso}$ |  |  |  |  |  |  |  |
| Codex code | Commodity description | Expr <br> as | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | $\begin{aligned} & \text { G07 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G07 } \\ \text { intake } \end{gathered}$ | G08 <br> diet | G08 <br> intake | $\begin{gathered} \text { G09 } \\ \text { diet } \end{gathered}$ | G09 <br> intake | G10 <br> diet | G10 <br> intake | G11 <br> diet | G11 <br> intake | G12 <br> diet | G12 <br> intake |
| $\begin{aligned} & \text { FC } \\ & 0001 \end{aligned}$ | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 114.42 | 2.29 | 62.91 | 1.26 | 26.97 | 0.54 | 96.72 | 1.93 | 96.22 | 1.92 | 563.19 | 11.26 |
| $\begin{aligned} & \text { FP } \\ & 0226 \end{aligned}$ | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 61.44 | 6.76 | 72.81 | 8.01 | 26.84 | 2.95 | 45.18 | 4.97 | 93.28 | 10.26 | 7.78 | 0.86 |
| $\begin{aligned} & \text { FB } \\ & 0020 \end{aligned}$ | Blueberries, raw | RAC | 2.27 | 0.10 | 0.23 | 0.23 | 0.52 | 0.10 | 0.23 | 0.83 | 1.88 | 0.33 | 0.75 | NC | - |
| $\begin{aligned} & \text { FB } \\ & 0269 \end{aligned}$ | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 142.23 | 22.76 | 105.77 | 16.92 | 7.87 | 1.26 | 52.44 | 8.39 | 109.22 | 17.48 | 10.96 | 1.75 |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry, raw | RAC | 0.25 | 4.49 | 1.12 | 5.66 | 1.42 | 0.10 | 0.03 | 6.63 | 1.66 | 5.75 | 1.44 | 0.10 | 0.03 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 19.69 | 4.53 | 29.83 | 6.86 | 24.64 | 5.67 | 31.35 | 7.21 | 9.72 | 2.24 | 12.59 | 2.90 |
| - | Onions, green, raw | RAC | 0.52 | 1.55 | 0.81 | 0.74 | 0.38 | 1.05 | 0.55 | 3.74 | 1.94 | 0.94 | 0.49 | 6.45 | 3.35 |
| $\begin{aligned} & \text { VC } \\ & 0424 \end{aligned}$ | Cucumber, raw | RAC | 0.02 | 6.72 | 0.13 | 11.03 | 0.22 | 32.10 | 0.64 | 15.10 | 0.30 | 4.05 | 0.08 | 9.57 | 0.19 |
| $\begin{aligned} & \text { VO } \\ & 0444 \end{aligned}$ | Peppers, chili, raw (incl dried) | RAC | 0.01 | 6.36 | 0.06 | 15.46 | 0.15 | 10.74 | 0.11 | 7.28 | 0.07 | 8.21 | 0.08 | 3.58 | 0.04 |
| $\begin{aligned} & \text { VO } \\ & 0445 \end{aligned}$ | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 0.82 | 0.01 | 1.53 | 0.02 | 10.85 | 0.11 | 4.59 | 0.05 | 1.84 | 0.02 | 2.00 | 0.02 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 11.43 | 0.11 | 3.71 | 0.04 | 0.74 | 0.01 | 13.63 | 0.14 | 3.07 | 0.03 | 1.50 | 0.02 |
| $\begin{aligned} & \text { VO } \\ & 0448 \end{aligned}$ | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 43.88 | 9.21 | 55.41 | 11.64 | 35.38 | 7.43 | 74.88 | 15.72 | 26.50 | 5.57 | 9.51 | 2.00 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 4.96 | 0.35 | 3.20 | 0.22 | 0.15 | 0.01 | 1.61 | 0.11 | 6.88 | 0.48 | 0.52 | 0.04 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.80 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 | 0.61 | 0.00 | 0.40 | 0.00 | 0.10 | 0.00 |
| VL | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | NC | - | NC | - | NC | - | NC | - | NC | - | 0.13 | 0.01 |


| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G07 } \\ \text { intake } \end{gathered}$ | G08 <br> diet | G08 intake | G09 <br> diet | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ | G10 <br> diet | $\begin{gathered} \text { G10 } \\ \text { intake } \end{gathered}$ | G11 diet | G11 <br> intake | G12 <br> diet | G12 intake |
| 0485 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach, raw | RAC | 0.35 | 2.20 | 0.77 | 1.76 | 0.62 | 13.38 | 4.68 | 2.94 | 1.03 | 5.53 | 1.94 | 0.10 | 0.04 |
| $\begin{aligned} & \text { VL } \\ & 0506 \end{aligned}$ | Turnip greens, raw (i.e. Namenia, Tendergreen) | RAC | 1.2 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { VP } \\ & 0061 \end{aligned}$ | Beans, green, with pods, raw: beans except broad bean \& soya bean (i.e. immature seeds + pods) (Phaseolus spp) | RAC | 0.31 | 5.07 | 1.57 | 0.83 | 0.26 | 0.17 | 0.05 | 3.70 | 1.15 | NC | - | NC | - |
| $\begin{aligned} & \hline \mathrm{VD} \\ & 0071 \end{aligned}$ | Beans, dry, raw (Phaseolus spp) | RAC | 0.36 | 1.51 | 0.54 | 1.50 | 0.54 | 1.90 | 0.68 | 5.11 | 1.84 | 1.36 | 0.49 | 23.43 | 8.43 |
| $\begin{aligned} & \text { VR } \\ & 0506 \end{aligned}$ | Garden turnip, raw | RAC | 0.05 | 5.78 | 0.29 | 15.35 | 0.77 | NC | - | 6.54 | 0.33 | 1.95 | 0.10 | 4.73 | 0.24 |
| VS <br> 0621 | Asparagus | RAC | 0.305 | 0.84 | 0.26 | 2.08 | 0.63 | 7.11 | 2.17 | 1.01 | 0.31 | 1.69 | 0.52 | 0.10 | 0.03 |
| $\begin{aligned} & \hline \mathrm{GC} \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0.01 | 18.51 | 0.19 | 26.18 | 0.26 | 26.04 | 0.26 | 39.99 | 0.40 | 7.36 | 0.07 | 64.58 | 0.65 |
| $\begin{aligned} & \mathrm{GC} \\ & 0651 \end{aligned}$ | Sorghum, raw (incl flour, incl beer) | RAC | 0.235 | NC | - | NC | - | 1.44 | 0.34 | 1.15 | 0.27 | NC | - | 7.12 | 1.67 |
| $\begin{aligned} & \mathrm{GC} \\ & 0654 \end{aligned}$ | Wheat, raw (incl bulgur, incl fermented beverages, excl germ, excl wholemeal bread, excl white flour products, excl white bread) | RAC | 10 | 0.37 | 3.70 | 0.10 | 1.00 | 0.10 | 1.00 | 0.10 | 1.00 | NC | - | 0.10 | 1.00 |
| $\begin{aligned} & \text { CF } \\ & 0654 \end{aligned}$ | Wheat, bran | PP | 25 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { CF } \\ & 1212 \end{aligned}$ | Wheat, wholemeal flour | PP | 7.5 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| CP <br> 1212 | Wheat, wholemeal bread | PP | 1.2 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 |
| CP | Wheat, white bread | PP | 0.2 | 1.30 | 0.26 | 0.46 | 0.09 | 0.10 | 0.02 | 0.22 | 0.04 | 2.44 | 0.49 | 0.77 | 0.15 |

Annex 3


| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |
|  |  |  |  | G13 <br> diet | $\begin{gathered} \text { G13 } \\ \text { intake } \end{gathered}$ | G14 diet | G14 <br> intake | G15 <br> diet | G15 <br> intake | G16 <br> diet | G16 <br> intake | G17 <br> diet | G17 <br> intake |
| FC 0001 | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 21.16 | 0.42 | 2.94 | 0.06 | 58.52 | 1.17 | 0.44 | 0.01 | 5.13 | 0.10 |
| FP 0226 | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 66.71 | 7.34 | 2.19 | 0.24 | 65.63 | 7.22 | 188.34 | 20.72 | 1.38 | 0.15 |
| FB 0020 | Blueberries, raw | RAC | 2.27 | NC | - | NC | - | 0.20 | 0.45 | NC | - | NC | - |
| FB 0269 | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 0.60 | 0.10 | 1.26 | 0.20 | 103.25 | 16.52 | 0.74 | 0.12 | 44.23 | 7.08 |
| FB 0275 | Strawberry, raw | RAC | 0.25 | 0.10 | 0.03 | 0.10 | 0.03 | 3.35 | 0.84 | 0.10 | 0.03 | 0.10 | 0.03 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 9.01 | 2.07 | 20.24 | 4.66 | 30.90 | 7.11 | 9.61 | 2.21 | 2.11 | 0.49 |
| - | Onions, green, raw | RAC | 0.52 | 1.43 | 0.74 | 0.10 | 0.05 | 0.20 | 0.10 | NC | - | 6.30 | 3.28 |
| VC 0424 | Cucumber, raw | RAC | 0.02 | 0.68 | 0.01 | 1.81 | 0.04 | 10.40 | 0.21 | 0.10 | 0.00 | 0.10 | 0.00 |
| VO 0444 | Peppers, chili, raw (incl dried) | RAC | 0.01 | 7.55 | 0.08 | 12.48 | 0.12 | 24.78 | 0.25 | 0.87 | 0.01 | NC | - |
| VO 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 5.49 | 0.05 | 10.57 | 0.11 | 8.84 | 0.09 | 0.91 | 0.01 | NC | - |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 3.63 | 0.04 | 20.50 | 0.21 | 8.78 | 0.09 | 0.10 | 0.00 | 0.17 | 0.00 |
| VO 0448 | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 13.10 | 2.75 | 4.90 | 1.03 | 62.16 | 13.05 | 1.04 | 0.22 | 0.10 | 0.02 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 0.58 | 0.04 | 0.22 | 0.02 | 2.21 | 0.15 | 0.24 | 0.02 | 3.10 | 0.22 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.10 | 0.00 | 0.10 | 0.00 | 0.42 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 |
| VL 0485 | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | NC | - |
| VL 0502 | Spinach, raw | RAC | 0.35 | 0.17 | 0.06 | 0.10 | 0.04 | 0.81 | 0.28 | 0.10 | 0.04 | NC | - |
| VL 0506 | Turnip greens, raw (i.e. Namenia, Tendergreen) | RAC | 1.2 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP 0061 | Beans, green, with pods, raw: beans except broad bean \& soya bean (i.e. immature seeds + pods) (Phaseolus spp) | RAC | 0.31 | NC | - | NC | - | NC | - | NC | - | NC | - |

International Estimated Daily Intake (IEDI)

IESTI

|  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\begin{gathered} 100 \% \\ \text { all } \end{gathered}$ | $\begin{aligned} & 100 \% \\ & \text { gen pop } \end{aligned}$ | $\begin{aligned} & 100 \% \\ & \text { child } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR- } \\ \mathrm{P} \\ \mathrm{mg} / \mathrm{kg} \end{array}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, g/person | Unit weight, edible portion, g | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ |
| FP 0226 | Apple <br> (all commodities) | highest utilization: Total | 0.004 | 0.24 | 1.000 | US | Child, 1-6 <br> yrs | - | 624.45 | 127.0 | 3 | 2a | $\begin{aligned} & 0.04- \\ & 14.05 \end{aligned}$ | 0-50\% | 0-20\% | 0-50\% |
| FP 0227 | Crab-apple (all commodities) | highest utilization: raw with peel | 0 | 0.24 | 1.000 | CN | Gen pop, $>1 \mathrm{yrs}$ | 204 | 488.33 | - | - | - | 0-0 | 0-0\% | 0-0\% | 0-0\% |
| FP 0228 | Loquat (Japanese medlar) <br> (all commodities) | highest utilization: raw without peel | 0 | 0.24 | 1.000 | JP | Gen pop, $>1 \mathrm{yrs}$ | 113 | 326.40 | 49.0 | 3 | 2a | $\begin{array}{r} 0.42- \\ 1.88 \end{array}$ | 1-6\% | 1-6\% | 0-0\% |
| FP 0229 | Medlar | Total |  | 0.24 | 1.000 | - | - | - | - | - | - | - | - | - | - | - |
| FP 0230 | Pear (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0.004 | 0.24 | 1.000 | CN | Child, 1-6 yrs | 413 | 418.33 | 255.0 | 3 | 2a | 0-13.81 | 0-50\% | 0-20\% | 0-50\% |
| FT 0307 | Persimmon, Japanese (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0 | 0.24 | 1.000 | TH | Child, 3-6 yrs | 20 | 264.88 | 227.5 | 3 | 2a | $\begin{array}{r} 3.53- \\ 10.1 \end{array}$ | $\begin{gathered} 10- \\ 30 \% \end{gathered}$ | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ | $\begin{gathered} 30- \\ 30 \% \end{gathered}$ |
| FP 0231 | Quince <br> (all commodities) | highest utilization: Total | 0 | 0.24 | 1.000 | DE | Child, 2-4 yrs | 16 | 26.30 | 301.2 | 3 | 2b | $\begin{array}{r} 1.17- \\ 1.17 \end{array}$ | 4-4\% | 0-0\% | 4-4\% |
| FS 0013 | Cherries (all commodities) | highest utilization: raw | 0 | 0.73 | 1.000 | DE | Child, 2-4 <br> yrs | 24 | 187.50 | 7.2 | NR | 1 | $\begin{array}{r} 0.83- \\ 8.48 \end{array}$ | 3-30\% | 3-30\% | 8-30\% |
| FS 0014 | Plums <br> (all commodities) | highest utilization: raw with peel (incl consumption withou | 0 | $\begin{array}{r} 0.78- \\ \hline 1.9 \end{array}$ | 1.000 | TH | Child, 3-6 yrs | 11 | 376.88 | 93.0 | 3 | 2a | $\begin{aligned} & 3.04- \\ & 25.68 \end{aligned}$ | $\begin{gathered} 10- \\ 90 \% \end{gathered}$ | 4-40\% | $\begin{gathered} 10- \\ 90 \% \end{gathered}$ |


|  |  |  | DIAZINON (22) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | ARfD $=0.03 \mathrm{mg} / \mathrm{kg} \mathrm{bw}(30 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$ |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $\begin{gathered} 100 \% \\ \text { all } \end{gathered}$ |  | $100 \%$ <br> child |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | STMR <br> or STMR$\mathrm{mg} / \mathrm{kg}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, g/person | Unit weight, edible portion, g | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | ARfD rounded | ARfD rounded | ARfD rounded |
|  |  | peel) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FS 0247 | Peach (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0 | 0.2 | 1.000 | JP | Child, 1-6 yrs | 76 | 306.00 | 255.0 | 3 | 2a | $\begin{aligned} & 1.09- \\ & 10.53 \end{aligned}$ | 4-40\% | 2-10\% | 4-40\% |
| $\begin{aligned} & \text { FB } \\ & 0264 \end{aligned}$ | Blackberries (all commodities) | highest utilization: raw with skin | 0 | 0.1 | 1.000 | DE | Gen pop, $14-80 \mathrm{yrs}$ | 35 | 460.00 | 2.4 | NR | 1 | 0.02-0.6 | 0-2\% | 0-2\% | 0-2\% |
| $\begin{aligned} & \text { FB } \\ & 0266 \end{aligned}$ | Dewberries, incl boysen- \& loganberry | Total |  | 0.1 | 1.000 | AU | Child, 2-6 yrs | 328 | 3.23 | $<25$ | NR | 1 | 0.02 | 0\% | - | 0\% |
| $\begin{aligned} & \text { FB } \\ & 0272 \end{aligned}$ | Raspberries, red, black (all commodities) | highest utilization: Total | 0 | 0.2 | 1.000 | FR | Child, 3-6 yrs | 0 | 157.50 | 4.3 | NR | 1 | $\begin{array}{r} 0.07- \\ 1.67 \end{array}$ | 0-6\% | 1-3\% | 0-6\% |
| $\begin{aligned} & \text { FB } \\ & 0021 \end{aligned}$ | Currants, red, black, white (all commodities) | highest utilization: Total | 0 | 0.21 | 1.000 | AU | $\begin{aligned} & \text { Gen pop, } \\ & >2 \mathrm{yrs} \end{aligned}$ | 322 | 797.60 | 14.9 | NR | 1 | 0.14-2.5 | 0-8\% | 0-8\% | 0-7\% |
| $\begin{aligned} & \text { FB } \\ & 0265 \end{aligned}$ | Cranberry <br> (all commodities) | highest utilization: Total | 0 | 0.13 | 1.000 | AU | $\begin{aligned} & \text { Child, 2- } \\ & 16 \text { yrs } \end{aligned}$ | 103 | 279.66 | 1.8 | NR | 1 | $\begin{array}{r} 0.08- \\ 0.96 \end{array}$ | 0-3\% | 0-2\% | 3-3\% |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry (all commodities) | highest utilization: Total | 0 | 0.12 | 1.000 | FR | Child, 3-6 yrs | 0 | 339.40 | 13.4 | NR | 1 | $\begin{array}{r} 0.14- \\ 2.15 \end{array}$ | 0-7\% | $0-4 \%$ | 0-7\% |
| FI 0353 | Pineapple <br> (all commodities) | highest utilization: raw without peel | 0 | $\begin{array}{r} 0.07- \\ 0.2 \end{array}$ | 1.000 | JP | Child, 1-6 yrs | 67 | 499.80 | 1116.0 | 3 | 2b | $\begin{array}{r} 2.33- \\ 6.17 \end{array}$ | 8-20\% | 4-10\% | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ |
| $\begin{aligned} & \text { VA } \\ & 0385 \end{aligned}$ | Onion, bulb <br> (all commodities) | highest utilization: raw without skin | 0 | 0.05 | 1.000 | JP | Child, 1-6 yrs | 748 | 102.00 | 244.4 | 3 | 2b | $\begin{array}{r} 0.08- \\ 0.93 \end{array}$ | 0-3\% | 0-1\% | 0-3\% |
| $\begin{aligned} & \text { VA } \\ & 0389 \end{aligned}$ | Spring onion <br> (all commodities) | highest utilization: cooked/boiled | 0 | 0.65 | 1.000 | NL | Child, 2-6 yrs | E | 20.30 | 30.0 | 3 | 2b | $\begin{array}{r} 1.66- \\ 2.15 \end{array}$ | 6-7\% | 3-3\% | 6-7\% |
| $\begin{aligned} & \text { VB } \\ & 0041 \end{aligned}$ | Cabbage, head (all commodities) | highest utilization: raw | 0 | 0.35 | 1.000 | CN | Child, 1-6 yrs | 287 | 255.54 | 1402.5 | 3 | 2b | $\begin{array}{r} 13.35- \\ 16.63 \end{array}$ | $\begin{gathered} 40- \\ 60 \% \end{gathered}$ | $\begin{gathered} 20- \\ 30 \% \end{gathered}$ | $\begin{gathered} 40- \\ 60 \% \end{gathered}$ |

DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$


|  |  |  | DIAZINON (22) |  |  |  |  |  |  | IESTI |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw ( $30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw}$ ) |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $100 \%$ <br> all | $100 \%$ <br> gen pop | $100 \%$ <br> child |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR- } \\ P \\ \mathrm{mg} / \mathrm{kg} \end{array}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n |  | Unit weight, edible portion, g | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | ARfD rounded | ARfD rounded | ARfD rounded |
| $\begin{aligned} & \text { VL } \\ & 0467 \end{aligned}$ | Chinese cabbage, type pe-tsai <br> (all commodities) | highest utilization: Total | 0 | 0.02 | 1.000 | CN | Child, 1-6 <br> yrs | 2788 | 336.16 | 1500.0 | 3 | 2b | $\begin{array}{r} 0.25- \\ 1.25 \end{array}$ | 1-4\% | 1-3\% | 1-4\% |
| $\begin{aligned} & \text { VL } \\ & 0480 \end{aligned}$ | Kale (borecole, collards) <br> (all commodities) | highest utilization: Total | 0 | 0.02 | 1.000 | DE | Gen pop, 14-80 yrs | 123 | 669.80 | 672.0 | 3 | 2b | $\begin{array}{r} 0.33- \\ 0.53 \end{array}$ | 1-2\% | 1-2\% | 1-2\% |
| $\begin{aligned} & \text { VL } \\ & 0482 \end{aligned}$ | Lettuce, head (all commodities) | highest utilization: raw | 0 | 0.5 | 1.000 | NL | Child, 2-6 <br> yrs | 91 | 140.10 | 338.9 | 3 | 2b | $\begin{array}{r} 4.7- \\ 11.42 \end{array}$ | $\begin{gathered} 20- \\ 40 \% \end{gathered}$ | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ | $\begin{gathered} 20- \\ 40 \% \end{gathered}$ |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | Total |  | 0.5 | 1.000 | CN | Child, 1-6 <br> yrs | 243 | 387.25 | 305.4 | 3 | 2a | 30.92 | 100\% | 30\% | 100\% |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | raw |  | 0.5 | 1.000 | NL | Child, 2-6 <br> yrs | 91 | 140.10 | 117.8 | 3 | 2a | 10.21 | 30\% | 10\% | 30\% |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | cooked/boiled |  | 0.5 | 1.000 | NL | Gen pop, $>1$ yrs | 2 | 220.89 | 79.0 | 3 | 2a | 2.88 | 10\% | 10\% | NC |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach <br> (all commodities) | highest utilization: Total | 0 | 0.5 | 1.000 | ZA | Child, 1-5 <br> yrs | - | 237.48 | 197.8 | 3 | 2a | $\begin{aligned} & 2.22- \\ & 22.29 \end{aligned}$ | 7-70\% | 7-20\% | 7-70\% |
| $\begin{aligned} & \text { VP } \\ & 0061 \end{aligned}$ | Beans, green, with pods, raw: beans except broad bean \& soya bean (i.e. immature seeds + pods) (Phaseolus spp) (all commodities) | highest utilization: canned/preserved | 0 | 0.2 | 1.000 | NL | Toddler, 8-20 m | E | 127.90 | 2.3 | NR | 1 | $\begin{array}{r} 0.76- \\ 2.51 \end{array}$ | 3-8\% | 3-5\% | 8-8\% |
| $\begin{aligned} & \text { VP } \\ & 0064 \end{aligned}$ | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) (all commodities) | highest utilization: Total | 0 | 0.2 | 1.000 | UK | Child, $1.5-4.5$ <br> yrs | 57 | 174.00 | $<25$ | NR | 1 | 0.76-2.4 | 3-8\% | 2-6\% | 4-8\% |

DIAZINON (22)
$\mathrm{ARfD}=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$


|  |  |  | DIAZINON (22) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | ARfD $=0.03 \mathrm{mg} / \mathrm{kg} \mathrm{bw}(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$ |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $\begin{gathered} 100 \% \\ \text { all } \end{gathered}$ | $\begin{gathered} 100 \% \\ \text { gen pop } \end{gathered}$ | $\begin{aligned} & 100 \% \\ & \text { child } \end{aligned}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR- } \\ \mathrm{P} \\ \mathrm{mg} / \mathrm{kg} \\ \hline \end{array}$ | $\begin{gathered} \text { HR or } \\ \text { HR-P } \\ \mathrm{mg} / \mathrm{kg} \end{gathered}$ | DCF | Coun- <br> try | Population group | n | Large portion, g/person | Unit weight, edible portion, | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \text { day } \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ | $\begin{array}{r} \% \\ \text { ARfD } \\ \text { rounded } \end{array}$ |
| 0105 | (mammalian) |  |  |  |  |  | yrs |  |  |  |  |  |  |  |  |  |
| ML 0106 | Milks | Total |  | 0.02 | 1.000 | NL | Toddler, 8-20 m | 1882 | 1060.67 | NR | NR | 3 | ND | - | - | - |
| $\begin{aligned} & \text { PM } \\ & 0110 \end{aligned}$ | Poultry meat | Total | NA | NA | 1.000 | CN | $\begin{aligned} & \text { Child, 1-6 } \\ & \text { yrs } \\ & \hline \end{aligned}$ | 175 | 347.00 | NR | NR | 1 | NA | 1\% | 1\% | 1\% |
| $\begin{aligned} & \text { PM } \\ & 0110 \end{aligned}$ | Poultry meat: $10 \%$ as fat | Total |  | 0.02 | 1.000 | CN | Child, 1-6 <br> yrs | 175 | 34.70 | NR | NR | 1 | 0.04 | 0\% | 0\% | 0\% |
| $\begin{aligned} & \text { PM } \\ & 0110 \end{aligned}$ | Poultry meat: $90 \%$ as muscle | Total |  | 0.02 | 1.000 | CN | Child, 1-6 <br> yrs | 175 | 312.30 | NR | NR | 1 | 0.39 | 1\% | 1\% | 1\% |
| $\begin{aligned} & \text { PO } \\ & 0111 \end{aligned}$ | Poultry, edible offal (includes kidney, liver and skin) | Total |  | 0.02 | 1.000 | CN | $\begin{aligned} & \text { Gen pop, } \\ & >1 \text { yrs } \end{aligned}$ | 421 | 345.63 | NR | NR | 1 | 0.13 | 0\% | 0\% | 0\% |
| PE 0112 | Eggs | Total |  | 0.02 | 1.000 | CN | Child, 1-6 yrs | 136 | 195.82 | NR | NR | 1 | 0.24 | 1\% | 0\% | 1\% |

Annex 4
IESTI

|  |  |  | ARfD $=2 \mathrm{mg} / \mathrm{kg} \mathrm{bw}(2000 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$ |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $9 \%$ <br> all | $5 \%$ <br> gen pop | $\begin{gathered} 9 \% \\ \text { child } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR-P } \\ \mathrm{mg} / \mathrm{kg} \end{array}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, g/person | Unit weight, edible portion, g | Variability factor | Case | IESTI <br> $\mu \mathrm{g} / \mathrm{kg}$ <br> bw/day | ARfD rounded | ARfD rounded | ARfD rounded |
| 001 | CITRUS FRUITS | - |  | 0.22 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { FC } \\ & 0303 \end{aligned}$ | Kumquats <br> (all commodities) | highest utilization: Total | 0 | 0.22 | 1.000 | JP | Gen pop, $>1$ yrs | 135 | 120.00 | $<25$ | NR | 1 | $\begin{array}{r} 0.04- \\ 0.53 \end{array}$ | 0-0\% | $0-0 \%$ | 0-0\% |
| FC $0204$ | Lemon <br> (all commodities) | highest utilization: <br> Total | 0.02 | 0.22 | 1.000 | FR | Child, 3-6 yrs | 0 | 58.15 | 64.0 | 3 | 2b | $\begin{array}{r} 0.01- \\ 2.03 \end{array}$ | 0-0\% | $0-0 \%$ | 0-0\% |
| FC $0205$ | Lime <br> (all commodities) | highest utilization: Total | 0.02 | 0.22 | 1.000 | AU | Gen pop, $>2 \mathrm{yrs}$ | 579 | 259.21 | 49.0 | 3 | 2a | 0-1.17 | 0-0\% | $0-0 \%$ | 0-0\% |
| 001B | Mandarins | - |  | 0.22 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \mathrm{FC} \\ & 0003 \end{aligned}$ | Mandarins (incl mandarin-like hybrids) (all commodities) | highest utilization: raw, without peel | 0.02 | 0.22 | 1.000 | CN | Child, 1-6 yrs | 151 | 586.75 | 124.3 | 3 | 2a | 0-11.39 | 0-1\% | $0-0 \%$ | 0-1\% |
| $\begin{aligned} & \text { FC } \\ & 0004 \end{aligned}$ | Oranges, sweet, sour (incl orange-like hybrids) (all commodities) | highest utilization: Total | 0.02 | 0.22 | 1.000 | AU | Child, 2-6 <br> yrs | 1735 | 800.83 | 155.8 | 3 | 2a | $\begin{aligned} & 0.01- \\ & 12.88 \end{aligned}$ | 0-1\% | $0-0 \%$ | 0-1\% |
| $\begin{aligned} & \text { FC } \\ & 0005 \end{aligned}$ | Pummelo and Grapefruits (incl Shaddock-like hybrids, among others Grapefruit) (all commodities) | highest utilization: raw, without peel | 0.02 | 0.22 | 1.000 | DE | Child, 2-4 yrs | 12 | 358.60 | 178.5 | 3 | 2a | 0-9.75 | 0-0\% | $0-0 \%$ | 0-0\% |
| FP 0226 | Apple <br> (all commodities) | highest utilization: Total | 0.11 | 0.37 | 1.000 | US | Child, 1-6 yrs | - | 624.45 | 127.0 | 3 | 2a | $\begin{aligned} & 0.33- \\ & 21.67 \end{aligned}$ | 0-1\% | $0-0 \%$ | 0-1\% |
| FB | Blueberries | highest utilization: | 0 | 7.5 | 1.000 | DE | Gen pop, | 70 | 388.00 | 1.8 | NR | 1 | 17.6- | 1-2\% | 0-2\% | 1-2\% |


|  |  |  | MALATHION (49) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | ARfD $=2 \mathrm{mg} / \mathrm{kg}$ bw $(2000 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$ |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $\begin{gathered} 9 \% \\ \text { all } \end{gathered}$ | $5 \%$ <br> gen pop | $9 \%$ <br> child |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | STMR or STMR-P $\mathrm{mg} / \mathrm{kg}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n |  | Unit weight, edible portion, g | Variability factor | Case | IESTI <br> $\mu \mathrm{g} / \mathrm{kg}$ <br> bw/day | \% <br> ARfD <br> rounded | ARfD rounded | ARfD rounded |
| 0020 | (all commodities) | raw with skin |  |  |  |  | 14-80 yrs |  |  |  |  |  | 38.1 |  |  |  |
| $\begin{aligned} & \text { FB } \\ & 0269 \end{aligned}$ | Grape <br> (all commodities) | highest utilization: raw with skin | 0 | 2.6 | 1.000 | CN | Child, 1-6 yrs | 232 | 366.72 | 636.6 | 3 | 2b | $\begin{array}{r} 8.24 \\ 177.27 \end{array}$ | 0-9\% | 0-4\% | 0-9\% |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry <br> (all commodities) | highest utilization: Total | 0 | 0.59 | 1.000 | FR | Child, 3-6 yrs | 0 | 339.40 | 13.4 | NR | 1 | $\begin{array}{r} 0.67- \\ 10.6 \end{array}$ | 0-1\% | 0-0\% | 0-1\% |
| $\begin{aligned} & \text { VA } \\ & 0385 \end{aligned}$ | Onion, bulb (all commodities) | highest utilization: raw without skin | 0 | 0.59 | 1.000 | JP | Child, 1-6 yrs | 748 | 102.00 | 244.4 | 3 | 2b | $\begin{aligned} & 0.94- \\ & 11.01 \end{aligned}$ | 0-1\% | 0-0\% | 0-1\% |
| $\begin{aligned} & \text { VA } \\ & 0389 \end{aligned}$ | Spring onion <br> (all commodities) | highest utilization: cooked/boiled | 0 | 5 | 1.000 | NL | Child, 2-6 yrs | E | 20.30 | 30.0 | 3 | 2b | $\begin{array}{r} 12.74- \\ 16.55 \end{array}$ | 1-1\% | 0-0\% | 1-1\% |
| $\begin{aligned} & \text { VC } \\ & 0424 \end{aligned}$ | Cucumber <br> (all commodities) | highest utilization: raw with skin | 0 | 0.1 | 1.000 | CN | Child, 1-6 yrs | 340 | 212.11 | 458.1 | 3 | 2b | $\begin{array}{r} 0.91- \\ 3.94 \end{array}$ | 0-0\% | 0-0\% | $0-0 \%$ |
| VO <br> 0444 | Peppers, chili (all commodities) | highest utilization: raw with skin | 0 | 0.08 | 1.000 | CN | Gen pop, $>1 \mathrm{yrs}$ | 1743 | 295.71 | 43.2 | 3 | 2a | $\begin{array}{r} 0.06- \\ 0.57 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { VO } \\ & 0445 \end{aligned}$ | Peppers, sweet (incl. pim(i)ento) (bell pepper, paprika) (all commodities) | highest utilization: raw with skin | 0 | 0.08 | 1.000 | CN | Child, 1-6 yrs | 1002 | 169.85 | 170.0 | 3 | 2b | $\begin{array}{r} 0.48- \\ 2.53 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn (corn-on-the-cob) (all commodities) | highest utilization: cooked/boiled | 0 | 0.02 | 1.000 | TH | Child, 3-6 yrs | 1383 | 196.99 | 191.1 | 3 | 2a | $\begin{array}{r} 0.08- \\ 0.68 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { VO } \\ & 0448 \end{aligned}$ | Tomato <br> (all commodities) | highest utilization: dried | 0 | $\begin{array}{r} 0.0123- \\ 0.41 \end{array}$ | 5.000 | AU | Gen pop, $>2$ yrs | 61 | 861.10 | 8.0 | NR | 1 | $\begin{aligned} & 8.84- \\ & 26.35 \end{aligned}$ | 0-1\% | 0-1\% | $0-1 \%$ |
| $\begin{aligned} & \text { VL } \\ & 0485 \end{aligned}$ | Mustard greens <br> (all commodities) | highest utilization: raw | 0 | 1.1 | 1.000 | CN | Child, 1-6 <br> yrs | 635 | 299.31 | 244.8 | 3 | 2a | $\begin{aligned} & 8.12- \\ & 53.78 \end{aligned}$ | 0-3\% | 0-1\% | 0-3\% |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach <br> (all commodities) | highest utilization: Total | 0 | 2.2 | 1.000 | ZA | Child, 1-5 yrs | - | 237.48 | 197.8 | 3 | 2a | $\begin{aligned} & 9.79- \\ & 98.07 \end{aligned}$ | 0-5\% | 0-2\% | 0-5\% |

MALATHION (49)
ARfD $=2 \mathrm{mg} / \mathrm{kg}$ bw $(2000 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$


## ANNEX 5: REPORTS AND OTHER DOCUMENTS RESULTING FROM PREVIOUS JOINT MEETINGS OF THE FAO PANEL OF EXPERTS ON PESTICIDE RESIDUES IN FOOD AND THE ENVIRONMENT AND THE WHO CORE ASSESSMENT GROUP ON PESTICIDE RESIDUES

1. Principles governing consumer safety in relation to pesticide residues. Report of a meeting of a WHO Expert Committee on Pesticide Residues held jointly with the FAO Panel of Experts on the Use of Pesticides in Agriculture. FAO Plant Production and Protection Division Report, No. PL/1961/11; WHO Technical Report Series, No. 240, 1962.
2. Evaluation of the toxicity of pesticide residues in food. Report of a Joint Meeting of the FAO Committee on Pesticides in Agriculture and the WHO Expert Committee on Pesticide Residues. FAO Meeting Report, No. PL/1963/13; WHO/Food Add./23, 1964.
3. Evaluation of the toxicity of pesticide residues in food. Report of the Second Joint Meeting of the FAO Committee on Pesticides in Agriculture and the WHO Expert Committee on Pesticide Residues. FAO Meeting Report, No. PL/1965/10; WHO/Food Add./26.65, 1965.
4. Evaluation of the toxicity of pesticide residues in food. FAO Meeting Report, No. PL/1965/10/1; WHO/Food Add./27.65, 1965.
5. Evaluation of the hazards to consumers resulting from the use of fumigants in the protection of food. FAO Meeting Report, No. PL/1965/10/2; WHO/Food Add./28.65, 1965.
6. Pesticide residues in food. Joint report of the FAO Working Party on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 73; WHO Technical Report Series, No. 370, 1967.
7. Evaluation of some pesticide residues in food. FAO/PL:CP/15; WHO/Food Add./67.32, 1967.
8. Pesticide residues. Report of the 1967 Joint Meeting of the FAO Working Party and the WHO Expert Committee. FAO Meeting Report, No. PL:1967/M/11; WHO Technical Report Series, No. 391, 1968.
9. 1967 Evaluations of some pesticide residues in food. FAO/PL:1967/M/11/1; WHO/Food Add./68.30, 1968.
10. Pesticide residues in food. Report of the 1968 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 78; WHO Technical Report Series, No. 417, 1968.
11. 1968 Evaluations of some pesticide residues in food. FAO/PL:1968/M/9/1; WHO/Food Add./69.35, 1969.
12. Pesticide residues in food. Report of the 1969 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Group on Pesticide Residues. FAO Agricultural Studies, No. 84; WHO Technical Report Series, No. 458, 1970.
13. 1969 Evaluations of some pesticide residues in food. FAO/PL:1969/M/17/1; WHO/Food Add./70.38, 1970.
14. Pesticide residues in food. Report of the 1970 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 87; WHO Technical Report Series, No. 4574, 1971.
15. 1970 Evaluations of some pesticide residues in food. AGP:1970/M/12/1; WHO/Food Add./71.42, 1971.
16. Pesticide residues in food. Report of the 1971 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 88; WHO Technical Report Series, No. 502, 1972.
17. 1971 Evaluations of some pesticide residues in food. AGP:1971/M/9/1; WHO Pesticide Residue Series, No. 1, 1972.
18. Pesticide residues in food. Report of the 1972 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 90; WHO Technical Report Series, No. 525, 1973.
19. 1972 Evaluations of some pesticide residues in food. AGP:1972/M/9/1; WHO Pesticide Residue Series, No. 2, 1973.
20. Pesticide residues in food. Report of the 1973 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 92; WHO Technical Report Series, No. 545, 1974.
21. 1973 Evaluations of some pesticide residues in food. FAO/AGP/1973/M/9/1; WHO Pesticide Residue Series, No. 3, 1974.
22. Pesticide residues in food. Report of the 1974 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Agricultural Studies, No. 97; WHO Technical Report Series, No. 574, 1975.
23. 1974 Evaluations of some pesticide residues in food. FAO/AGP/1974/M/11; WHO Pesticide Residue Series, No. 4, 1975.
24. Pesticide residues in food. Report of the 1975 Joint Meeting of the FAO Working Party of Experts on Pesticide Residues and the WHO Expert Committee on Pesticide Residues. FAO Plant Production and Protection Series, No. 1; WHO Technical Report Series, No. 592, 1976.
25. 1975 Evaluations of some pesticide residues in food. AGP:1975/M/13; WHO Pesticide Residue Series, No. 5, 1976.
26. Pesticide residues in food. Report of the 1976 Joint Meeting of the FAO Panel of Experts on Pesticide Residues and the Environment and the WHO Expert Group on Pesticide Residues. FAO Food and Nutrition Series, No. 9; FAO Plant Production and Protection Series, No. 8; WHO Technical Report Series, No. 612, 1977.
27. 1976 Evaluations of some pesticide residues in food. AGP:1976/M/14, 1977.
28. Pesticide residues in food - 1977. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues and Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 10 Rev, 1978.
29. Pesticide residues in food: 1977 evaluations. FAO Plant Production and Protection Paper 10 Suppl., 1978.
30. Pesticide residues in food - 1978. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues and Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 15, 1979.
31. Pesticide residues in food: 1978 evaluations. FAO Plant Production and Protection Paper 15 Suppl., 1979.
32. Pesticide residues in food - 1979. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 20, 1980.
33. Pesticide residues in food: 1979 evaluations. FAO Plant Production and Protection Paper 20 Suppl., 1980
34. Pesticide residues in food - 1980. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 26, 1981.
35. Pesticide residues in food: 1980 evaluations. FAO Plant Production and Protection Paper 26 Suppl., 1981.
36. Pesticide residues in food - 1981. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 37, 1982.
37. Pesticide residues in food: 1981 evaluations. FAO Plant Production and Protection Paper 42, 1982.
38. Pesticide residues in food - 1982. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 46, 1982.
39. Pesticide residues in food: 1982 evaluations. FAO Plant Production and Protection Paper 49, 1983.
40. Pesticide residues in food - 1983. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 56, 1985.
41. Pesticide residues in food: 1983 evaluations. FAO Plant Production and Protection Paper 61, 1985.
42. Pesticide residues in food - 1984. Report of the Joint Meeting on Pesticide Residues. FAO Plant Production and Protection Paper 62, 1985.
43. Pesticide residues in food - 1984 evaluations. FAO Plant Production and Protection Paper 67, 1985.
44. Pesticide residues in food - 1985. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and a WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 68, 1986.
45. Pesticide residues in food - 1985 evaluations. Part I. Residues. FAO Plant Production and Protection Paper 72/1, 1986.
46. Pesticide residues in food - 1985 evaluations. Part II. Toxicology. FAO Plant Production and Protection Paper 72/2, 1986.
47. Pesticide residues in food - 1986. Report of the Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and a WHO Expert Group on Pesticide Residues. FAO Plant Production and Protection Paper 77, 1986.
48. Pesticide residues in food - 1986 evaluations. Part I. Residues. FAO Plant Production and Protection Paper 78, 1986.
49. Pesticide residues in food - 1986 evaluations. Part II. Toxicology. FAO Plant Production and Protection Paper 78/2, 1987.
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#### Abstract

A Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Core Assessment Group on Pesticide Residues (JMPR) was held in Geneva, Switzerland, from 9 to 13 May 2016. The three pesticides evaluated at the meeting were placed on the agenda by the JMPR Secretariat following the recommendation of an electronic task force of the WHO Core Assessment Group that they be re-evaluated due to public health concerns identified by the International Agency for Research on Cancer (IARC) and the availability of a significant number of new studies. During the meeting, the WHO Core Assessment Group was responsible for reviewing epidemiological, toxicological and related data in order to establish acceptable daily intakes (ADIs) and acute reference doses (ARfDs) of the pesticides for humans, where necessary. As no residue data were requested, the FAO Expert was responsible for estimating the dietary exposures (both short-term and long-term) to the pesticides reveiewed and, on this basis, performed dietary risk assessments in relation to their ADIs or ARfDs. This report contains information on ADIs, ARfDs and general principles for the evaluation of pesticides. The recommendations of the Joint Meeting, including further research and information, are proposed for use by Member governments of the respective agencies and other interested parties.


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## T, toxicological evaluation

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Geneva, 9-13 May 2016

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[^12]|  | ABBREVIATIONS |
| :---: | :---: |
| ADI | acceptable daily intake |
| AHS | Agricultural Health Study |
| AMPA | aminomethylphosphonic acid |
| ARfD | acute reference dose |
| BMD | benchmark dose |
| bw | body weight |
| CAS | Chemical Abstracts Service |
| CCPR | Codex Committee on Pesticide Residues |
| CI | confidence interval |
| CYP | cytochrome P450 |
| DCF | diet correction factor |
| DNA | deoxyribonucleic acid |
| $\mathrm{F}_{0}$ | parental generation |
| $\mathrm{F}_{1}$ | first filial generation |
| FAO | Food and Agriculture Organization of the United Nations |
| GEMS/Food | Global Environment Monitoring System - Food Contamination Monitoring and Assessment Programme |
| GLP | good laboratory practice |
| HR | highest residue in the edible portion of a commodity found in trials used to estimate a maximum residue level in the commodity |
| HR-P | highest residue in a processed commodity calculated by multiplying the HR of the raw commodity by the corresponding processing factor |
| IARC | International Agency for Research on Cancer |
| IEDI | international estimated daily intake |
| IESTI | international estimate of short-term dietary intake |
| ISO | International Organization for Standardization |
| IUPAC | International Union of Pure and Applied Chemistry |
| JMPR | Joint FAO/WHO Meeting on Pesticide Residues |
| $\mathrm{LC}_{50}$ | median lethal concentration |
| $\mathrm{LD}_{50}$ | median lethal dose |
| LOAEL | lowest-observed-adverse-effect level |
| MDCA | malathion dicarboxylic acid |
| MIC | minimum inhibitory concentration |
| MMCA | malathion monocarboxylic acid |
| MRL | maximum residue limit |
| NC | no national consumption data available |


| NES | not elsewhere specified |
| :--- | :--- |
| NHL | non-Hodgkin lymphoma |
| NOAEC | no-observed-adverse-effect concentration |
| NOAEL | no-observed-adverse-effect level |
| OR | odds ratio |
| PP | processed product |
| ppm | parts per million |
| RAC | relative risk agricultural commodity |
| RR | supervised trials median residue <br> STMR |
| STMR-P | multiplying the STMR of the raw commodity by the corresponding processing <br> factor |
| TAF | toxicity adjustment factor |
| WHO | World Health Organization |

## USE OF JMPR REPORTS AND EVALUATIONS BY REGISTRATION AUTHORITIES

Most of the summaries and evaluations contained in this report are based on unpublished proprietary data submitted for use by JMPR in making its assessments. A registration authority should not grant a registration on the basis of an evaluation unless it has first received authorization for such use from the owner of the data submitted for the JMPR review or has received the data on which the summaries are based, either from the owner of the data or from a second party that has obtained permission from the owner of the data for this purpose.

## PESTICIDE RESIDUES IN FOOD

## REPORT OF THE MAY 2016 JOINT FAO/WHO MEETING OF EXPERTS

## 1. INTRODUCTION

A Joint Meeting of the Food and Agriculture Organization of the United Nations (FAO) Panel of Experts on Pesticide Residues in Food and the Environment and the World Health Organization (WHO) Core Assessment Group on Pesticide Residues (JMPR) was held at WHO Headquarters, Geneva (Switzerland), from 9 to 13 May 2016.

The meeting was opened by Dr Kazuaki Miyagishima, Director of the Department of Food Safety and Zoonoses, WHO, who welcomed participants on behalf of the Directors General of WHO and FAO. Dr Miyagishima stated that the meeting was convened to re-evaluate three compounds for which new studies had become available since their last full assessments. He reminded the participants of the importance of the functional separation between risk assessment and risk management and of the role that JMPR plays as the expert risk assessment body providing scientific advice to Codex and to Member States. He urged the participants to be guided by JMPR's standing rules and procedures based on the weight of evidence approach. Dr Miyagishima thanked the participants for devoting significant time and effort to the work of JMPR, including the preparatory work of paramount importance that had taken place in the past months. He reminded the experts that they were invited as independent experts acting in their own individual capacities and not as representatives of their countries or organizations. He also reminded the participants of the confidential nature of the meeting, in order to allow experts to freely express their opinions.

During the meeting, the WHO Core Assessment Group was responsible for reviewing epidemiological, toxicological and related data in order to establish acceptable daily intakes (ADIs) and acute reference doses (ARfDs), where necessary. As no residue data were requested, the FAO Expert was responsible for estimating the dietary exposures (both short-term and long-term) to the pesticides reviewed and, on this basis, performed dietary risk assessments in relation to their ADIs or ARfDs.

The Meeting re-evaluated three pesticides, established ADIs and ARfDs and recommended them for use by the Codex Committee on Pesticide Residues (CCPR). The Meeting also considered issues related to the evaluation of genotoxicity and epidemiological studies in relation to the risk assessment of chemicals.

### 1.1 DECLARATION OF INTERESTS

The Secretariat informed the Meeting that all experts participating in the May 2016 JMPR had completed declaration of interest forms and that no conflicts had been identified.

## 2. GENERAL CONSIDERATIONS

### 2.1 GENERAL CONSIDERATIONS ON THE EVALUATION OF GENOTOXICITY STUDIES

A large number of genotoxicity studies were evaluated during the present meeting. These were identified through direct submission to JMPR, searches of the publicly available literature and requests to the International Agency for Research on Cancer (IARC) Monographs Secretariat and industry groups. The studies evaluated included unpublished (primarily guideline) studies submitted to support pesticide registration as well as peer-reviewed studies published in the scientific literature. The number, quality and relevance of studies differed widely for each chemical and necessitated that a somewhat different approach be used to evaluate each pesticide. As a general strategy, the studies were separated into categories based largely on phylogenetic relevance and significance of the genetic end-point measured. The categories used were human biomonitoring, in vivo mammals, in vitro mammalian cells, in vitro bacteria, phylogenetically distant organisms, metabolites in vivo and metabolites in vitro. The evaluation was conducted for the pesticide active ingredient, its formulation products and prominent metabolites, as data were available. For the three pesticides evaluated, the human biomonitoring studies were most often confounded by exposures to other pesticides or considered to have other limitations. Among the genotoxicity studies, in vivo studies in mammals were given the greatest weight, compared with cell culture studies or investigations in phylogenetically distant organisms. Studies of gene mutations and chromosomal alterations were also given more weight than studies measuring other less serious or transient types of genotoxic damage. With regard to route of exposure, studies in which chemicals were administered by the oral route were considered to be of most relevance for evaluating low-level dietary exposures.

Following an evaluation and weighting of the studies, taking the criteria described above and the quality of the studies into account, an overall weight of evidence approach was used to reach conclusions about the genotoxicity of the individual pesticides. An important aspect of the evaluation was whether the genotoxic effect would be likely to occur in humans exposed to low levels of the pesticide present as residues in food.

The Meeting recommended that a guidance document be developed for the evaluation of genotoxicity studies, taking the experience gained from this meeting into account.

### 2.2 METHODS FOR THE EVALUATION OF EPIDEMIOLOGICAL EVIDENCE FOR RISK ASSESSMENT

## Identification of compound/cancer sites and screening of papers

There is a large body of literature regarding pesticide exposures and non-cancer outcomes (neurodevelopmental, neurodegenerative and reproductive outcomes, among other health outcomes), but the assessment of the epidemiological evidence on diazinon, glyphosate and malathion was restricted to studies of cancer outcomes. This restriction was partly driven by feasibility reasons: a clinically relevant adverse effect size (or an acceptable level of risk) for a non-cancer outcome must be defined, and the methodologies for hazard identification and characterization based on observational epidemiological findings of non-carcinogenic adverse effects are less well established than those for cancer. ${ }^{1}$

[^13]The IARC Monographs on malathion, diazinon and glyphosate referred to a total of 45 epidemiological studies. ${ }^{1}$ Databases were searched for any relevant articles published after the studies cited in these Monographs using the following search terms: [(diazinon OR glyphosate OR malathion) AND cancer] and [(diazinon OR glyphosate OR malathion) AND (NHL OR lymphoma OR leukemia OR "lung cancer" OR "prostate cancer")] in PubMed (limited to Humans; published in the last 5 years) and Scopus (limited to 2014-2016). Two studies published since the publication of the IARC Monographs that evaluated at least one of malathion, diazinon or glyphosate were identified in relation to cancer outcomes. ${ }^{2}$ An additional study on prostate cancer, ${ }^{3}$ which was not included in the IARC Monographs, was also identified.

The pre-agreed evaluation process shown in Fig. 1 was used to (1) select compound/cancer site combinations to include in this evaluation; (2) screen papers for inclusion/exclusion in this evaluation (Tier 1 screening criteria); and (3) evaluate the information available for risk assessment. In this process, it was noted that there were stand-alone analyses for specific subtypes of nonHodgkin lymphoma (NHL). The risk for subtypes of NHL was not evaluated separately, as there was insufficient evidence (too few studies or small numbers of cases); the risk for other haematopoietic and lymphoid tumours was also not evaluated separately, as the positive associations identified by IARC were for total NHL.


[^14]
## Evaluation of evidence for the compound/cancer site associations

Several aspects of each study and of all studies combined were considered in this evaluation, including factors that decrease the level of confidence in the body of evidence, such as risk of bias, unexplained inconsistency and imprecision; and factors that increase the level of confidence, such as large magnitude of effect, dose-response and consistency. ${ }^{1}$ The findings for each study were summarized in tables, and risk estimates for non-quantitative exposure assessment (predominantly ever versus never use) were summarized in forest plots.

## Evaluation of information available for risk assessment/hazard characterization

To evaluate overall evidence for dose-response relationships, risk estimates were plotted against quantitative exposure measures (for studies that had used these). The most commonly used quantitative exposure metric was days of use per year. Where studies had used other quantitative exposure metrics (e.g. lifetime days of exposure), data were requested from the authors on median "days of use per year" for the participants in each of the original exposure categories, although this information was not always forthcoming. These additional data allowed the translation and plotting of risk estimates from different studies on the same exposure scale (days of use per year).

[^15]
## 3. EVALUATION OF DATA FOR ACCEPTABLE DAILY INTAKE AND ACUTE REFERENCE DOSE FOR HUMANS

### 3.1 DIAZINON (22)

## TOXICOLOGY

Diazinon is the common name approved by the International Organization for Standardization (ISO) for $O, O$-diethyl $O$-2-isopropyl-6-methylpyrimidin-4-yl phosphorothioate (International Union of Pure and Applied Chemistry [IUPAC]), with the Chemical Abstracts Service (CAS) number 333-41-5.

Diazinon is a contact organophosphorus insecticide with a wide range of insecticidal activity. It is effective against adult and juvenile forms of flying insects, crawling insects, acarians and spiders. Diazoxon, the biologically active metabolite of diazinon, inhibits the activity of cholinesterases.

Diazinon is used mainly as a pesticide in agriculture and as a drug in veterinary medicine. Thus, the major source of diazinon residues in edible crops is from its use as an agricultural pesticide; residues in meat, offal and other animal products arise from its use as a veterinary drug containing active ingredient.

Diazinon has been evaluated by JMPR on several occasions since the first evaluation in 1963. In the most recent evaluation, in 2006, the Meeting established an ADI of $0-0.005 \mathrm{mg} / \mathrm{kg}$ body weight (bw), based on a no-observed-adverse-effect level (NOAEL) of $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day for inhibition of erythrocyte acetylcholinesterase activity in a 92-day repeated-dose toxicity study in rats. The 2006 Meeting reaffirmed the ARfD of $0.03 \mathrm{mg} / \mathrm{kg}$ bw, established by the 2001 JMPR , based on a NOAEL of $2.5 \mathrm{mg} / \mathrm{kg}$ bw observed in a study of acute neurotoxicity in rats.

Diazinon was scheduled within the periodic review programme of CCPR for 2021. The compound was placed on the agenda by the JMPR Secretariat following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously considered toxicological data in addition to new published or unpublished toxicological studies and published epidemiological studies on cancer outcomes. Several study reports evaluated at previous JMPR meetings were not available to the present Meeting, as they were not submitted in the sponsor's dossier; for these studies, the evaluations in this report were summarized from the 1993 JMPR monograph without further review.

All critical unpublished studies contained statements of compliance with good laboratory practice (GLP), unless otherwise specified. The studies on human volunteers were conducted in accordance with the principles expressed in the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

Following oral administration to rats, diazinon was almost completely absorbed and rapidly eliminated, mainly in the urine. There was no evidence of accumulation.

Diazinon is metabolized by P 450 to diazoxon, the active metabolite. The main degradative pathway includes the oxidase/hydrolase-mediated cleavage of the ester bond, leading to the pyrimidinol derivative 2-isopropyl-6-methyl-4( 1 H$)$-pyrimidinone, which is further oxidized to more polar metabolites.

## Toxicological data

The oral median lethal dose $\left(\mathrm{LD}_{50}\right)$ for diazinon in rats ranged from 300 to greater than $2150 \mathrm{mg} / \mathrm{kg}$ bw, whereas the dermal $\mathrm{LD}_{50}$ was greater than $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The inhalation median lethal concentration ( $\mathrm{LC}_{50}$ ) was $3.1 \mathrm{mg} / \mathrm{L}$ in rats. Diazinon produced mild skin and eye irritation in rabbits. It caused skin sensitization in the guinea-pig Magnusson and Kligman maximization test.

The most sensitive end-point observed in all species given single and repeated doses of diazinon was inhibition of cholinesterase activity. Brain acetylcholinesterase activity was generally decreased at doses higher than those that inhibited erythrocyte acetylcholinesterase activity. Clinical signs of cholinergic toxicity occurred at doses causing more than $50 \%$ inhibition of brain acetylcholinesterase activity. Female rats were more sensitive than male rats.

Many repeated-dose toxicity studies are available. In both rats and dogs, no effects other than those related to cholinesterase inhibition have been observed at the lowest-observed-adverse-effect level (LOAEL); in general, effects observed at the highest doses can be considered secondary to the cholinergic toxicity. In these studies, NOAELs ranged from 0.02 to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day, and LOAELs ranged from 1 to $15 \mathrm{mg} / \mathrm{kg}$ bw per day, based on erythrocyte acetylcholinesterase inhibition (i.e. $>20 \%$ ), with brain acetylcholinesterase inhibition (i.e. $>10 \%$ ) generally appearing at the next higher dose and clinical cholinergic signs appearing at doses above $23 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 28-day acetylcholinesterase inhibition study, rats received diazinon by dietary administration at a concentration of $0,0.3,30,300$ or 3000 parts per million ( ppm ) (equal to $0,0.02$, $2.3,23$ and $213 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.02,2.4,23$ and $210 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.3 ppm (equal to $0.02 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 30 ppm (equal to $2.3 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a short-term toxicity study, rats were fed diazinon at a concentration of 0 or 2 ppm (equivalent to 0 and $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 7 days or at a concentration of 0 or 25 ppm (equivalent to 0 and $2.5 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 30 days. The NOAEL was 2 ppm (equivalent to $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day), based on inhibition of erythrocyte acetylcholinesterase activity at 25 ppm (equivalent to $2.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 3-month toxicity study, rats were given diets containing diazinon at a concentration of 0 , $0.5,5,250$ or 2500 ppm (equal to $0,0.03,0.3,15$ and $168 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.04$, $0.4,19$ and $212 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 5 ppm (equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at 250 ppm (equal to $15 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 3-month toxicity study, rats were fed diets containing diazinon at a concentration of $0,0.3,30,300$ or 3000 ppm (equal to $0,0.017,1.7,17$ and $177 \mathrm{mg} / \mathrm{kg}$ bw per day for males and 0 , $0.019,1.9,19$ and $196 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.3 ppm (equal to $0.017 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 30 ppm (equal to $1.7 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a third 3-month toxicity study, female rats were fed diets containing diazinon at a concentration of $0,5,10$ or 15 ppm (equivalent to $0,0.5,1$ and $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 92 days. In the second phase, female rats were fed diets containing diazinon at a concentration of $0,1,2,3$ or 4 ppm (equivalent to $0,0.1,0.2,0.3$ and $0.4 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 42 days. In the third phase, female rats were fed diets containing diazinon at a concentration of $0,0.1$, $0.5,1$ or 2 ppm (equivalent to $0,0.01,0.05,0.1$ and $0.2 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) for 35 days. The NOAEL in the first phase was 5 ppm (equivalent to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day), based on inhibition of erythrocyte acetylcholinesterase activity at 10 ppm (equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day) after dosing for 92 days. The NOAEL for females in the second and third phases were the highest tested doses of 4 ppm (equivalent to $0.4 \mathrm{mg} / \mathrm{kg}$ bw per day) and 2 ppm (equivalent to $0.2 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per day) after dosing for 42 and 35 days, respectively.

In a fourth 3-month toxicity study, rats were fed diets containing diazinon at a concentration of $0,5,125$ or 2000 ppm (equal to $0,0.3,7.8$ and $198 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.3,8.9$ and $247 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 5 ppm (equal to $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte acetylcholinesterase activity at 125 ppm (equal to $7.8 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 90-day repeated-dose neurotoxicity study, rats were dosed in the diet at $0,25,125$ or 1000 ppm (equal to $0,1.7,8.4$ and $69.1 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1.8,9.3$ and $82.4 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). A NOAEL could not be identified, as erythrocyte acetylcholinesterase activity was inhibited at $1.7 \mathrm{mg} / \mathrm{kg}$ bw per day, the lowest dose tested.

In considering the NOAELs and LOAELs identified in the 28-day and 3-month (neuro)toxicity studies in rats measuring the inhibition of acetylcholinesterase activity, the Meeting concluded that the extent of acetylcholinesterase inhibition was not dependent on duration of dosing once steady state had been achieved (within 4 weeks). The overall NOAEL for the 28-day and 3month (neuro)toxicity studies in rats was 5 ppm , based on inhibition of erythrocyte acetylcholinesterase activity at the overall LOAEL of 10 ppm . In studies where feed consumption data were used to calculate test substance intake, 5 ppm was equal to $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day. These substance intake data are considered to be more accurate than those calculated using a default conversion factor, in which the NOAEL of 5 ppm is equivalent to $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 90-day toxicity study, dogs were given diets containing diazinon at a concentration of 0 , $0.1,0.5,150$ or 300 ppm (equal to $0,0.0034,0.020,5.9$ and $10.9 \mathrm{mg} / \mathrm{kg}$ bw per day for males and 0 , $0.0037,0.021,5.6$ and $11.6 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively). The NOAEL was 0.5 ppm (equal to $0.020 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at a dietary concentration of 150 ppm (equal to $5.6 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 90 -day toxicity study, dogs were given diazinon at $0,0.3,3$ or $10 \mathrm{mg} / \mathrm{kg}$ bw per day by gelatine capsule. The NOAEL was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at $3 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 1-year toxicity study in dogs given diazinon in the diet at a concentration of $0,0.1,0.5$, 150 or 300 ppm (equal to $0,0.0032,0.015,4.7$ and $7.7 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.0037$, $0.020,4.5$ and $9.1 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 0.5 ppm (equal to $0.015 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte (males and females) and brain (females only) acetylcholinesterase activities at 150 ppm (equal to $4.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL for the 90-day and 1-year toxicity studies in dogs was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte and brain acetylcholinesterase activities at $3 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a pre-GLP carcinogenicity study in mice that was considered adequate to evaluate carcinogenicity but not chronic toxicity, diazinon was administered at a dietary concentration of 0 , 100 or 200 ppm (equivalent to 0,15 and $30 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) over 103 weeks. No treatment-related tumours were observed.

In another pre-GLP carcinogenicity study in mice, diazinon was administered at a dietary concentration of $0,100,200,300$ (males) or 400 (females) ppm (equal to $0,16,31 \mathrm{and} 46 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per day for males and $0,22,43$ and $86 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively) for 104 weeks. Cholinesterase activity was not measured in this study. The NOAEL for chronic toxicity was 200 ppm (equal to $31 \mathrm{mg} / \mathrm{kg}$ bw per day), based on depression of body weight and lower feed consumption at 300 ppm (equal to $46 \mathrm{mg} / \mathrm{kg}$ bw per day). No treatment-related tumours were observed.

In a pre-GLP carcinogenicity study in rats that was considered adequate to evaluate carcinogenicity but not chronic toxicity, diazinon was administered at a dietary concentration of 0 , 400 or 800 ppm (equivalent to 0,20 and $40 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively) over 103 weeks. No treatment-related tumours were observed.

In a chronic toxicity study, rats received diazinon in the diet at a concentration of 0 (untreated and vehicle controls), $0.1,1.5,125$ or 250 ppm (equal to $0,0.004,0.06,5$ and $10 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.005,0.07,6$ and $12 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively) for 98/99 weeks. The NOAEL was 1.5 ppm (equal to $0.06 \mathrm{mg} / \mathrm{kg}$ bw per day), on the basis of inhibition of erythrocyte and brain acetylcholinesterase activities at 125 ppm (equal to $5 \mathrm{mg} / \mathrm{kg}$ bw per day). From the available data, there was no evidence of a tumorigenic response; however, the group size $(N=20)$ was too small to allow a conclusion to be reached on carcinogenicity.

In a combined chronic toxicity and carcinogenicity study in rats, diazinon was fed in the diet at concentrations adjusted to achieve target concentrations of $0,0.025,0.1,1.5$ and $22.5 \mathrm{mg} / \mathrm{kg}$ bw per day for 104 weeks. The NOAEL for long-term toxicity was $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte acetylcholinesterase activity at $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day. No treatment-related tumours were observed.

The overall NOAEL for chronic toxicity in rats was $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day, based on inhibition of erythrocyte acetylcholinesterase activity at $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that diazinon is not carcinogenic in mice or rats.
Given the similarity of the sensitivities of mammalian species, an overall NOAEL in all studies of repeated-dose (neuro)toxicity in rats and dogs could be identified. The overall NOAEL was $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day, on the basis of inhibition of acetylcholinesterase activity in erythrocytes at $1 \mathrm{mg} / \mathrm{kg}$ bw per day.

In studies submitted by the sponsors, diazinon was tested for genotoxicity in an adequate range of assays, both in vitro and in vivo. In addition, many studies with diazinon were described in the published literature, but most of these were considered by the Meeting as inappropriate to evaluate the genotoxicity of diazinon, as they had major deficiencies in study design or reliability (e.g. lack of statistical analysis, testing of mixtures of diazinon with other chemicals and similarity between negative and positive control values). Overall, these studies provided no convincing evidence of genotoxic effects.

The Meeting concluded that diazinon is unlikely to be genotoxic.
In the multigeneration and developmental toxicity studies, cholinesterase activity was not measured.

In a two-generation study on reproductive toxicity, rats received diazinon in the diet at a concentration of $0,10,100$ or 500 ppm over the course of two generations ( $\mathrm{F}_{0}$ and $\mathrm{F}_{1}$ ). Mean diazinon intakes for the $F_{0}$ generation during the premating period were $0,0.77,7.48$ and $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,0.77,7.48$ and $40.26 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively. The NOAEL for reproductive effects was 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day), based on prolonged gestation duration, decrease in the number of pregnancies, and reduced fertility and mating indices at 500 ppm (equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for parental effects was 10 ppm (equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day), based on reduced parental body weight gain at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for offspring toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on reduced viability of pups and pup weights at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day).

In another two-generation study on reproductive toxicity, rats received diazinon in the diet at a concentration of $0,0.1,1.0$ or $10 \mathrm{mg} / \mathrm{kg}$ (equivalent to $0,0.0067,0.067$ and $0.67 \mathrm{mg} / \mathrm{kg}$ bw per day, assuming concentrations are in $\mathrm{mg} / \mathrm{kg}$ feed or ppm ) over the course of two generations ( $\mathrm{F}_{0}$ and $\mathrm{F}_{1}$ ). A rationale for the dose selection was not provided. There were no treatment-related effects observed in $F_{0}$ or $F_{1}$ parental animals or pups. The NOAEL for reproductive, parental and offspring toxicity was 10 ppm (equivalent to $0.67 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

In a range of studies on estrogenic and androgenic activities, no estrogenic, androgenic or antiandrogenic activity was observed at concentrations relevant to human exposure via the diet.

Overall NOAELs from the multigeneration studies in rats were identified. The overall NOAEL for reproductive effects was 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 500 ppm (equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for parental toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for offspring toxicity was 10 ppm (equal to $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day), based on effects at 100 ppm (equal to $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a study of developmental toxicity evaluated by the 1993 JMPR, rats were administered diazinon via gavage at a dose of $0,15,50$ or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. A marked decrease in maternal feed consumption correlating with weight loss at the beginning of the treatment period and a slightly higher incidence of incomplete ossification at different sites in the fetuses were observed at 100 $\mathrm{mg} / \mathrm{kg}$ bw per day. As limited information was available from the previous JMPR monograph, the Meeting was unable to identify a NOAEL for this study.

In a study of developmental toxicity, rats were administered diazinon via gavage at a dose of 0 , 10,20 or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was $20 \mathrm{mg} / \mathrm{kg}$ bw per day, based on body weight loss on gestation days $6-10$, reduced body weight/body weight gains throughout treatment and decreased feed consumption on gestation days $6-9$ at $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo/fetal toxicity was $20 \mathrm{mg} / \mathrm{kg}$ bw per day, based on an increased incidence of rudimentary 14 th ribs at $100 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a study of developmental toxicity, rabbits were dosed with diazinon via gavage at $0,7,25$ or $100 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was $25 \mathrm{mg} / \mathrm{kg}$ bw per day, based on mortality, tremors, convulsions, hypoactivity, anorexia and reduced body weight gain observed at 100 $\mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo/fetal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

In another developmental toxicity study, diazinon was administered to pregnant rabbits by gavage at a dose level of $0,2.5,10$ or $40 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for maternal toxicity was 10 $\mathrm{mg} / \mathrm{kg}$ bw per day, based on clinical signs, decreased body weight and reduced feed consumption. The NOAEL for embryo/fetal toxicity was $10 \mathrm{mg} / \mathrm{kg}$ bw per day, based on decreased fetal weight at $40 \mathrm{mg} / \mathrm{kg}$ bw per day.

The overall NOAEL for maternal toxicity in developmental toxicity studies in rabbits was 25 $\mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $40 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall NOAEL for embryo/fetal toxicity was $10 \mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $40 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that diazinon is not teratogenic.
In a limited acute neurotoxicity study in which acetylcholinesterase activity was not measured, rats were dosed with diazinon at $0,100,300$ or $500 \mathrm{mg} / \mathrm{kg}$ bw by gavage. The NOAEL was 100 $\mathrm{mg} / \mathrm{kg}$ bw, based on systemic toxicity and clinical signs of neurotoxicity observed at 300 or 500 $\mathrm{mg} / \mathrm{kg}$ bw. In another acute toxicity study, rats were administered a single dose of diazinon by gavage at $0,2.5,150,300$ or $600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, on the basis of depressed erythrocyte acetylcholinesterase activity and behavioural changes at $150 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. In a third study, rats were administered a single dose of diazinon by gavage at 100,250 or $500 \mathrm{mg} / \mathrm{kg}$ bw for males or $0,0.05,0.12,0.25,2.5,25$ or $250 \mathrm{mg} / \mathrm{kg}$ bw for females. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg}$ bw, on the basis of inhibition of brain and erythrocyte acetylcholinesterase activities in females at $25 \mathrm{mg} / \mathrm{kg}$ bw.

In a study that investigated the time course of acute inhibition of acetylcholinesterase activity, rats were given a single dose of diazinon by gavage at $0,2.5,150,300$ or $600 \mathrm{mg} / \mathrm{kg}$ bw, and brain and blood samples were collected at 3,9 and 24 hours after dosing. The NOAEL was $2.5 \mathrm{mg} / \mathrm{kg}$ bw, based on inhibition of brain and erythrocyte acetylcholinesterase activities at $150 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. Inhibition was observed beginning at 3 hours post-dosing, with maximal inhibition at 9 hours postdosing.

The overall NOAEL in all studies of acute toxicity was $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, on the basis of inhibition of acetylcholinesterase activity in erythrocytes and in the brain at $25 \mathrm{mg} / \mathrm{kg}$ bw in rats of both sexes.

Three studies were performed on delayed neurotoxicity in the hen. Oral doses of diazinon technical ranging from 10 to $100 \mathrm{mg} / \mathrm{kg}$ bw were administered to hens. Inhibition of cholinesterase activity was observed from $20 \mathrm{mg} / \mathrm{kg}$ bw, but there was no evidence that diazinon caused acute delayed neurotoxicity in the hen.

No specific studies on immunotoxicity were submitted. A study in the open literature with intraperitoneal injection of diazinon in mice was not informative. The submitted repeated-dose toxicity studies do not indicate an immunotoxic potential for diazinon after oral exposure.

## Toxicological data on metabolites and/or degradates

No toxicological data were available on any metabolites of diazinon other than diazoxon, which is the active metabolite of diazinon. However, the Meeting concluded that none of the other metabolites would be of toxicological concern at the levels present in the diet.

## Human data

In a study of acute toxicity in male volunteers given ascending doses of diazinon (seven volunteers per group given $0.03,0.12,0.20$ or $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$; one volunteer given $0.30 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ), acetylcholinesterase activity was not inhibited in erythrocytes at $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, the second highest dose tested. The highest dose ( $0.30 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ) was not informative, as it was tested in a single volunteer only. Plasma cholinesterase activity was inhibited by more than $20 \%$ at doses above 0.12 $\mathrm{mg} / \mathrm{kg}$ bw.

Repeated-dose studies in four male volunteers given diazinon for 28-37 days showed that, although there was some inhibition of plasma cholinesterase activity at the highest tested dose of $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day (actual administered doses varied slightly, i.e. $0.03,0.027,0.022 / 0.027$ and $0.026 \mathrm{mg} / \mathrm{kg}$ bw per day), no inhibition of erythrocyte acetylcholinesterase activity was observed.

Diazinon was evaluated in four male volunteers who received diazinon in capsules at 0.025 $\mathrm{mg} / \mathrm{kg}$ bw per day for 37-43 days. There were no consistent treatment-related effects on erythrocyte acetylcholinesterase activity, blood chemistry or urine analysis. No clinical effects were reported. The NOAEL was $0.025 \mathrm{mg} / \mathrm{kg}$ bw per day, the only dose tested.

The overall NOAEL from repeated-dose studies in humans was $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day.
Several epidemiological studies on cancer outcomes following occupational exposure to diazinon were available. The review of these studies focused on the occurrence of three cancer types: NHL, leukaemia and lung cancer (see section 2.2). One prospective cohort study was available, the Agricultural Health Study (AHS), with a large sample size and detailed exposure assessment. Cohort studies are considered a powerful design, as recall bias is avoided. All other studies were casecontrol studies, usually retrospective, which are more prone to recall and selection biases.

There was no significant evidence of a positive association of NHL with diazinon exposure and no evidence of an exposure-response relationship in the AHS. In a large pooled case-control study, the unadjusted estimates showed a significant elevated risk of NHL (relative risk [RR] $=1.7 ; 95 \%$ confidence interval $[\mathrm{CI}]=1.2-2.5$ ) associated with ever versus never use of diazinon. However, these risks were attenuated and/or no longer significant when proxy respondents were excluded and analyses were mutually adjusted for other pesticides (malathion, fonofos). Although increasing risk across exposure duration categories was observed, which was suggestive of a duration-response pattern, confidence intervals were non-significant, wide and overlapping between categories. Two other studies reported elevated risks of NHL for ever versus never use of diazinon or high versus low
diazinon use, but confidence intervals were wide, reflecting uncertainty in the risk estimates, and chance could not be excluded as an explanation for the findings. Overall, there was no convincing evidence of a positive association between NHL and exposure to diazinon.

A significantly increased risk of leukaemia in the highest exposure category (> 38.8 lifetime days of diazinon exposure; $\mathrm{RR}=3.36 ; 95 \% \mathrm{CI}=1.08-10.49$ ) and a significant exposure-response relationship were observed in the AHS. Findings for intensity-weighted lifetime exposure days demonstrated a similar pattern, but did not reach significance. Two other studies reported nonsignificantly elevated risks of leukaemia for high versus low diazinon use and ever versus never use of diazinon, with a non-significant dose-response relationship observed using days of use per year. Overall, there is weak evidence of a positive association between leukaemia and exposure to diazinon from the AHS only. It is noted that the number of diazinon-exposed cases was low or not reported in all three available studies.

A significant $60 \%$ excess risk of lung cancer in the highest exposure category ( $>38.8$ lifetime days of diazinon exposure) and a significant trend across exposure categories were observed in the AHS. Findings for intensity-weighted lifetime exposure days demonstrated a similar pattern, but did not reach significance. A separate analysis of ever use of diazinon versus never use from the AHS found no evidence of elevated risk of lung cancer among spouses of farmers/pesticide applicators; however, there were only 15 exposed cases. One other study reported a non-significant elevated risk of lung cancer for ever versus never use of diazinon (based on 17 exposed cases). Overall, there is weak evidence of a positive association between lung cancer and exposure to diazinon from the AHS cohort study only.

In view of the lack of genotoxicity and the absence of carcinogenicity in mice and rats and considering the available epidemiological data from occupational exposure, the Meeting concluded that diazinon is unlikely to pose a carcinogenic risk to humans via exposure from the diet.

The Meeting concluded that the existing database on diazinon was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The Meeting identified inhibition of acetylcholinesterase activity as the most sensitive end-point after single or repeated doses of diazinon in all species. After considering all previously evaluated data and the new studies, the Meeting established an ADI of $0-0.003 \mathrm{mg} / \mathrm{kg}$ bw, based on the overall NOAEL of $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day from all repeated-dose toxicity studies, and using a safety factor of 100. This ADI was supported by the NOAEL of $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested, identified in repeated-dose studies that involved a limited number of male volunteers, with application of a safety factor of 10 .

In 2006, the Meeting established an ADI of $0-0.005 \mathrm{mg} / \mathrm{kg}$ bw, based on the highest NOAEL of $0.5 \mathrm{mg} / \mathrm{kg}$ bw per day for inhibition of erythrocyte acetylcholinesterase activity at $1 \mathrm{mg} / \mathrm{kg}$ bw per day in a 92 -day repeated-dose toxicity study in rats and using a safety factor of 100 . In this study, the dietary concentrations of diazinon were converted to units of milligrams per kilogram body weight per day using a default conversion factor; the present Meeting considers this less reliable than the conversion using feed consumption data.

The Meeting reaffirmed the ARfD of $0.03 \mathrm{mg} / \mathrm{kg}$ bw established by the 2006 JMPR. This ARfD was based on the NOAEL of $2.5 \mathrm{mg} / \mathrm{kg}$ bw identified in studies of acute (neuro)toxicity in rats, and using a safety factor of 100 . This ARfD was supported by the NOAEL of $0.21 \mathrm{mg} / \mathrm{kg}$ bw, the highest dose tested, identified in the study in which a limited number of male volunteers were given a single dose of diazinon, with application of a safety factor of 10 .

A toxicological monograph was prepared.

## Levels relevant to risk assessment of diazinon

| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
| Mouse | Two-year study of carcinogenicity ${ }^{\mathrm{a}, \mathrm{b}}$ | Toxicity | 200 ppm , equal to 31 $\mathrm{mg} / \mathrm{kg}$ bw per day | 300 ppm , equal to 46 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 300 ppm , equal to 46 $\mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
| Rat | Acute (neuro)toxicity studies ${ }^{\mathrm{d}, \mathrm{e}}$ (acetylcholinesterase inhibition) | Toxicity | $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ | $25 \mathrm{mg} / \mathrm{kg}$ bw |
|  | Four-week or 3-month studies of (neuro)toxicity ${ }^{\text {a,e }}$ | Toxicity | 5 ppm , equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{f}}$ | 10 ppm , equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-year studies of toxicity and carcinogenicity ${ }^{\text {a,e }}$ | Toxicity <br> Carcinogenicity | $0.1 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{f}}$ <br> 800 ppm , equivalent to $40 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\mathrm{c}}$ | $1.5 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-generation studies of reproductive toxicity ${ }^{\text {a,b, }}$ | Reproductive toxicity | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day | 500 ppm , equal to $32.85 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Parental toxicity | 10 ppm , equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | 10 ppm , equal to 0.77 $\mathrm{mg} / \mathrm{kg}$ bw per day | 100 ppm , equal to 7.48 $\mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity study ${ }^{\text {b,d }}$ | Maternal toxicity <br> Embryo and fetal toxicity | $20 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $20 \mathrm{mg} / \mathrm{kg}$ bw per day | $100 \mathrm{mg} / \mathrm{kg}$ bw per day $100 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity studies ${ }^{\text {b,d,e }}$ | Maternal toxicity <br> Embryo and fetal toxicity | $25 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $10 \mathrm{mg} / \mathrm{kg}$ bw per day | $40 \mathrm{mg} / \mathrm{kg}$ bw per day <br> $40 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Dog | Ninety-day and 1-year studies of toxicity ${ }^{\text {a,e }}$ | Toxicity | $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {f }}$ | $3 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rat, dog | Repeat-dose (neuro)toxicity studies ${ }^{\text {e }}$ | Toxicity | 5 ppm , equal to 0.3 $\mathrm{mg} / \mathrm{kg}$ bw per day | 10 ppm , equivalent to $1 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Human | Acute toxicity study ${ }^{\text {d }}$ | Toxicity | $0.21 \mathrm{mg} / \mathrm{kg} \mathrm{bw}^{\text {c }}$ | - |
|  | Four/five-week studies of toxicity ${ }^{\text {d,e }}$ | Toxicity | $0.03 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |

[^16]Estimate of acceptable daily intake (ADI)
$0-0.003 \mathrm{mg} / \mathrm{kg}$ bw

## Estimate of acute reference dose (ARfD)

$0.03 \mathrm{mg} / \mathrm{kg}$ bw

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

## Critical end-points for setting guidance values for exposure to diazinon

| Absorption, distribution, excretion and metabolism in mammals |  |
| :--- | :--- |
| Rate and extent of oral absorption | Nearly complete and rapid $(\sim 90 \%$ at $10 \mathrm{mg} / \mathrm{kg}$ bw within 24 h$)$ |
| Dermal absorption | No data |
| Distribution | Widely distributed at low concentrations |
| Potential for accumulation | No potential for accumulation |
| Rate and extent of excretion | Predominantly in urine ( $86-93 \%$ at $10 \mathrm{mg} / \mathrm{kg}$ bw within 24 h$)$ <br> Metabolism in animals |
| Rapidly degraded to diazoxon and subsequently mainly via <br> oxidase/hydrolase-mediated cleavage of the ester bond, and <br> further oxidation at the isopropyl substituent to yield hydroxy <br> pyrimidinols |  |
| Toxicologically significant compounds in <br> animals and plants | Parent compound and diazoxon |

Acute toxicity

| Rat, $\mathrm{LD}_{50}$, oral | 300 to $>2150 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Rat, $\mathrm{LD}_{50}$, dermal | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| Rat, $\mathrm{LC}_{50}$, inhalation | $3.1 \mathrm{mg} / \mathrm{L}$ |
| Rabbit, dermal irritation | Mildly irritating |
| Rabbit, ocular irritation | Mildly irritating |
| Guinea-pig, dermal sensitization | Sensitizing (Magnusson and Kligman maximization test) |
| Repeat-dose studies of (neuro)toxicity |  |
| Target/critical effect | Acetylcholinesterase inhibition |
| Overall oral NOAEL | $0.3 \mathrm{mg} / \mathrm{kg}$ bw per day (rat, dog) |
| Lowest relevant dermal NOAEL | $3 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rat) |
| Lowest relevant inhalation NOAEC | $0.46 \mathrm{mg} / \mathrm{m}^{3}(21$ days; rat) |

Long-term studies of carcinogenicity
Carcinogenicity Not carcinogenic in mice or rats ${ }^{a}$
Genotoxicity
No evidence of genotoxicity by the oral route ${ }^{a}$

| Reproductive toxicity |  |  |
| :---: | :---: | :---: |
| Target/critical effect | Mortality, reduced parental body weig of pups and pup weights, prolonged ge decrease in number of pregnancies, and mating indices | reduced viability duration, ed fertility and |
| Lowest relevant parental NOAEL | $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Lowest relevant offspring NOAEL | $0.77 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Lowest relevant reproductive NOAEL | $7.48 \mathrm{mg} / \mathrm{kg}$ bw per day (rat) |  |
| Developmental toxicity |  |  |
| Target/critical effect | Clinical signs, reduced maternal body consumption, and reduced fetal weight | and feed |
| Lowest relevant maternal NOAEL | $25 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) |  |
| Lowest relevant embryo/fetal NOAEL | $10 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) |  |
| Neurotoxicity ${ }^{\text {b }}$ |  |  |
| Acute neurotoxicity NOAEL | $2.5 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ (acetylcholinesterase inh | rat) |
| Developmental neurotoxicity NOAEL | No data |  |
| Acute delayed neurotoxicity | No evidence (hens) |  |
| Human data | Acetylcholinesterase inhibition: |  |
|  | Acute toxicity NOAEL: $0.21 \mathrm{mg} / \mathrm{kg}$ bw, | st dose tested |
|  | Subchronic toxicity NOAEL: 0.03 mg / dose tested ( $4 / 5$ weeks) | per day, highest |
| ${ }^{\text {a }}$ Unlikely to pose a carcinogenic risk to humans via exposure from the diet. <br> ${ }^{\mathrm{b}}$ Ninety-day neurotoxicity study in rats is covered by the overall NOAEL for repeated-dose studies of |  |  |
| Summary |  |  |
| Value | Study | Safety factor |
| ADI $\quad 0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ | Repeated-dose toxicity studies (rat, dog) | 100 |
| ARfD $\quad 0.03 \mathrm{mg} / \mathrm{kg}$ bw | Acute (neuro)toxicity studies (rat) | 100 |

## DIETARY RISK ASSESSMENT

## Long-term dietary exposure

The ADI for diazinon is $0-0.003 \mathrm{mg} / \mathrm{kg}$ bw. The international estimated daily intakes (IEDIs) for diazinon were estimated for the 17 Global Environment Monitoring System - Food Contamination Monitoring and Assessment Programme (GEMS/Food) cluster diets using the supervised trials median residue (STMR) or STMR in a processed commodity (STMR-P) values estimated by the 1996 (animal commodities), 1999 (pome fruit, cabbage head) and 2006 (cranberries) JMPRs. An STMR value for tomato was estimated using the data reported in the 1993 JMPR evaluation monograph. For all other commodities, the maximum residue limits (MRLs) were used, as STMR values were not available. The results are shown in Annex 3. The IEDI ranged from $3 \%$ to $50 \%$ of the
maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of diazinon from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The ARfD for diazinon is $0.03 \mathrm{mg} / \mathrm{kg}$ bw. The international estimate of short-term dietary intake (IESTI) was calculated. The calculation employed highest residue (HR) values where these could be identified in the relevant JMPR reports; otherwise, the MRL was used. In the case of meat, the Meeting noted that residues in fat are approximately 15 times higher than those in muscle and used the MRL value of $2 \mathrm{mg} / \mathrm{kg}$ for fat and $0.1333 \mathrm{mg} / \mathrm{kg}$ for muscle. The results are shown in Annex 4. The IESTI represented a maximum of $100 \%$ of the ARfD for both children and the general population. The Meeting concluded that the short-term dietary exposure to diazinon residues from uses considered by JMPR was unlikely to present a public health concern.

### 3.2 GLYPHOSATE (158)

## TOXICOLOGY

Glyphosate is the ISO-approved common name for $N$-(phosphonomethyl)glycine (IUPAC), with CAS number 1071-83-6. It is a broad-spectrum systemic herbicide.

Glyphosate was previously evaluated by JMPR for toxicology in 1986, 1997 (evaluation of the metabolite aminomethylphosphonic acid, or AMPA), 2004 and 2011 (evaluation of new plant metabolites in genetically modified maize and soya beans).

Glyphosate was last re-evaluated for toxicology within the periodic review programme of CCPR in 2004. The compound was reviewed by the present Meeting following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously considered toxicological data in addition to new published or unpublished toxicological studies and published epidemiological studies on cancer outcomes. The evaluation of the biochemical aspects and systemic toxicity of glyphosate was based on previous JMPR evaluations, updated as necessary with additional information. The particular focus of the current meeting was on genotoxicity, carcinogenicity, reproductive and developmental toxicity and epidemiological studies on cancer outcomes. The scope was restricted to the active ingredient.

All critical unpublished studies contained statements of compliance with GLP, unless otherwise specified. The studies on human volunteers were conducted in accordance with the principles expressed in the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

In studies with radiolabelled glyphosate in rats, glyphosate was rapidly absorbed from the gastrointestinal tract following oral intake, but only to a limited extent (about 20-30\%). Elimination was fast and virtually complete within $72-168$ hours, with the majority being excreted during the first 48 hours. Most of the excretion occurred in faeces, largely as unabsorbed dose, and in the urine. Biliary excretion of glyphosate was negligible. Less than $1 \%$ of the administered dose was retained in tissues 168 hours post-administration. Highest residues were detected in bone, followed by kidney and liver. This pattern of absorption, distribution and elimination was independent of dose, treatment regimen and sex of the test animals. Peak plasma concentrations of radiolabel were observed at 6 and 2 hours after administration in male and female rats, respectively. The estimated half-life for wholebody elimination of the radiolabel was about 5.9-8.3 hours.

There was very little biotransformation of glyphosate; the only metabolite, AMPA, accounted for $0.2-0.7 \%$ of the administered dose in excreta; the rest was unchanged glyphosate.

## Toxicological data

Glyphosate has low acute oral toxicity in mice $\left(\mathrm{LD}_{50}>2000\right.$ to $>10000 \mathrm{mg} / \mathrm{kg}$ bw; no lethality at $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ ) and rats $\left(\mathrm{LD}_{50} 5600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}\right)$, low acute dermal toxicity in rats $\left(\mathrm{LD}_{50}>2000\right.$ $\mathrm{mg} / \mathrm{kg} \mathrm{bw})$ and rabbits $\left(\mathrm{LD}_{50}>5000 \mathrm{mg} / \mathrm{kg}\right.$ bw$)$, and low acute inhalation toxicity in rats $\left(\mathrm{LC}_{50}>5.48\right.$ $\mathrm{mg} / \mathrm{L})$. Glyphosate was not irritating to the skin of rabbits. Glyphosate produced moderate to severe eye irritation in rabbits, with irreversible corneal opacity in one study as a consequence of the low pH of the test material in solution. Glyphosate was not sensitizing in guinea-pigs or mice as determined by the Magnusson and Kligman maximization test, the Buehler test and the local lymph node assay.

In short-term studies of toxicity in different species, the most notable effects were clinical signs related to gastrointestinal irritation, decreased body weight, salivary gland changes (hypertrophy and increase in basophilia of cytoplasm of acinar cells), histological findings in the caecum and hepatotoxicity.

In short-term studies in mice, reduced body weight was seen at a dietary concentration of 50000 ppm (equal to $9710 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for decreased body weight was 10000 ppm (equal to $1221 \mathrm{mg} / \mathrm{kg}$ bw per day). Effects on the salivary glands were observed in mice in only one study out of four at 6250 ppm (equal to $1065 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for the salivary gland effects in mice was 3125 ppm (equal to $507 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL in shortterm studies in mice was 3125 ppm (equal to $507 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 6250 ppm (equal to $1065 \mathrm{mg} / \mathrm{kg}$ bw per day).

In 90-day toxicity studies in rats, common findings included soft faeces, diarrhoea, reduced body weight gain and decreased food utilization at dietary concentrations of 20000 ppm (equal to $1262.1 \mathrm{mg} / \mathrm{kg}$ bw per day) and above. The lowest NOAEL was $371.9 \mathrm{mg} / \mathrm{kg}$ bw per day. A decrease in urine pH was frequently noted owing to the acidic nature of the compound and excretion as glyphosate in the urine. In two 90-day dietary toxicity studies, an increase in caecum weight (at 10000 ppm , equal to $569 \mathrm{mg} / \mathrm{kg}$ bw per day) and histological findings in the caecum (at 50000 ppm , equal to $3706 \mathrm{mg} / \mathrm{kg}$ bw per day) were observed. In rats, effects on the salivary gland were seen in two out of seven 90-day studies starting at 12500 ppm (equal to $811 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAELs for effects on the salivary gland were 300 and $410 \mathrm{mg} / \mathrm{kg}$ bw per day. The overall NOAEL in short-term studies in rats was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall LOAEL was 10000 ppm (equal to $569 \mathrm{mg} / \mathrm{kg}$ bw per day).

In four 90-day toxicity studies in dogs, the most notable effects were loose stools, decreased body weight and reduced feed consumption. In one study, there were no treatment-related effects at doses up to 40000 ppm (equal to $1015 \mathrm{mg} / \mathrm{kg}$ bw per day). The lowest NOAEL and LOAEL were $300 \mathrm{mg} / \mathrm{kg}$ bw per day and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively.

Seven 1-year toxicity studies in dogs are available. In one study, changes in faeces were observed at $100 \mathrm{mg} / \mathrm{kg}$ bw per day and above. The NOAEL was $30 \mathrm{mg} / \mathrm{kg}$ bw per day. However, these results were not reproduced in four other studies with administration via capsules at 300 or 500 $\mathrm{mg} / \mathrm{kg}$ bw per day. In the remaining six studies, the NOAELs ranged from 8000 ppm (equal to 182 $\mathrm{mg} / \mathrm{kg}$ bw per day) to $500 \mathrm{mg} / \mathrm{kg}$ bw per day, and the LOAELs ranged from 30000 ppm (equal to 926 $\mathrm{mg} / \mathrm{kg}$ bw per day) to $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

The overall NOAEL in the 90-day and 1-year toxicity studies in dogs was 15000 ppm (equal to $448 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 30000 ppm (equal to $926 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting compiled the tumour incidence data for all relevant mouse and rat studies in order to undertake statistical analysis and investigate any potential pattern of occurrence across studies. In addition, incidences of tumours of lymphatic tissues were summarized, as these were identified as possible targets of relevance from the review of epidemiological cancer studies. However, the Meeting recognized that the relationship between tumours of lymphatic tissues in rodents and humans has not been clearly established.

Nine carcinogenicity studies in mice were available. Two studies were considered to be of insufficient quality to be included in the assessment. Effects such as loose stools, reduced body weights and decreased feed consumption were noted in most of the studies. The overall NOAEL for systemic toxicity in mice was 1600 ppm (equal to $153 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 8000 ppm (equal to $787 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that there is equivocal evidence of induction of lymphomas in male mice in three out of seven studies and in female mice in one out of seven studies at high doses (500040000 ppm , equal to $814-4348 \mathrm{mg} / \mathrm{kg}$ bw per day). The Meeting also noted that in the other three studies in which even higher doses (up to 50000 ppm , equal to $7470 \mathrm{mg} / \mathrm{kg}$ bw per day) had been used, no effect was observed.

The Meeting concluded that there is some indication, by a trend test, and not by pairwise comparison, of induction of kidney adenomas in male mice in four out of seven studies. The Meeting noted that the increases were marginal and occurred at the highest dose only and that other studies that used appreciably higher doses did not find any excess. However, the Meeting noted that kidney adenomas are uncommon in male mice.

Eleven combined chronic toxicity and carcinogenicity studies in rats were available. One study was considered to be inadequate for carcinogenicity assessment due to its exposure duration (12 months). Toxicities variously reported in some of these studies included increased incidences of clinical signs, reduced body weights, degenerative lens changes (cataracts) in males, microscopic findings in the salivary gland, increased incidence of basophilia of parotid acinar cells, and microscopic findings in liver, prostate and kidneys. The overall NOAEL for systemic toxicity in rats was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, and the overall LOAEL was $300 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting discussed the increased incidence of a variety of tumours observed in one or, in one case, two of the 10 studies in rats. The Meeting concluded that these findings were incidental, based on the following considerations:

- interstitial cell tumours of the testes: occurred in only one study; and other studies that used appreciably higher doses did not find any excess;
- pancreatic islet cell adenoma: occurred in only one study in males only; other studies that used appreciably higher doses did not find any excess; there was no dose-response relationship; and the incidence in controls was unusually low (less than the lower bound of the historical control data); the Meeting also noted that there was a negative dose-response relationship in females;
- thyroid C-cell tumours: occurred in only one study; other studies that used appreciably higher doses did not find any excess; and these tumours are considered not to be relevant for humans;
- skin keratoma: occurred in two studies in males only; other studies that used appreciably higher doses did not find any excess; in one study, there was no dose-response relationship; and in the other study, only the test for trend was statistically significant, not the pairwise test at any dose;
- lymphoma (in spleen and kidney): no evidence of induction in any of the studies.

The Meeting concluded that there is no reliable evidence for treatment-related tumours in rats at doses up to 32000 ppm (equal to $1750 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that glyphosate is not carcinogenic in rats but could not exclude the possibility that it is carcinogenic in mice at very high doses.

Glyphosate and its formulation products have been extensively tested for genotoxic effects using a variety of tests in a wide range of organisms. While no mutational effects have been detected in bacterial test systems, DNA damage and chromosomal effects have commonly been seen in cell culture models and in organisms that are phylogenetically distant from humans. However, these effects have not been seen in vivo in orally treated mammalian models. The overall weight of evidence indicates that administration of glyphosate and its formulation products at doses as high as $2000 \mathrm{mg} / \mathrm{kg}$ bw by the oral route, the route most relevant to human dietary exposure, was not associated with genotoxic effects in an overwhelming majority of studies conducted in mammals, a model considered to be appropriate for assessing genotoxic risks to humans.

The Meeting concluded that glyphosate is unlikely to be genotoxic at anticipated dietary exposures.

Seven reproductive toxicity studies in rats were available. No evidence of reproductive toxicity was observed at doses up to 30000 ppm (equal to $1983 \mathrm{mg} / \mathrm{kg}$ bw per day). In one study, an
increased incidence of histopathological findings in the parotid (males) and submaxillary salivary glands in females was observed in both generations at 10000 ppm (equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL was 3000 ppm (equal to $197 \mathrm{mg} / \mathrm{kg}$ bw per day). In a separate study, an increased incidence of loose stools and caecum distension was observed in both generations at 30000 ppm (equal to $2150 \mathrm{mg} / \mathrm{kg}$ bw per day), and the NOAEL was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day). Slight reductions in pup weight or weight gain were observed in most studies, but were confined to very high, parentally toxic dose levels. In addition, a significant delay in sexual maturation in male pups ( $\mathrm{F}_{1}$ ) was seen at 15000 ppm (equal to $1063 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for parental toxicity was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 10000 ppm (equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day). The overall NOAEL for offspring toxicity was 6000 ppm (equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day), and the overall LOAEL was 10000 ppm (equal to $985 \mathrm{mg} / \mathrm{kg}$ bw per day).

No evidence of teratogenicity was observed in four developmental toxicity studies in rats at doses up to $3500 \mathrm{mg} / \mathrm{kg}$ bw per day. There was some variation in the extent of toxicity observed in the four studies. The lowest NOAEL for maternal toxicity was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, based on loose stools and reduced body weights seen at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day. The lowest NOAEL for embryo and fetal toxicity was $300 \mathrm{mg} / \mathrm{kg}$ bw per day, based on delayed ossification and an increased incidence of fetuses with skeletal anomalies observed at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

Seven developmental toxicity studies in the rabbit were available. Maternal toxicity was primarily manifested as an increased incidence of soft stools and diarrhoea at doses of $175 \mathrm{mg} / \mathrm{kg}$ bw per day and above. The overall NOAEL for maternal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day. In three studies, the occurrences of a variety of low-incidence fetal effects (e.g. cardiac malformation, absent kidney) were slightly increased at higher dose levels. These increases are considered secondary to maternal toxicity. The overall NOAEL for embryo and fetal toxicity was $250 \mathrm{mg} / \mathrm{kg}$ bw per day, based on effects at $450 \mathrm{mg} / \mathrm{kg}$ bw per day. The Meeting considered that these effects were secondary to local irritation from unabsorbed glyphosate in the colon administered by gavage dosing and concluded that they were not relevant for establishing health-based guidance values.

The Meeting concluded that glyphosate is not teratogenic.
Glyphosate was tested in a range of validated in vivo and in vitro assays for its potential to interact with the endocrine system. The studies that the Meeting considered adequate for the evaluation clearly demonstrate that there is no interaction with estrogen or androgen receptor pathways or thyroid pathways.

There was no evidence of neurotoxicity in an acute neurotoxicity study in rats at doses up to $2000 \mathrm{mg} / \mathrm{kg}$ bw. The NOAEL for systemic toxicity was $1000 \mathrm{mg} / \mathrm{kg}$ bw, based on a single death and general signs of toxicity at $2000 \mathrm{mg} / \mathrm{kg}$ bw. In a 90 -day neurotoxicity study in rats, no evidence of neurotoxicity or systemic toxicity was seen at doses up to 20000 ppm (equal to $1546.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

No evidence of immunotoxicity was seen in a 28 -day dietary study in female mice at doses up to 5000 ppm (equal to $1448 \mathrm{mg} / \mathrm{kg}$ bw per day).

Effects on the salivary glands were observed in several repeated-dose toxicity studies in rats. The pH of glyphosate in solution is low, and it has been shown that exposure to organic acids can cause such changes in salivary glands. Therefore, the changes are likely secondary to the effects caused by the pH of the test compound in solution.

In many of the long-term repeated-dose studies reviewed, glyphosate was reported to have an impact on the gastrointestinal tract at high doses. Although this is not uncommon with high-dose chemical substance administration, this was investigated further, as glyphosate is known to be poorly absorbed in mammalian models, and alterations in gut microbiota profiles, specifically reductions in the beneficial microbiota and increases in pathogenic bacteria, are known to have impacts on carcinogenesis. There is evidence from livestock species that pathogenic bacteria are more resistant to glyphosate, whereas beneficial microbiota are more sensitive, and thus more vulnerable.

This is an emerging area of scientific investigation. The extent to which glyphosate adversely affects the normal functioning of the microbiota in the human gastrointestinal tract or the gastrointestinal tract of mammalian models is unclear. However, it is unlikely, given the available information on minimum inhibitory concentration (MIC) values, that this would occur from glyphosate residues in the diet.

## Toxicological data on metabolites and/or degradates

AMPA is the only identified metabolite found in the urine and faeces of orally treated rats. AMPA was of low acute oral and dermal toxicity in rats $\left(\mathrm{LD}_{50}>5000\right.$ and $>2000 \mathrm{mg} / \mathrm{kg}$ bw, respectively) and was not sensitizing in guinea-pigs, as determined by the Magnusson and Kligman maximization test. In a 90-day study of toxicity in rats, the NOAEL was $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested. AMPA administered orally in mammalian test systems showed no evidence of genotoxicity. Largely negative results were seen in studies in vitro. The Meeting concluded that AMPA is unlikely to be genotoxic in vivo by the oral route. In a study of developmental toxicity in rats, no evidence for embryo or fetal toxicity was observed; the NOAEL for maternal and embryo/fetal toxicity was 1000 $\mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

Following single gavage administration of radiolabelled $N$-acetyl-glyphosate, a plant-specific metabolite, at $15 \mathrm{mg} / \mathrm{kg}$ bw in rats, about $66.1 \%$ of the administered dose was excreted in urine ( $61.3 \%$ within 12 hours post-dosing), $26.4 \%$ in faeces ( $25.8 \%$ within 48 hours post-dosing), $2.79 \%$ in cage wash and wipe, and $0.23 \%$ in residual carcass. Radioactivity was eliminated rapidly from blood and plasma, with half-life values of 20.1 and 15.6 hours, respectively. Unchanged $\left[{ }^{14} \mathrm{C}\right] \mathrm{N}$-acetylglyphosate recovered in urine and faeces represented over $99 \%$ of the administered radioactivity. Glyphosate, a metabolite of $N$-acetyl-glyphosate, was detected in faeces and represented less than $1 \%$ of the total radioactivity.

The acute oral toxicity $\mathrm{LD}_{50}$ of $N$-acetyl-glyphosate in rats is greater than $5000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, expressed as the free acid. In a 90-day toxicity study in rats, the NOAEL was 18000 ppm (equal to $1157 \mathrm{mg} / \mathrm{kg}$ bw per day).
$N$-Acetyl-glyphosate was tested for genotoxicity in vitro and in vivo in an adequate range of assays; it was not found to be genotoxic in mammalian or microbial test systems.

The Meeting concluded that $N$-acetyl-glyphosate is unlikely to be genotoxic.
$N$-Acetyl-AMPA, another plant-specific metabolite, was of low acute oral toxicity; the $\mathrm{LD}_{50}$ was greater than $5000 \mathrm{mg} / \mathrm{kg}$ bw in rats.
$N$-Acetyl-AMPA was tested for genotoxicity in vitro and in vivo in an adequate range of assays; it was not found to be genotoxic in mammalian or microbial test systems.

The Meeting concluded that $N$-acetyl-AMPA is unlikely to be genotoxic.

## Human data

Routine medical surveillance of workers in production and formulation plants revealed no adverse health effects attributable to glyphosate. In operators applying glyphosate products, cases of eye, skin and/or respiratory tract irritation have been reported. Acute intoxication was reported in humans after accidental or intentional ingestion of concentrated glyphosate formulations, resulting in gastrointestinal, cardiovascular, pulmonary and renal effects and, occasionally, death. The acute toxicity of glyphosate formulations was likely caused by the surfactant in these products.

Several epidemiological studies on cancer outcomes following occupational exposure to glyphosate were available. The evaluation of these studies focused on the occurrence of NHL, as outlined in section 2.2. One meta-analysis and one prospective cohort study, the AHS, with a large sample size and detailed exposure assessment, were available. Cohort studies are considered a
powerful design, as recall bias is avoided. All other studies were case-control studies, usually retrospective, which are more prone to recall and selection biases.

The AHS cohort study found no evidence of a positive association of NHL with glyphosate exposure or an exposure-response relationship. Elevated risks were reported in various case-control studies. A significant elevated risk of NHL associated with ever versus never use of glyphosate (odds ratio $[\mathrm{OR}]=2.1 ; 95 \% \mathrm{CI}=1.1-4.0$ ) was reported. Ever use of glyphosate was not associated with risk of NHL in the cross-Canada case-control study of pesticides and health, but when analysing days of use per year, there was a significant elevated risk in the highest usage category ( $\mathrm{OR}=2.12$; $95 \%$ $\mathrm{CI}=1.20-3.73$; for $>2$ days/year glyphosate use). There was, however, no indication of an exposure-response relationship across exposure usage categories. In another case-control study, a significant increased risk of NHL associated with ever use ( $\mathrm{OR}=2.02 ; 95 \% \mathrm{CI}=1.10-3.71$ ) as well as the highest usage category ( $\mathrm{OR}=2.36 ; 95 \% \mathrm{CI}=1.04-5.37$; for greater than 10 days/year glyphosate use) was observed, with some suggestion of an exposure-response gradient. Two smaller case-control studies with few exposed cases and limited statistical power reported a non-significant elevated risk and no association, respectively, for risk of NHL and ever use of glyphosate. The metaanalysis, including the AHS, found a significant $50 \%$ excess risk ratio for ever versus never use of glyphosate.

Overall, there is some evidence of a positive association between glyphosate exposure and risk of NHL from the case-control studies and the overall meta-analysis. However, it is notable that the AHS, which is the only cohort study and is large and of high quality, found no evidence of association at any exposure level.

In view of the absence of carcinogenic potential in rodents at human-relevant doses and the absence of genotoxicity by the oral route in mammals, and considering the epidemiological evidence from occupational exposures, the Meeting concluded that glyphosate is unlikely to pose a carcinogenic risk to humans via exposure from the diet.

The Meeting concluded that the existing database on glyphosate was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The Meeting reaffirmed the group ADI for the sum of glyphosate, AMPA, $N$-acetyl-glyphosate and $N$-acetyl-AMPA of $0-1 \mathrm{mg} / \mathrm{kg}$ bw on the basis of the NOAEL of $100 \mathrm{mg} / \mathrm{kg}$ bw per day for effects on the salivary gland in a long-term study of toxicity and carcinogenicity in rats and application of a safety factor of 100 . The Meeting noted that these effects may be secondary to local irritation due to the low pH of glyphosate in solution, but was unable to establish this unequivocally.

The Meeting concluded that it was not necessary to establish an ARfD for glyphosate, AMPA, $N$-acetyl-glyphosate and $N$-acetyl-AMPA in view of their low acute toxicity, the absence of relevant developmental toxicity in rats and rabbits that could have occurred as a consequence of acute exposure, and the absence of any other toxicological effect that would be elicited by a single dose.

A toxicological monograph was prepared.

Levels relevant to risk assessment of glyphosate

| Species | Study | Effect | NOAEL | LOAEL |
| :--- | :--- | :--- | :--- | :--- |
| Mouse | Eighteen- to 24-month | Toxicity | 1600 ppm, equal to | 8000 ppm, equal to |
|  | studies of toxicity and |  | $153 \mathrm{mg} / \mathrm{kg}$ bw per <br> carcinogenicity ${ }^{\text {a,b }}$ |  |
|  | day $^{\mathrm{c}}$ | $787 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ per |  | day |


| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Carcinogenicity | The Meeting could not exclude the possibility that glyphosate is carcinogenic in mice at very high doses. |  |
| Rat | Acute neurotoxicity study ${ }^{\text {a }}$ | Neurotoxicity | $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}^{\text {c }}$ | - |
|  | Two-year studies of toxicity and carcinogenicity ${ }^{b}$ | Toxicity | $100 \mathrm{mg} / \mathrm{kg}$ bw per day | $300 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 32000 ppm , equal to $1750 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  | Two-generation studies of reproductive toxicity ${ }^{\text {a,b }}$ | Reproductive toxicity | 30000 ppm , equal to $1983 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  |  | Parental toxicity | 6000 ppm , equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day | 10000 ppm , equal to $668 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | 6000 ppm , equal to $417 \mathrm{mg} / \mathrm{kg}$ bw per day | 10000 ppm , equal to $985 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity studies ${ }^{\text {b,d }}$ | Maternal toxicity | $300 \mathrm{mg} / \mathrm{kg}$ bw per day | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $300 \mathrm{mg} / \mathrm{kg}$ bw per day | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity studies ${ }^{\text {b,d }}$ | Maternal toxicity ${ }^{\text {e }}$ | $100 \mathrm{mg} / \mathrm{kg}$ bw per day | $175 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity ${ }^{\text {e }}$ | $250 \mathrm{mg} / \mathrm{kg}$ bw per day | $450 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Dog | Thirteen-week and 1-year studies of toxicity ${ }^{\text {b,f }}$ | Toxicity | 15000 ppm , equal to $448 \mathrm{mg} / \mathrm{kg}$ bw per day | 30000 ppm , equal to $926 \mathrm{mg} / \mathrm{kg}$ bw per day |
| AMPA |  |  |  |  |
| Rat | Thirteen-week study of toxicity ${ }^{\text {d }}$ | Toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  | Developmental toxicity study ${ }^{\text {d }}$ | Maternal toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |
|  |  | Embryo and fetal toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {c }}$ | - |

${ }^{\text {a }}$ Dietary administration.
${ }^{\text {b }}$ Two or more studies combined.
${ }^{\text {c }}$ Highest dose tested.
${ }^{\text {d }}$ Gavage administration.
${ }^{\mathrm{e}}$ Secondary to local irritation of the colon.
${ }^{f}$ Capsule administration.

## Estimate of acceptable daily intake (ADI)

$0-1 \mathrm{mg} / \mathrm{kg}$ bw (for sum of glyphosate, $N$-acetyl-glyphosate, AMPA and $N$-acetyl-AMPA)

## Estimate of acute reference dose (ARfD)

## Unnecessary

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

## Critical end-points for setting guidance values for exposure to glyphosate

Absorption, distribution, excretion and metabolism in mammals
Rate and extent of oral absorption Rapidly, but only to a limited extent (about 20-30\%)
Dermal absorption
Distribution
Potential for accumulation
Rate and extent of excretion

Metabolism in animals
Toxicologically significant compounds in animals and plants

## About 1-3\%

Widely distributed (low levels occurring in all tissues)
No evidence of accumulation
Rapid and nearly complete in 48 h (about 20-30\% in urine and about $60-70 \%$ in faeces)

Very limited ( $<0.7 \%$ ), by hydrolysis leading to AMPA
Parent compound, AMPA, $N$-acetyl-glyphosate, $N$-acetylAMPA

Acute toxicity
Rat, $\mathrm{LD}_{50}$, oral $5600 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$
Rat, $\mathrm{LD}_{50}$, dermal $\quad>2000 \mathrm{mg} / \mathrm{kg}$ bw
Rat, $\mathrm{LC}_{50}$, inhalation $\quad>5.48 \mathrm{mg} / \mathrm{L}$
Rabbit, dermal irritation Not irritating
Rabbit, ocular irritation Moderately to severely irritating
Guinea-pig, dermal sensitization Not sensitizing (Magnusson and Kligman test, Buehler test)
Mouse, dermal sensitization Not sensitizing (local lymph node assay)
Short-term studies of toxicity
Target/critical effect Clinical signs (loose stools, diarrhoea), liver, salivary glands and reduced body weights

Lowest relevant oral NOAEL
$300 \mathrm{mg} / \mathrm{kg}$ bw per day ( 90 days; rat)
Lowest relevant dermal NOAEL $\quad>5000 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rabbit)
Lowest relevant inhalation NOAEC
No data
Long-term studies of toxicity and carcinogenicity
Target/critical effect Reduced body weights, loose stools, liver (toxicity), salivary glands (organ weight, histology), eye (cataracts, lens fibre degeneration)

Lowest relevant NOAEL
$100 \mathrm{mg} / \mathrm{kg}$ bw per day (2 years; rat)
Carcinogenicity Not carcinogenic in rats; could not exclude possibility of carcinogenicity in mice at very high doses ${ }^{\text {a }}$

## Genotoxicity

No genotoxic potential via oral route in mammals ${ }^{\mathrm{a}}$

## Reproductive toxicity

Target/critical effect Reduced body weights and delayed development (absence of maternal toxicity)
Lowest relevant parental NOAEL
Lowest relevant offspring NOAEL
$417 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)
$417 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)

Lowest relevant reproductive NOAEL
Developmental toxicity

Target/critical effect
Lowest relevant maternal NOAEL
Lowest relevant embryo/fetal NOAEL
$1983 \mathrm{mg} / \mathrm{kg}$ bw per day (rat)

Neurotoxicity
Acute neurotoxicity NOAEL
Subchronic neurotoxicity NOAEL
Developmental neurotoxicity NOAEL
Slight increase in malformations at maternally toxic doses
$100 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) ${ }^{\text {b }}$
$250 \mathrm{mg} / \mathrm{kg}$ bw per day (rabbit) ${ }^{\text {b }}$

Other toxicological studies
Immunotoxicity
No immunotoxicity; NOAEL $1448 \mathrm{mg} / \mathrm{kg}$ bw per day, highest dose tested (28 days; mouse)
Studies on toxicologically relevant metabolites
Toxicological studies on AMPA, N -acetyl-glyphosate and N -acetyl-AMPA reveal the metabolites to be less toxic than the parent compound

Human data
Medical surveillance of workers in plants producing and formulating glyphosate did not reveal any adverse health effects. In operators applying glyphosate products, cases of eye, skin and/or respiratory irritation have been reported. Cases of acute intoxication have been observed after accidental or intentional ingestion of glyphosate formulation.

[^17]
## Summary

|  | Value | Study | Safety factor |
| :--- | :--- | :--- | :--- |
| ADI | $0-1 \mathrm{mg} / \mathrm{kg}$ bw | Two-year studies of toxicity (rat) | 100 |
| ARfD | Unnecessary | - | - |

## DIETARY RISK ASSESSMENT

## Long-term dietary exposure

The ADI for glyphosate is $0-1 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$. The IEDIs for glyphosate were estimated for the 17 GEMS/Food cluster diets using the STMR or STMR-P values estimated by JMPR. The results are
shown in Annex 3. The IEDI ranged from $0 \%$ to $1 \%$ of the maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of glyphosate from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The Meeting concluded that it was unnecessary to establish an ARfD for glyphosate, and therefore an IESTI for glyphosate was not calculated. The Meeting therefore concluded that short-term dietary exposure to glyphosate residues is unlikely to present a risk to consumers.

### 3.3 MALATHION (49)

## TOXICOLOGY

Malathion is the ISO-approved common name for $S$-1,2-bis(ethoxycarbonyl)ethyl $O, O$-dimethyl phosphorothioate (IUPAC), with the CAS number 121-75-5.

Malathion is a non-systemic organophosphorus insecticide whose mode of pesticidal action is the inhibition of cholinesterase activity. It is used to control insects on agricultural crops and stored commodities and for vector control.

The toxicity of malathion was evaluated by JMPR in 1963, 1965, 1966, 1997 and 2003. Malathion was listed in the periodic review programme of CCPR but was not yet scheduled for review. The compound was reviewed by the present Meeting following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that it be reevaluated due to public health concerns identified by IARC and the availability of a significant number of new studies.

The current Meeting evaluated all previously submitted toxicological data in addition to new published and unpublished toxicological studies and published epidemiological studies on cancer outcomes. All critical unpublished studies contained certificates of compliance with GLP, unless otherwise specified. Human volunteer studies were conducted according to the Declaration of Helsinki or equivalent ethical standards.

## Biochemical aspects

In a study conducted in rats using $\left[{ }^{14} \mathrm{C}\right]$ malathion, gastrointestinal absorption was at least $77 \%$ in males and $86 \%$ in females. The majority (up to $90 \%$ ) of radioactivity was excreted in urine within 24 hours. Less than $1 \%$ of radioactivity was detected in tissues, with the highest proportions in the liver, skin, fat and gastrointestinal tract. There was no evidence that malathion or its metabolites accumulated in any tissue.

Malathion is extensively metabolized via desulfuration, oxidation, hydrolysis, dealkylation and demethylation reactions. In particular, the oxidative desulfuration of malathion in the liver generates malaoxon, which is a more potent inhibitor of acetylcholinesterase compared with malathion. The major metabolites detected in rat urine ( $>80 \%$ of urinary radioactivity) were $\alpha$ - and $\beta$ monocarboxylic acids (MMCA) and the dicarboxylic acid (MDCA) of malathion. Other urinary metabolites include desmethyl malathion, $O, O$-dimethyl phosphorothioic acid, fumaric acid, 2mercaptosuccinic acid, $O, O$-dimethyl phosphorodithioic acid, monoethyl fumarate and malaoxon. Malaoxon was observed only in urine samples and accounted for less than $2 \%$ of total urinary radioactivity. Similar metabolites were detected in human studies.

Published in vitro studies have further investigated the metabolism of malathion. In human liver microsomes, the metabolism of malathion to malaoxon was catalysed by CYP1A2, CYP2B6 or CYP3A4, their respective contributions depending on the concentration of malathion. Isomalathion, a storage impurity, was a potent non-competitive inhibitor of hepatic carboxylesterase activity, important for the formation of MMCA by human liver microsomes.

Estimates of in vitro dermal absorption through human skin ranged from $1.44 \%$ to $8.74 \%$ and from $8 \%$ to $20.7 \%$. In a volunteer study, dermal absorption was $4.48 \%$ following a single application and $3.53 \%$ following a second application.

## Toxicological data

Consistent with other organophosphorus insecticides, the most sensitive toxicological effect following acute and repeated exposures to malathion is the inhibition of acetylcholinesterase activity in erythrocytes and brain. At higher doses, cholinergic signs become evident.

In rats, the oral $\mathrm{LD}_{50}$ ranged from 1539 to $8227 \mathrm{mg} / \mathrm{kg}$ bw, the dermal $\mathrm{LD}_{50}$ was greater than $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ and the inhalation $\mathrm{LC}_{50}$ was greater than $5.2 \mathrm{mg} / \mathrm{L}$. The dermal $\mathrm{LD}_{50}$ in rabbits was $8790 \mathrm{mg} / \mathrm{kg}$ bw. Malathion was slightly irritating to rabbit skin and eyes. In a Buehler test conducted in guinea-pigs, malathion did not cause skin sensitization, whereas malathion caused skin sensitization in the guinea-pig maximization test. Malathion was not sensitizing in the mouse local lymph node assay.

In a 14 -day range-finding study conducted in juvenile rats, which tested gavage malathion doses of $0,250,450$ and $600 \mathrm{mg} / \mathrm{kg}$ bw per day, salivation occurred at 450 and $600 \mathrm{mg} / \mathrm{kg}$ bw per day. In males, erythrocyte and brain acetylcholinesterase activities were reduced at every dose, whereas in females, erythrocyte and brain acetylcholinesterase activities were reduced at 450 and 600 $\mathrm{mg} / \mathrm{kg}$ bw per day.

In a 28 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000$ and 10000 ppm (equal to $0,9.2,46.1,457.5$ and $947.8 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,9.4,47.4,461.3$ and $910.1 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $46.1 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of erythrocyte and brain acetylcholinesterase activities at 5000 ppm (equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day). Nasal toxicity, consisting of goblet cell depletion and hyperplasia of the olfactory epithelium, was noted at the highest dose.

In a 30 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,50,100,500,10000$ and 20000 ppm (equal to $0,5.1,10.4,51.9,1036$ and $2008 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,5.7,11.6,57.6,1134$ and $2193 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 10000 ppm (equal to $1036 \mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL from these two 1-month repeated-dose toxicity studies in rats was 500 ppm (equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day), with an overall LOAEL of 5000 ppm (equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 90 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000,10000$ and 20000 ppm (equal to $0,7,34,340,680$ and $1390 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,8,39,384,784$ and $1597 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 500 ppm (equal to $34 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 5000 ppm (equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a second 90 -day repeated-dose toxicity study in rats, which tested dietary malathion concentrations of $0,100,500,5000$ and 10000 ppm (equal to $0,7.2,35.0,353.6$ and $733.8 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,7.5,35.9,363.1$ and $719.0 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 100 ppm (equal to $7.2 \mathrm{mg} / \mathrm{kg}$ bw per day) for goblet cell depletion at 500 ppm (equal to $35.0 \mathrm{mg} / \mathrm{kg}$ bw per day). This is considered to be an atypical result, as the effect is likely to have arisen through non-dietary exposure.

In a 13 -week neurotoxicity study in rats, which tested dietary malathion concentrations of 0 , 50,5000 and 20000 ppm (equal to $0,4,352$ and $1486 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,4,395$ and $1575 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL was 50 ppm (equal to $4 \mathrm{mg} / \mathrm{kg}$ bw per day), based on the inhibition of erythrocyte acetylcholinesterase activity at 5000 ppm (equal to 352 $\mathrm{mg} / \mathrm{kg}$ bw per day).

The overall NOAEL for the 90 -day (neuro)toxicity studies in rats was 500 ppm (equal to 34 $\mathrm{mg} / \mathrm{kg}$ bw per day) for effects at 5000 ppm (equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day).

In a 28 -day range-finding study in dogs in which malathion was administered orally in capsules at doses of $0,125,250$ and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, inhibition of erythrocyte acetylcholinesterase occurred at 250 and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, with deaths, cholinergic signs and reduced body weight and feed consumption occurring at the highest dose.

In a 12-month repeated-dose toxicity study in dogs in which malathion was administered orally in capsules at doses of $0,62.5,125$ and $250 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $125 \mathrm{mg} / \mathrm{kg}$ bw per day for reduced body weight and haematological changes at $250 \mathrm{mg} / \mathrm{kg}$ bw per day. Inhibition of erythrocyte acetylcholinesterase activity occurred at every dose but was of marginal toxicological significance in the absence of brain acetylcholinesterase inhibition.

In a 3-week repeated-dose dermal toxicity study in rabbits, which tested malathion doses of 0 , 50,300 and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $300 \mathrm{mg} / \mathrm{kg}$ bw per day for the inhibition of brain acetylcholinesterase activity at $1000 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 21-day repeated-dose dermal toxicity study in rabbits, which tested malathion doses of 0 , $75,100,150$ and $500 \mathrm{mg} / \mathrm{kg}$ bw per day, the NOAEL was $150 \mathrm{mg} / \mathrm{kg}$ bw per day for the inhibition of brain acetylcholinesterase activity at $500 \mathrm{mg} / \mathrm{kg}$ bw per day.

In a 13 -week repeated-dose inhalational toxicity study in which rats were exposed whole body to an aerosol malathion concentration of $0,0.1,0.45$ or $2.0 \mathrm{mg} / \mathrm{L}$, a no-observed-adverse-effect concentration (NOAEC) was not determined, as laryngeal hyperplasia and degeneration and/or hyperplasia of the olfactory epithelium occurred at every concentration.

In an 18 -month pre-GLP study conducted in mice, which tested dietary malathion concentrations of 0,8000 and 16000 ppm (equivalent to 0,1200 and $2400 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), a NOAEL for chronic toxicity was not identified, because clinical signs during the second year of exposure and reduced body weight occurred at both doses. Although no treatmentrelated tumours were observed, this study was considered unreliable for assessing carcinogenicity because of the small number of concurrent control mice $(n=10)$ compared with the treated groups ( $n=50$ ).

In a second 18 -month study conducted in mice, which tested dietary malathion concentrations of $0,100,800,8000$ and 16000 ppm (equal to $0,17,143,1476$ and $2978 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,21,167,1707$ and $3448 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for chronic toxicity was 800 ppm (equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of brain acetylcholinesterase activity at 8000 ppm (equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day). Increases in liver carcinomas in males at the low dose and second highest dose were not considered treatment related because of the lack of a dose-response relationship, the lack of corroboration in females and the fact that liver carcinomas are a common age-related tumour in this strain of mouse (B6C3F1). The NOAEL for carcinogenicity was 800 ppm (equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day) for an increased incidence of liver adenomas at 8000 ppm (equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day).

In an 80 -week pre-GLP study conducted in rats, which tested dietary malathion concentrations of 0,4700 and 8150 ppm (equivalent to 0,1200 and $2400 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), it was not possible to identify a NOAEL for chronic toxicity because of the lack of reporting detail. While there was an increase in proliferative lesions of the thyroid in both sexes at both doses, these increases were not statistically significant in males and were significant in females only in a trend test and not by pairwise comparison when compared with groups of pooled controls. Overall, this study is not considered acceptable for the assessment of carcinogenicity because of the small number of rats in the concurrent control group ( 15 versus 50 in the treated groups) and the short duration of exposure.

In a subsequent 24 -month pre-GLP study conducted in rats, which tested dietary malathion concentrations of $0,100,1000$ and 5000 ppm (equivalent to $0,5,50$ and $250 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively, as calculated by a previous Meeting), the NOAEL was 100 ppm (equivalent to $5 \mathrm{mg} / \mathrm{kg}$ bw per day) for the inhibition of erythrocyte acetylcholinesterase activity at 1000 ppm (equivalent to $50 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for carcinogenicity was 5000 ppm (equivalent to $250 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

In a 24 -month chronic toxicity and carcinogenicity study in rats, which tested dietary malathion concentrations of $0,100,500,6000$ and 12000 ppm (equal to $0,7,29,359$ and $729 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,8,35,415$ and $868 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for
chronic toxicity was 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) for reduced red cell parameters, inhibition of brain acetylcholinesterase activity and the occurrence of nasal toxicity at 6000 ppm (equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day). The nasal toxicity was characterized by olfactory epithelial degeneration, hyperplasia and cyst formation, goblet cell hyperplasia, congestion, oedema and inflammation. Four nasal adenomas were observed, one in each sex at the two highest doses. In females, but not males, the incidence of liver adenomas was increased slightly at 6000 and 12000 ppm, but the incidences were within the performing laboratory's historical control range. A NOAEL of 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) was identified for carcinogenicity, based on the increase in nasal adenomas at 6000 ppm (equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day).

The Meeting concluded that there is some evidence that malathion is carcinogenic in rats and mice.

The Meeting noted that the mouse liver adenomas observed in the second 18 -month study occurred at doses exceeding the maximum tolerated dose and were not replicated in other mouse studies. The increases in liver adenomas in rats observed in the 24 -month chronic toxicity and carcinogenicity study occurred only in females and were within the performing laboratory's historical control range. Whereas the rodent liver adenomas were co-incident with liver hypertrophy, there were no findings in these or other studies to suggest a possible mode of action, such as liver enzyme induction or cytotoxicity. Malathion showed no peroxisome proliferator-activated receptor alpha or gamma activity and also showed no aryl hydrocarbon receptor activity. Overall, the Meeting considered that there was equivocal evidence to suggest a tumorigenic response in the liver, but this had a clear threshold and was likely to be secondary to the effects on the liver of prolonged exposure to very high dietary concentrations of malathion.

Based on consistent observations of nasal toxicity in dietary studies of various durations ranging from 28 days to 2 years and in a short-term inhalational toxicity study, the Meeting concluded that the formation of nasal adenomas in rats was due to a local mechanism of irritancy and cytotoxicity caused by prolonged exposure of the nasal epithelium to high concentrations of malathion absorbed via inhaled food particles or as a vapour arising from food. This produces a state of reactive hyperplasia, a causative factor in tumour formation. Scenarios of prolonged, direct and excessive exposure of human nasal tissue to malathion or malathion metabolites following ingestion of residues is unlikely, and therefore these tumours would not occur in humans following exposure to malathion in the diet.

Malathion has been extensively tested for genotoxicity using a broad range of in vitro and in vivo assays. In 1997, the Meeting evaluated the available unpublished and published genotoxicity studies and noted that the majority of studies indicated that malathion is not genotoxic, although a small number of studies indicated that it can induce chromosomal aberrations and sister chromatid exchanges in vitro. However, there was no evidence that malathion induced chromosomal aberrations in vivo. Therefore, the 1997 Meeting concluded that malathion does not induce genotoxic damage in vivo. The 2003 Meeting evaluated supplementary genotoxicity studies and found that malathion caused chromosomal aberrations in cultured human lymphocytes and gene mutations in the mouse lymphoma assay at cytotoxic concentrations, but did not cause unscheduled DNA synthesis in vivo in male rats. The 2003 Meeting reaffirmed its previous conclusion that although the results of some in vitro tests were positive, malathion was not considered to induce genotoxic damage in vivo.

In addition to the studies considered at previous meetings, the current Meeting considered a number of new published and unpublished genotoxicity studies, including studies that involved the assessment of genotoxic damage in exposed workers. Many of the published studies do not provide adequate experimental detail, do not specify the purity of the malathion tested or were conducted on commercial formulations, or used in vivo test systems or exposure routes less relevant to the risk assessment of dietary residues of pesticides. The following discussion is limited to studies that evaluated technical malathion or malathion at purities above $90 \%$ and provided adequate experimental and data analysis details to allow interpretation of the findings.

Using standard genotoxicity test systems, malathion was not mutagenic in assays using prokaryotes or lower eukaryotes when tested with or without metabolic activation. In contrast, in in vitro assays using either human or non-human cells, malathion was generally positive for the induction of (1) chromosomal damage, as measured by increased frequencies of chromosomal aberrations or micronuclei; (2) mutations; and (3) DNA damage, as measured by increases in DNA migration in the alkaline comet assay and increased frequencies of sister chromatid exchanges. Negative findings were reported for the induction of micronuclei in Molt-4 T-lymphocytes, unscheduled DNA synthesis in WI-38 cells and primary rat liver hepatocytes, and mutations in a mouse lymphoma assay (reported to be equivocal without metabolic activation and negative with metabolic activation).

Using in vivo non-mammalian systems, malathion was active for micronucleus induction in a bird model and for induction of reciprocal translocations and sex-linked recessive lethals in one Drosophila melanogaster study, but not for sex-linked recessive lethals, sex chromosome loss or wing spot mutations in another study.

Based on the criteria mentioned in section 2.1, very few of the 34 in vivo mammalian study/end-point combinations were considered adequate for this review. In reports submitted by the sponsor, malathion was negative in a rat liver unscheduled DNA synthesis study when administered by gavage, in a rat bone marrow chromosomal aberration study when administered by gavage and in a mouse bone marrow erythrocyte micronucleus assay when administered intraperitoneally. However, the unscheduled DNA synthesis assay is insensitive for detecting genotoxic compounds; the micronucleus assay, as conducted, suffers from concerns about scoring criteria; and the chromosomal aberration test appears to be significantly underpowered, based on the frequency of chromosomal aberrations detected among control and treated animals. A negative mouse dominant lethal test was also reported when malathion was administered in feed for 7 weeks, and a negative mouse bone marrow chromosomal aberration study was reported in intraperitoneally treated mice. In contrast, malathion-induced micronuclei and chromosomal aberrations were reported in bone marrow immature erythrocytes and proliferating cells, respectively. A positive alkaline comet assay using blood leukocytes sampled from rats treated intraperitoneally once a day for 5 days was reported.

The Meeting evaluated a number of human studies that examined genotoxicity end-points. Patients treated for acute intoxication with a malathion-based product exhibited increased levels of chromosomal damage in lymphocytes. The frequency of micronuclei and glycophorin A mutations in erythrocytes or micronuclei in lymphocytes was not increased in workers exposed selectively to malathion. However, DNA damage and chromosomal aberrations have been reported in workers exposed to a mixture of pesticides, including malathion. These studies are of limited value for examining the specific effect of malathion on genotoxicity end-points in humans.

The Meeting noted that malathion has been reported to have genotoxic activity in multiple assay systems at multiple genetic end-points. In several studies where evaluated, reactive oxygen species appear to have been responsible for the increased damage, as demonstrated by the detection of malathion-induced 8 -hydroxy-2'-deoxyguanosine and increased malondialdehyde concentrations in isolated human peripheral blood mononuclear cells treated in vitro, an effect attenuated by cotreatment with $N$-acetylcysteine or curcumin; by increased intracellular levels of reactive oxygen species and reduced levels of catalase, superoxide dismutase and glutathione in rat PC12 cells treated in vitro, an effect ameliorated by co-treatment with vitamin E; and by the detection of oxidative damage using the comet assay in isolated rat lymphocytes treated in vitro with malathion. Supportive of this hypothesis, malathion appears to selectively induce markers of oxidative stress in Tox21/ToxCast high-throughput screening assays. The Meeting concluded that the observed genotoxic effects occur secondary to the formation of reactive oxygen species, which will exhibit a threshold.

The Meeting concluded that malathion is unlikely to be genotoxic at anticipated dietary exposures.

In the multigeneration and developmental toxicity studies, cholinesterase activity was not measured.

In a two-generation reproductive toxicity study conducted in rats, which tested dietary malathion concentrations of $0,550,1700,5000$ and 7500 ppm (equal to $0,43,130,393$ and 595 $\mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,50,152,438$ and $655 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for both reproductive toxicity and parental toxicity was 7500 ppm (equal to $595 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested. The NOAEL for offspring toxicity was 1700 ppm (equal to 130 $\mathrm{mg} / \mathrm{kg}$ bw per day) for reduced pup weights at 5000 ppm (equal to $393 \mathrm{mg} / \mathrm{kg}$ bw per day).

Two published studies reported potential testicular toxicity in rats exposed to malathion orally, but these studies had a number of methodological limitations that reduced their utility. Further, the reported observations are not corroborated by the preceding GLP-compliant multigenerational rat study in which no effects on the testes were observed.

A variety of in vivo and in vitro assays in mammalian and non-mammalian models indicated that malathion is unlikely to affect the endocrine system.

In a pilot developmental toxicity study in rats, which tested gavage malathion doses of 0,300 , 600,800 and $1000 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 15 of gestation, no embryo or fetal toxicity occurred, whereas maternal toxicity occurred at and above $600 \mathrm{mg} / \mathrm{kg}$ bw per day. In the main developmental toxicity study in rats, which tested gavage doses of $0,200,400$ and $800 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 15 of gestation, the NOAEL for maternal toxicity was $400 \mathrm{mg} / \mathrm{kg}$ bw per day for clinical signs and reduced body weight gain and feed consumption at $800 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo and fetal toxicity was $800 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

In a range-finding developmental toxicity study in rabbits, which tested gavage malathion doses of $0,25,50,100,200$ and $400 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 18 of gestation, no embryo or fetal toxicity occurred, whereas maternal toxicity occurred at 200 and $400 \mathrm{mg} / \mathrm{kg}$ bw per day. In the main study, which tested malathion doses of $0,25,50$ and $100 \mathrm{mg} / \mathrm{kg}$ bw per day from days 6 to 18 of gestation, the NOAEL for maternal toxicity was $25 \mathrm{mg} / \mathrm{kg}$ bw per day for a marginal effect on body weight gain at $50 \mathrm{mg} / \mathrm{kg}$ bw per day. The NOAEL for embryo and fetal toxicity was $100 \mathrm{mg} / \mathrm{kg}$ bw per day, the highest dose tested.

The Meeting concluded that malathion is not teratogenic.
In a study conducted in hens, there was no evidence that malathion caused delayed peripheral neuropathy.

In an acute neurotoxicity study in rats, which tested gavage malathion doses of $0,500,1000$ and $2000 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, the NOAEL was $1000 \mathrm{mg} / \mathrm{kg}$ bw for reduced erythrocyte acetylcholinesterase activity in females and reduced ambulatory activity in males at $2000 \mathrm{mg} / \mathrm{kg}$ bw.

A 13-week neurotoxicity study in rats is described above together with the other 13-week toxicity studies in rats, and an overall NOAEL is identified for these studies.

In a developmental neurotoxicity study in rats, which tested gavage malathion doses of $0,5,50$ and $150 \mathrm{mg} / \mathrm{kg}$ bw per day from day 6 of gestation to day 10 of lactation, the NOAEL for both maternal toxicity and offspring toxicity was $50 \mathrm{mg} / \mathrm{kg}$ bw per day for clinical signs at $150 \mathrm{mg} / \mathrm{kg}$ bw per day.

Administration of malathion from day 6 of gestation to day 21 of lactation had no effect on the thickness of the corpus callosum in rat pups at doses up to $150 \mathrm{mg} / \mathrm{kg}$ bw per day.

The Meeting concluded that malathion is neurotoxic.
Studies in rats have examined the time to peak effect and compared the effects of malathion and malaoxon on the inhibition of acetylcholinesterase activity. The time to peak effect in juvenile rats following dosing with malathion ranged from 30 to 90 minutes for the inhibition of erythrocyte acetylcholinesterase activity and from 60 to 90 minutes for the inhibition of brain
acetylcholinesterase activity. Malaoxon was a more potent inhibitor of acetylcholinesterase activity compared with malathion. Comparison of benchmark doses (BMDs) following acute oral dosing indicated that the toxicity adjustment factor (TAF) for malaoxon was 21.5 in males and 17.4 in females for the inhibition of erythrocyte acetylcholinesterase activity and 14.8 in males and 11.0 in females for the inhibition of brain acetylcholinesterase activity. Comparison of BMDs for the inhibition of erythrocyte acetylcholinesterase activity from chronic toxicity studies indicated that TAFs for malaoxon ranged from 37 to 38 in males and from 65 to 69 in females.

In a 6 -week immunotoxicity study in female rats, which tested dietary malathion concentrations of $0,50,100,700$ and 7000 ppm (equal to $0,8.9,17.6,126.8$ and $1215.8 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), the NOAEL for immunotoxicity was 7000 ppm (equal to $1215.8 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested.

The Meeting concluded that malathion is not immunotoxic.
An extensive literature search did not identify any potential adverse effects on intestinal microbiota or any evidence that intestinal microbiota can metabolize malathion.

## Toxicological data on metabolites, degradates and/or impurities

Current FAO specifications for malathion prescribe maximum limits for isomalathion (CAS No. 3344-12-5), malaoxon (CAS No. 152-20-05), O,O,S-trimethyl phosphorothioate (CAS No. 2953-299 ) and $O, S, S$-trimethyl phosphorodithoate (CAS No. 152-18-1).

Toxicity tests were conducted on malaoxon, isomalathion, desmethyl malathion, desmethyl malathion monocarboxylic acid, MMCA, MDCA and desmethyl malaoxon dicarboxylic acid.

## Malaoxon

The oral $\mathrm{LD}_{50}$ in rats for malaoxon was $50 \mathrm{mg} / \mathrm{kg}$ bw.
In a 14 -day range-finding study in rats, which tested malaoxon at dietary concentrations of 0 , $10,25,100,2500$ and 3500 ppm (equal to $0,1.1,3.0,12.1,293$ and $387 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1.1,3.1,12.5,281.6$ and $294.7 \mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), inhibition of erythrocyte acetylcholinesterase activity occurred at and above 100 ppm (equal to $12.1 \mathrm{mg} / \mathrm{kg}$ bw per day). At the two highest doses, inhibition of brain acetylcholinesterase activity and reduced body weight gain and feed consumption occurred.

In a 103 -week carcinogenicity study conducted in mice, which tested dietary malaoxon concentrations of 0,500 and 1000 ppm (estimated by a previous Meeting to be equal to 0,75 and 150 $\mathrm{mg} / \mathrm{kg}$ bw per day, respectively), survival and body weight were reduced at the highest dose. There were no treatment-related neoplastic or non-neoplastic lesions. In a parallel study conducted in rats, which tested the same dietary concentrations of malathion (equal to 0,25 and $50 \mathrm{mg} / \mathrm{kg}$ bw per day, respectively), the combined incidence of C-cell adenomas and carcinomas of the thyroid in females was increased, although this was comparable to historical control values. The incidence of gastric ulcers, commonly observed in the forestomach, was increased in treated rats.

In a 24 -month toxicity study in rats, which tested malaoxon at dietary concentrations of 0,20 , 1000 and 2000 ppm (equal to $0,1,57$ and $110 \mathrm{mg} / \mathrm{kg}$ bw per day for males and $0,1,68$ and 140 $\mathrm{mg} / \mathrm{kg}$ bw per day for females, respectively), the NOAEL for chronic toxicity was 20 ppm (equal to 1 $\mathrm{mg} / \mathrm{kg}$ bw per day), based on mortality and the inhibition of brain acetylcholinesterase activity at 1000 ppm (equal to $57 \mathrm{mg} / \mathrm{kg}$ bw per day). The NOAEL for carcinogenicity was 2000 ppm (equal to $110 \mathrm{mg} / \mathrm{kg}$ bw per day), the highest dose tested. Similar to studies conducted on malathion, inflammatory changes in the nasal mucosa occurred at 1000 and 2000 ppm ; these changes were likely attributable to inhaled food particles containing malaoxon, resulting in tissue injury and inflammation of the nasal cavity, with secondary effects on the lungs and middle ear.

The Meeting concluded that malaoxon is not carcinogenic in mice or rats.
Malaoxon was negative for mutagenicity in bacterial assays and in lower eukaryotes, both with and without metabolic activation. Malaoxon was reported to be active for induction of sister chromatid exchanges but not chromosomal aberrations in Chinese hamster ovary cells, with or without metabolic activation. An increase in sister chromatid exchanges when tested in the absence of metabolic activation only was also reported; it was also reported that malaoxon was more potent than malathion in this assay. Malaoxon was also reported to induce DNA damage as measured by the comet assay in rat adrenal gland PC12 cells when tested in the absence of metabolic activation only and was mutagenic in mouse lymphoma (L5178Y) cells in the absence but not the presence of metabolic activation. In this study, there seemed to be a preference for the induction of small colonies, generally considered to be indicative of chromosomal damage rather than gene mutations.

Malaoxon induced DNA damage in isolated lymphocytes in the absence of metabolic activation, as measured by the alkaline comet assay; studies with metabolic activation were not conducted. Further, a follow-up study concluded that the malaoxon-mediated damage was likely induced by reactive oxygen species. Also, malaoxon is more potent than malathion in inducing intracellular levels of reactive oxygen species and reducing levels of catalase, superoxide dismutase and glutathione in rat PC12 cells treated in vitro, an effect ameliorated by co-treatment with vitamin E. Also, similar to malathion, malaoxon appears to selectively induce markers of oxidative stress in Tox21/ToxCast high-throughput screening assays. When provided in food, malaoxon induced an increase in reciprocal translocations and sex-linked recessive lethals in D. melanogaster, but not for sex-linked recessive lethals when administered by injection. Malaoxon was reported negative for the induction of chromosomal aberrations and sister chromatid exchanges in the bone marrow cells of male mice following a single intraperitoneal injection.

The Meeting concluded that the observed genotoxic effects occur secondary to the formation of reactive oxygen species, which will exhibit a threshold.

The Meeting concluded that malaoxon is unlikely to be genotoxic at anticipated dietary exposures.

## Other metabolites

The oral $\mathrm{LD}_{50}$ in rats was greater than $2000 \mathrm{mg} / \mathrm{kg}$ bw for desmethyl malathion, desmethyl malathion monocarboxylic acid, MMCA, MDCA and desmethyl malaoxon dicarboxylic acid. The oral $\mathrm{LD}_{50}$ in rats for desmethyl malaoxon dicarboxylic acid, trisodium salt, was greater than $2000 \mathrm{mg} / \mathrm{kg}$ bw.

There are a limited number of genotoxicity studies on other metabolites of malathion. MDCA, MMCA, desmethyl malathion monocarboxylic acid, potassium salt, and desmethyl malaoxon dicarboxylic acid, trisodium salt, as well as isomalathion, $O, O, O$-trimethyl phosphorothioate, $O, O, S$ trimethyl phosphorothioate and $O, S, S$-trimethyl phosphorodithioate, were reported negative for bacterial mutagenicity, with and without metabolic activation. Isomalathion induced DNA damage in isolated lymphocytes in the absence of metabolic activation, as measured by the alkaline comet assay; studies with metabolic activation were not conducted. Isomalathion was also reported to induce micronuclei in the human liver-derived HepaRG cell line.

Using quantitative structure-activity relationships, the storage impurity, 2-mercaptosuccinic acid diethyl ester, was determined to have no greater toxicity than malathion.

The potential of malathion metabolites to inhibit acetylcholinesterase activity has been studied in rats. Comparisons of erythrocyte acetylcholinesterase activities indicated that desmethyl malathion, MMCA and MDCA are at least 2.75-, 1.9- and 4.6-fold less potent than malathion.

Based on a comparison of the inhibitions of acetylcholinesterase activities over acute and chronic exposure durations and a comparison of BMDs (see above), the Meeting concluded that malaoxon is approximately 30 -fold more potent than malathion.

## Human data

As in laboratory animals, the inhibition of acetylcholinesterase activity is the most sensitive adverse effect in humans exposed to malathion, mediated through the metabolite malaoxon, which is a more potent inhibitor of acetylcholinesterase activity compared with malathion. A comparative in vitro study indicated that malaoxon was a slightly less potent inhibitor ( $<2$-fold) of human compared with rat acetylcholinesterase activity.

In a study conducted in male and female volunteers, which tested single oral doses of malathion at $0,0.5,1.5,5,10$ and $15 \mathrm{mg} / \mathrm{kg}$ bw, the NOAEL was $15 \mathrm{mg} / \mathrm{kg}$ bw, the highest dose tested, based on the absence of any adverse effects, including the inhibition of erythrocyte acetylcholinesterase activity. In a subsequent study conducted in male and female volunteers, which tested single oral doses of malathion of $0,0.5,1.5,5.0,10.0$ and $15.0 \mathrm{mg} / \mathrm{kg}$ bw, there were no treatment-related adverse events or effects on erythrocyte acetylcholinesterase activity.

In a published study, application of malathion to the forearm of human volunteers increased blood flow, mediated via the inhibition of acetylcholinesterase activity.

In a published non-blinded study, slight inhibition of erythrocyte acetylcholinesterase activity occurred in children following two applications of a $1 \%$ malathion shampoo used to treat head lice.

In a 1994 summary report, there were no poisoning incidents and no inhibition of plasma cholinesterase activity in workers involved in the manufacture of malathion over a 20 -year period. In a subsequent (1999) summary report, biological monitoring of workers employed at dimethoate and malathion manufacturing plants from 1994 to 1999 detected no reduction in plasma cholinesterase activity.

Several epidemiological studies on cancer outcomes in relation to occupational exposure to malathion were available. The evaluation of these studies focused on the occurrence of NHL and prostate cancer, as outlined in section 2.2. One meta-analysis was available, as well as one prospective cohort study, the AHS, with a large sample size and detailed exposure assessment. Cohort studies are considered a powerful design, as recall bias is avoided. All other studies were case-control studies, usually retrospective, which are more prone to recall and selection biases.

The AHS found no evidence of a positive association of NHL with malathion exposure or of an exposure-response relationship. In contrast, various case-control studies reported excess risks of NHL associated with use of malathion. In a large pooled case-control study, the unadjusted estimates showed a significant increased risk of NHL $(R R=1.6 ; 95 \% C I=1.2-2.2)$ associated with ever versus never use of malathion. However, these were attenuated and/or no longer significant when proxy respondents were excluded and analyses were mutually adjusted for other pesticides. Significant elevated risks of NHL were reported from the cross-Canada case-control study of pesticides and health for ever versus never use of malathion ( $\mathrm{OR}=1.96 ; 95 \% \mathrm{CI}=1.42-2.70$ ) and when examining annual days of use, although there was no clear exposure-response relationship across exposure categories. Non-significant increased risks of NHL were reported by two other case-control studies, one of which had limited statistical power based on only five exposed cases. The meta-analysis, which did not include the AHS, found a significant $80 \%$ excess risk ratio for ever versus never use of malathion.

Overall, there is some very weak evidence of a positive association between malathion exposure and NHL from the case-control studies and the overall meta-analysis. However, it is notable that the AHS, which is the only cohort study and is large and of high quality, found no evidence of an association at any exposure level.

There was no evidence of an association with all prostate cancers and malathion exposure in the AHS. However, a significant excess risk of aggressive prostate cancer $(\mathrm{RR}=1.43 ; 95 \% \mathrm{CI}=$ 1.08-1.88) in the highest exposure category (highest quintile of intensity-weighted lifetime days of malathion exposure), along with a significant exposure-response relationship ( $P$ for trend $=0.04$ ), was observed. A significant elevated risk of all prostate cancer was observed in a case-control study
for ever use $(\mathrm{OR}=1.34 ; 95 \% \mathrm{CI}=1.01-1.78)$ and for highest lifetime cumulative exposure versus those unexposed $(\mathrm{OR}=1.49 ; 95 \% \mathrm{CI}=1.02-2.18)$. A significant trend across exposure categories $(P=0.03)$ was also reported. However, interpretation of results from this study is limited by potential for exposure misclassification in the job-exposure matrix used for exposure assessment and by the potential for residual confounding from lack of adjustment for other pesticide exposures. There was no evidence of an association between prostate cancer and malathion exposure in the United Farm Workers of America study, which was limited by the use of ecological rather than individual-level exposure assessment.

Overall, the evidence is suggestive of a positive association between malathion exposure and risk of aggressive prostate cancer; however, the evidence base is limited to the one large AHS cohort study.

Based on a consideration of the results of animal bioassays, genotoxicity assays and epidemiological data from occupational exposures, the Meeting concluded that malathion and its metabolites are unlikely to pose a carcinogenic risk to humans from exposure via the diet.

The Meeting concluded that the existing database on malathion was adequate to characterize the potential hazards to the general population, including fetuses, infants and children.

## Toxicological evaluation

The current Meeting reaffirmed the ADI of $0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$, based on the NOAEL of 500 ppm (equal to $29 \mathrm{mg} / \mathrm{kg}$ bw per day) in the 2-year study of toxicity and carcinogenicity in rats for the inhibition of brain acetylcholinesterase and using a 100-fold safety factor, established by the 1997 Meeting. The margins of exposure between this ADI and the doses causing liver adenomas in mice and nasal adenomas in rats are 5000 -fold and 1200 -fold, respectively.

The current Meeting reaffirmed the ARfD of $2 \mathrm{mg} / \mathrm{kg}$ bw, based on the NOAEL of $15 \mathrm{mg} / \mathrm{kg}$ bw for the inhibition of erythrocyte acetylcholinesterase activity in a study conducted in male and female volunteers with the application of a 10 -fold safety factor, established by the 2003 Meeting. This ARfD is supported by the NOAEL of $15 \mathrm{mg} / \mathrm{kg}$ bw in a second study conducted in male and female volunteers. The ARfD is considered to be a conservative value, because human acetylcholinesterase is slightly less sensitive ( $<2$-fold) than rat acetylcholinesterase to malaoxon.

The Meeting concluded that the metabolite malaoxon is approximately 30 -fold more toxic than malathion. On this basis, a 30 -fold potency factor should be applied to the residue levels for use in both the acute and chronic dietary exposure estimates for malaoxon, and these should be added to the dietary exposures for malathion and compared with the ARfD and ADI for malathion, respectively.

Both the ADI and ARfD are established for the sum of malathion and malaoxon (corrected for its potency), expressed as parent malathion. The other metabolites of malathion considered by the present Meeting are less potent than the parent compound and therefore would be covered by the ADI and ARfD for malathion. The impurity isomalathion may need to be taken into consideration in the risk assessment depending on its concentration in food commodities.

A toxicological monograph was prepared.

## Levels relevant to risk assessment of malathion

| Species | Study | Effect | NOAEL | LOAEL |
| :--- | :--- | :--- | :--- | :--- |
| Mouse | Two-year study of toxicity | Toxicity | 800 ppm, equal to | 8000 ppm, equal to |
|  | and carcinogenicity $^{\text {a }}$ |  | $143 \mathrm{mg} / \mathrm{kg}$ bw per <br>  |  |
|  |  |  | $1476 \mathrm{mg} / \mathrm{kg}$ bw per |  |


| Species | Study | Effect | NOAEL | LOAEL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Carcinogenicity | 800 ppm , equal to $143 \mathrm{mg} / \mathrm{kg}$ bw per day | 8000 ppm , equal to $1476 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rat | Acute neurotoxicity study ${ }^{\text {b }}$ | Toxicity | $1000 \mathrm{mg} / \mathrm{kg}$ bw per day | $2000 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | One-month studies of toxicity ${ }^{\mathrm{a}, \mathrm{c}}$ | Toxicity | 500 ppm , equal to $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $457.5 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Thirteen-week studies of toxicity and neurotoxicity ${ }^{\mathrm{a}, \mathrm{c}}$ | Toxicity | 500 ppm , equal to $34 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $340 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-year study of toxicity and carcinogenicity ${ }^{\text {a }}$ | Toxicity | 500 ppm , equal to 29 $\mathrm{mg} / \mathrm{kg}$ bw per day | 6000 ppm , equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Carcinogenicity | 500 ppm , equal to 29 $\mathrm{mg} / \mathrm{kg}$ bw per day | 6000 ppm , equal to $359 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Two-generation study of reproductive toxicity ${ }^{\text {a,e }}$ | Reproductive toxicity | $\begin{aligned} & 7500 \mathrm{ppm} \text {, equal to } \\ & 595 \mathrm{mg} / \mathrm{kg} \text { bw per } \\ & \text { day }^{\mathrm{d}} \end{aligned}$ | - |
|  |  | Parental toxicity | $\begin{aligned} & 7500 \mathrm{ppm} \text {, equal to } \\ & 595 \mathrm{mg} / \mathrm{kg} \text { bw per } \\ & \text { day }^{\mathrm{d}} \end{aligned}$ | - |
|  |  | Offspring toxicity | 1700 ppm , equal to $130 \mathrm{mg} / \mathrm{kg}$ bw per day | 5000 ppm , equal to $393 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  | Developmental toxicity study ${ }^{\text {b,e }}$ | Maternal toxicity | $400 \mathrm{mg} / \mathrm{kg}$ bw per day | $800 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $800 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {d }}$ | - |
|  | Developmental neurotoxicity study ${ }^{\text {b,e }}$ | Maternal toxicity | $50 \mathrm{mg} / \mathrm{kg}$ bw per day | $150 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Offspring toxicity | $50 \mathrm{mg} / \mathrm{kg}$ bw per day | $150 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Rabbit | Developmental toxicity study ${ }^{\text {be }}$ | Maternal toxicity | $25 \mathrm{mg} / \mathrm{kg}$ bw per day | $50 \mathrm{mg} / \mathrm{kg}$ bw per day |
|  |  | Embryo and fetal toxicity | $100 \mathrm{mg} / \mathrm{kg}$ bw per day ${ }^{\text {d }}$ | - |
| Dog | One-year study of toxicity ${ }^{\text {f }}$ | Toxicity | $125 \mathrm{mg} / \mathrm{kg}$ bw per day | $250 \mathrm{mg} / \mathrm{kg}$ bw per day |
| Human | Acute volunteer studies ${ }^{\text {c,f }}$ | Cholinesterase inhibition | $15 \mathrm{mg} / \mathrm{kg} \mathrm{bw}{ }^{\text {d }}$ | - |

${ }^{\text {a }}$ Dietary administration.
${ }^{\text {b }}$ Gavage administration.
c Two or more studies combined.
${ }^{\mathrm{d}}$ Highest dose tested.
${ }^{\text {e }}$ Acetylcholinesterase activity not measured.
${ }^{\mathrm{f}}$ Capsule administration.

## Estimate of acceptable daily intake (ADI)

$0-0.3 \mathrm{mg} / \mathrm{kg}$ bw (for sum of malathion and malaoxon, adjusted for its potency, and expressed as malathion)

## Estimate of acute reference dose (ARfD)

$2 \mathrm{mg} / \mathrm{kg}$ bw (for sum of malathion and malaoxon, adjusted for its potency, and expressed as malathion)

## Information that would be useful for the continued evaluation of the compound

Results from epidemiological, occupational health and other such observational studies of human exposure

Results from in vivo genotoxicity studies investigating oral dosing, because malathion genotoxicity data are highly variable and inconsistent and there is a lack of robust in vivo rodent studies using the oral route of exposure

Critical end-points for setting guidance values for exposure to malathion

| Rate and extent of oral absorption | Rapid; > 77\% |
| :---: | :---: |
| Dermal absorption | Estimates vary (1.44-20.7\% in human skin) |
| Distribution | Rapid tissue distribution |
| Potential for accumulation | No potential for accumulation |
| Rate and extent of excretion | Rapid and complete |
| Metabolism in animals | Extensive; oxidation, hydrolysis, dealkylation and demethylation reactions |
| Toxicologically significant compounds in animals and plants | Malathion, malaoxon, desmethyl malathion, desmethyl malaoxon, MMCA, MDCA, isomalathion |
| Acute toxicity |  |
| Rat, $\mathrm{LD}_{50}$, oral | $>1539$ to <8227 mg/kg bw |
| Rat, $\mathrm{LD}_{50}$, dermal | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| Rat, $\mathrm{LC}_{50}$, inhalation | $>5.2 \mathrm{mg} / \mathrm{L}$ |
| Rabbit, dermal irritation | Slightly irritating |
| Rabbit, ocular irritation | Slightly irritating |
| Guinea-pig, dermal sensitization | Not sensitizing (Buehler assay) |
|  | Sensitizing (maximization assay) |
| Mouse, dermal sensitization | Not sensitizing (local lymph node assay) |
| Short-term studies of toxicity |  |
| Target/critical effect | Acetylcholinesterase inhibition |
| Lowest relevant oral NOAEL | $51.9 \mathrm{mg} / \mathrm{kg}$ bw per day (28 days; rat) |
| Lowest relevant dermal NOAEL | $150 \mathrm{mg} / \mathrm{kg}$ bw per day (21 days; rabbit) |
| Lowest relevant inhalation NOAEC | $<0.1 \mathrm{mg} / \mathrm{L}$ (13 weeks; rat) |

Long-term studies of toxicity and carcinogenicity
Target/critical effect Acetylcholinesterase inhibition


## Toxicological studies on MDCA

| Rat, $\mathrm{LD}_{50}$, oral | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Genotoxicity | Not mutagenic in prokaryotic assays |

Toxicological studies on desmethyl malaoxon
dicarboxylic acid

| Rat, $\mathrm{LD}_{50}$, oral | $>2000 \mathrm{mg} / \mathrm{kg}$ bw |
| :--- | :--- |
| Genotoxicity | Not mutagenic in prokaryotic assays |
| Human data | Acetylcholinesterase inhibition: |
|  | Acute NOAEL: $15 \mathrm{mg} / \mathrm{kg}$ bw, highest dose tested |
|  | No adverse effects in manufacturing personnel |
|  |  |

[^18]
## Summary

|  | Value | Studies | Safety factor |
| :--- | :--- | :--- | :--- |
| ADI | $0-0.3 \mathrm{mg} / \mathrm{kg}$ bw | Two-year chronic toxicity and carcinogenicity study <br> (rat) | 100 |
| ARfD | $2 \mathrm{mg} / \mathrm{kg}$ bw | Single-dose studies (humans) | 10 |

## DIETARY RISK ASSESSMENT

The current residue definition for the estimation of dietary exposure is malathion. The Meeting identified that malaoxon is approximately 30 times more potent than malathion based on the endpoint (acetylcholinesterase inhibition) on which the ADI and ARfD have been established. Malaoxon is generally present in food at concentrations that are approximately $3 \%$ of the malathion concentration. If malaoxon were included in the residue definition for dietary risk assessment, the exposures calculated below for comparison with the health-based guidance values would be approximately double.

## Long-term dietary exposure

The ADI for malathion is $0-0.3 \mathrm{mg} / \mathrm{kg}$ bw. The IEDIs for malathion were estimated for the 17 GEMS/Food cluster diets using the STMR or STMR-P values estimated by JMPR. The results are shown in Annex 3. The IEDI ranged from $0.1 \%$ to $0.5 \%$ of the maximum ADI. The Meeting concluded that the long-term dietary exposure to residues of malathion from uses that have been considered by JMPR is unlikely to present a public health concern.

## Short-term dietary exposure

The ARfD for malathion is $2 \mathrm{mg} / \mathrm{kg}$ bw. The IESTI for malathion was calculated for the plant commodities for which STMR and HR levels were estimated by the 1999, 2004 and 2008 JMPRs and for which consumption data were available. The results are shown in Annex 4. The calculated IESTIs were $0-5 \%$ of the ARfD for the general population and $0-9 \%$ of the ARfD for children. The Meeting concluded that the short-term dietary exposure to malathion residues from uses considered by the Meeting was unlikely to present a public health concern.

## 6. RECOMMENDATIONS

The Meeting recommended that a guidance document be developed for the evaluation of genotoxicity studies, taking the experience gained from this meeting into account.

## ANNEX 1: ACCEPTABLE DAILY INTAKES AND ACUTE REFERENCE DOSES RECORDED BY THE MAY 2016 MEETING

| Pesticide (Codex reference <br> number) | Acceptable daily intake (ADI) <br> $(\mathbf{m g} / \mathbf{k g}$ bw) | Acute reference dose (ARfD) <br> $(\mathbf{m g} / \mathbf{k g}$ bw) |
| :--- | :---: | :---: |
| Diazinon (22) | $0-0.003$ | 0.03 |
| Glyphosate (158) | $0-1^{\mathrm{a}}$ | Unnecessary |
| Malathion (49) | $0-0.3^{\mathrm{b}}$ | $2^{\mathrm{b}}$ |
| a |  |  |

[^19]
## ANNEX 2: INDEX OF REPORTS AND EVALUATIONS OF PESTICIDES BY THE JMPR

Numbers in parentheses after the names of pesticides are Codex classification numbers. The abbreviations used are:

T, evaluation of toxicology
R , evaluation of residue and analytical aspects
E, evaluation of effects on the environment

| Abamectin (177) | $\begin{aligned} & 1992 \text { (T,R), } 1994 \text { (T,R), } 1995 \text { (T), } 1997 \text { (T,R), } \\ & 2000 \text { (R), } 2015 \text { (R) } \end{aligned}$ |
| :---: | :---: |
| Acephate (095) | 1976 (T,R), 1979 (R), 1981 (R), 1982 (T), 1984 (T,R), 1987 (T), 1988 (T), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1994 (R), 1996 (R), 2002 (T), 2003 (R), 2004 (corr. to 2003 report), 2005 (T), 2006 (R), 2011 (R) |
| Acetamiprid (246) | 2011 (T,R), 2012 (R), 2015 (R) |
| Acetochlor (280) | 2015 (T,R) |
| Acrylonitrile | 1965 (T,R) |
| Aldicarb (117) | 1979 (T,R), 1982 (T,R), 1985 (R), 1988 (R), 1990 (R), 1991 (corr. to 1990 evaluation), 1992 (T), 1993 (R), 1994 (R), 1996 (R), 2001 (R), 2002 (R), 2006 (R) |
| Aldrin (001) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (R), } 1974 \text { (R), } 1975 \text { (R), } \\ & 1977 \text { (T), } 1990 \text { (R), } 1992 \text { (R) } \end{aligned}$ |
| Allethrin | 1965 (T,R) |
| Ametoctradin (253) | 2012 (T,R) |
| Aminocarb (134) | 1978 (T,R), 1979 (T,R) |
| Aminocyclopyrachlor (272) | 2014 (T,R) |
| Aminomethylphosphonic acid (AMPA, 198) | 1997 (T,R) |
| Aminopyralid (220) | 2006 (T,R), 2007 (T,R) |
| Amitraz (122) | 1980 (T,R), 1983 (R), 1984 (T,R), 1985 (R), 1986 (R), 1989 (R), 1990 (T,R), 1991 (R \& corr. to 1990 R evaluation), 1998 (T) |
| Amitrole (079) | 1974 (T,R), 1977 (T), 1993 (T,R), 1997 (T), 1998 (R) |
| Anilazine (163) | 1989 (T,R), 1992 (R) |
| Atrazine | 2007 (T) |
| Azinphos-ethyl (068) | 1973 (T,R), 1983 (R) |
| Azinphos-methyl (002) | $\begin{aligned} & 1965(\mathrm{~T}), 1968 \text { (T,R), } 1972(\mathrm{R}), 1973(\mathrm{~T}), 1974(\mathrm{R}), \\ & 1991 \text { (T,R), } 1992 \text { (corr. to } 1991 \text { report), } 1993(\mathrm{R}) \text {, } \\ & 1995(\mathrm{R}), 2007(\mathrm{~T}) \end{aligned}$ |

Azocyclotin (129)

Azoxystrobin (229)
Benalaxyl (155)

Bendiocarb (137)
Benomyl (069)

Bentazone (172)

Benzovinfiflupyr (261)
BHC (technical-grade)
Bifenazate (219)
Bifenthrin (178)

Binapacryl (003)

Bioresmethrin (093)
Biphenyl
Bitertanol (144)

Bixafen (262)
Boscalid (221)
Bromide ion (047)

Bromomethane (052)

Bromophos (004)

Bromophos-ethyl (005)
Bromopropylate (070)
Butocarboxim (139)
Buprofezin (173)
sec-Butylamine (089)

Cadusafos (174)
Campheclor (071)

1979 (R), 1981 (T), 1982 (R), 1983 (R), 1985 (R), 1989 (T,R), 1991 (R), 1994 (T), 2005 (T,R)

2008 (T,R), 2011 (R), 2012 (R), 2013 (R)
1986 (R), 1987 (T), 1988 (R), 1992 (R), 1993 (R), 2005 (T), 2009 (R)

1982 (T,R), 1984 (T,R), 1989 (R), 1990 (R)
1973 (T,R), 1975 (T,R), 1978 (T,R), 1983 (T,R), 1988 (R), 1990 (R), 1994 (R), 1995 (T,E), 1998 (R)

1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1995 (R), 1998 (T,R), 1999 (corr. to 1998 report), 2004 (T), 2012 (T), 2013 (R)

2013 (T), 2014 (R)
1965 (T), 1968 (T,R), 1973 (T,R) (see also Lindane)
2006 (T,R), 2008 (R), 2010 (R)
1992 (T,R), 1995 (R), 1996 (R), 1997 (R), 2009 (T), 2010 (R), 2015 (R)

1969 (T,R), 1974 (R), 1982 (T), 1984 (R), 1985 (T,R)

1975 (R), 1976 (T,R), 1991 (T,R)
See Diphenyl
1983 (T), 1984 (R), 1986 (R), 1987 (T), 1988 (R), 1989 (R), 1991 (R), 1998 (T), 1999 (R), 2002 (R)

2013 (T,R)
2006 (T,R), 2008 (R), 2010 (R)
1968 (R), 1969 (T,R), 1971 (R), 1979 (R), 1981 (R), 1983 (R), 1988 (T,R), 1989 (R), 1992 (R)

1965 (T,R), 1966 (T,R), 1967 (R), 1968 (T,R), 1971 (R), 1979 (R), 1985 (R), 1992 (R)

1972 (T,R), 1975 (R), 1977 (T,R), 1982 (R), 1984 (R), 1985 (R)

1972 (T,R), 1975 (T,R), 1977 (R)
1973 (T,R), 1993 (T,R)
1983 (R), 1984 (T), 1985 (T), 1986 (R)
1991 (T,R), 1995 (R), 1996 (corr. to 1995 report.), 1999 (R), 2008 (T,R), 2009 (R), 2012 (R), 2014 (R)

1975 (T,R), 1977 (R), 1978 (T,R), 1979 (R), 1980 (R), 1981 (T), 1984 (T,R: withdrawal of temporary ADI, but no evaluation)

1991 (T,R), 1992 (R), 1992 (R), 2009 (R), 2010 (R)
1968 (T,R), 1973 (T,R)

| Captafol (006) | $\begin{aligned} & 1969 \text { (T,R), } 1973 \text { (T,R), } 1974 \text { (R), } 1976 \text { (R), } \\ & 1977 \text { (T,R), } 1982 \text { (T), } 1985 \text { (T,R), } 1986 \text { (corr. to } \\ & 1985 \text { report), } 1990 \text { (R), } 1999 \text { (ARfD) } \end{aligned}$ |
| :---: | :---: |
| Captan (007) | 1965 (T), 1969 (T,R), 1973 (T), 1974 (R), 1977 (T,R), 1978 (T,R), 1980 (R), 1982 (T), 1984 (T,R), 1986 ( R ), 1987 ( R and corr. to 1986 R evaluation), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1994 (R), 1995 (T), 1997 (R), 2000 (R), 2004 (T), 2007 (T) |
| Carbaryl (008) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (T,R), } 1968 \text { (R), } \\ & 1969(\mathrm{~T}, \mathrm{R}), 1970 \quad(\mathrm{R}), 1973 \\ & 1976(\mathrm{R}), 1977(\mathrm{R}), 1979(\mathrm{R}), 1984(\mathrm{R}), 1996 \text { (T), }, 1975 \\ & 2001(\mathrm{~T}), 2002(\mathrm{R}), 2007(\mathrm{R}) \end{aligned}$ |
| Carbendazim (072) |  |
| Carbofuran (096) | $\begin{aligned} & 1976 \text { (T,R), } 1979 \text { (T,R), } 1980 \text { (T), } 1982 \text { (T), } \\ & 1991(\mathrm{R}), 1993 \text { (R), } 1996 \text { (T), } 1997 \text { (R), } 1999 \text { (corr. } \\ & \text { to } 1997 \text { report), } 2002 \text { (T,R), } 2003 \text { (R) (See also } \\ & \text { carbosulfan), } 2004 \text { (R), } 2008 \text { (T), } 2009 \text { (R) } \end{aligned}$ |
| Carbon disulfide (009) | 1965 (T,R), 1967 (R), 1968 (R), 1971 (R), 1985 (R) |
| Carbon tetrachloride (010) | $\begin{aligned} & 1965 \text { (T,R), } 1967 \text { (R), } 1968 \text { (T,R), } 1971 \text { (R), } \\ & 1979 \text { (R), } 1985(\mathrm{R}) \end{aligned}$ |
| Carbophenothion (011) | $\begin{aligned} & 1972 \text { (T,R), } 1976 \text { (T,R), } 1977 \text { (T,R), } 1979 \text { (T,R), } \\ & 1980 \text { (T,R), } 1983 \text { (R) } \end{aligned}$ |
| Carbosulfan (145) | 1984 (T,R), 1986 (T), 1991 (R), 1992 (corr. to 1991 report), 1993 (R), 1997 (R), 1999 (R), 2002 (R), 2003 (T,R), 2004 ( R , corr. to 2003 report) |
| Cartap (097) | 1976 (T,R), 1978 (T,R), 1995 (T,R) |
| Chinomethionat (080) | ```1968 (T,R) (as oxythioquinox), 1974 (T,R), 1977 (T,R), 1981 (T,R), 1983 (R), }1984\mathrm{ (T,R), 1987 (T)``` |
| Chlorantraniliprole (230) | 2008 (T,R), 2010 (R), 2013 (R), 2014 (R) |
| Chlorbenside | 1965 (T) |
| Chlordane (012) | $\begin{array}{lllll} 1965(\mathrm{~T}), \quad 1967 & (\mathrm{~T}, \mathrm{R}), \quad 1969 & (\mathrm{R}), 1970 & (\mathrm{~T}, \mathrm{R}), \\ 1972(\mathrm{R}), & 1974 & (\mathrm{R}), \quad 1977 & (\mathrm{~T}, \mathrm{R}), & 1982 \\ 1984(\mathrm{~T}, \mathrm{R}), 1986 & (\mathrm{~T}) & & & \end{array}$ |
| Chlordimeform (013) | 1971 (T,R), 1975 (T,R), 1977 (T), 1978 (T,R), 1979 (T), 1980 (T), 1985 (T), 1986 (R), 1987 (T) |
| Chlorfenapyr (254) | 2013 (T) |
| Chlorfenson | 1965 (T) |
| Chlorfenvinphos (014) | 1971 (T,R), 1984 (R), 1994 (T), 1996 (R) |
| Chlormequat (015) | $\begin{aligned} & 1970 \text { (T,R), } 1972 \text { (T,R), } 1976 \text { (R), } 1985 \text { (R), } \\ & 1994 \text { (T,R), } 1997 \text { (T), } 1999 \text { (ARfD), } 2000 \text { (R) } \end{aligned}$ |

Chlorobenzilate (016)

Chloropicrin
Chloropropylate
Chlorothalonil (081)

Chlorpropham (201)
Chlorpyrifos (017)

Chlorpyrifos-methyl (090)

Chlorthion
Clethodim (187)
Clofentezine (156)

Clothianidin (238)
Coumaphos (018)

Crufomate (019)
Cyanophenfos (091)

Cyantraniliprole (263)
Cyazofamid (281)
Cycloxydim (179)
Cyflumetofen (273)
Cyfluthrin (157)

Cyhalothrin (146)

Cyhexatin (067)

Cypermethrin (118)

1965 (T), 1968 (T,R), 1972 (R), 1975 (R), 1977 (R), 1980 (T)

1965 (T,R)
1968 (T,R), 1972 (R)
1974 (T,R), 1977 (T,R), 1978 (R), 1979 (T,R), 1981 (T,R), 1983 (T,R), 1984 (corr. to 1983 report and T evaluation), 1985 (T,R), 1987 (T), 1988 (R), 1990 (T,R), 1991 (corr. to 1990 evaluation), 1992 (T), 1993 (R), 1997 (R), 2009 (T), 2010 (R), 2012 (R), 2015 (R)

1965 (T), 2000 (T), 2001 (R), 2005 (T), 2008 (R)
1972 (T,R), 1974 (R), 1975 (R), 1977 (T,R), 1981 (R), 1982 (T,R), 1983 (R), 1989 (R), 1995 (R), 1999 (T), 2000 (R), 2004 (R), 2006 (R)

1975 (T,R), 1976 (R, Annex I only), 1979 (R), 1990 (R), 1991 (T,R), 1992 (T and corr. to 1991 report), 1993 (R), 1994 (R), 2001 (T), 2009 (R)

1965 (T)
1994 (T,R), 1997 (R), 1999 (R), 2002 (R)
1986 (T,R), 1987 (R), 1989 (R), 1990 (R), 1992 (R), 2005 (T), 2007 (R)

2010 (T,R), 2011 (R), 2014 (R)
1968 (T,R), 1972 (R), 1975 (R), 1978 (R), 1980 (T,R), 1983 (R), 1987 (T), 1990 (T,R)

1968 (T,R), 1972 (R)
1975 (T,R), 1978 (T: ADI extended, but no evaluation), 1980 (T), 1982 (R), 1983 (T)

2013 (T,R), 2015 (R)
2015 (T, R)
1992 (T,R), 1993 (R), 2009 (T), 2012 (R)
2014 (T,R)
1986 (R), 1987 ( T and corr. to 1986 report), 1989 (R), 1990 (R), 1992 (R), 2006 (T), 2007 (R)

1984 (T,R), 1986 (R), 1988 (R), 2007 (T), 2008 (R), 2015 (R)

1970 (T,R), 1973 (T,R), 1974 (R), 1975 (R), 1977 (T), 1978 (T,R), 1980 (T), 1981 (T), 1982 (R), 1983 (R), 1985 (R), 1988 (T), 1989 (T), 1991 (T,R), 1992 (R), 1994 (T), 2005 (T,R)

1979 (T,R), 1981 (T,R), 1982 (R), 1983 (R), 1984 (R), 1985 (R), 1986 (R), 1987 (corr. to 1986 evaluation), 1988 (R), 1990 (R), 2006 (T), 2008 (R), 2009 (R), 2011 (R)

Cyproconazole (239)
Cyprodinil (207)

Cyromazine (169)

2,4-D (020)

Daminozide (104)
DDT (021)

Deltamethrin (135)

Demeton (092)
Demeton-S-methyl (073)

Demeton-S-methylsulfon (164)

Dialifos (098)
Diazinon (022)

1,2-Dibromoethane (023)

Dicamba (240)
Dichlobenil (274)
Dicloran (083)
Dichlorfluanid (082)

1,2-Dichloroethane (024)
Dichlorvos (025)

Dicloran (083)
Dicofol (026)

Dieldrin (001)

2010 (T,R), 2013 (R)
2003 (T,R), 2004 (corr. to 2003 report), 2013 (R), 2015 (R)

1990 (T,R), 1991 (corr. to 1990 R evaluation), 1992 (R), 2006 (T), 2007 (R), 2012 (R)

1970 (T,R), 1971 (T,R), 1974 (T,R), 1975 (T,R), 1980 (R), 1985 (R), 1986 (R), 1987 (corr. to 1986 report, Annex I), 1996 (T), 1997 (E), 1998 (R), 2001 (R)

1977 (T,R), 1983 (T), 1989 (T,R), 1991 (T)
1965 (T), 1966 (T,R), 1967 (T,R), 1968 (T,R), 1969 (T,R), 1978 (R), 1979 (T), 1980 (T), 1983 (T), 1984 (T), 1993 (R), 1994 (R), 1996 (R)

1980 (T,R), 1981 (T,R), 1982 (T,R), 1984 (R), 1985 (R), 1986 (R), 1987 (R), 1988 (R), 1990 (R), 1992 (R), 2000 (T), 2002 (R)

1965 (T), 1967 (R), 1975 (R), 1982 (T)
1973 (T,R), 1979 (R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R), 1998 (R)

1973 (T,R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R)

1976 (T,R), 1982 (T), 1985 (R)
1965 (T), 1966 (T), 1967 (R), 1968 (T,R), 1970 (T,R), 1975 (R), 1979 (R), 1993 (T,R), 1994 (R), 1996 (R), 1999 (R), 2001 (T), 2006 (T,R), 2016 (T)

1965 (T,R), 1966 (T,R), 1967 (R), 1968 (R), 1971 (R), 1979 (R), 1985 (R)

2010 (T,R), 2011 (R), 2012 (R), 2013 (R)
2014 (T,R)
2003 (R)
1969 (T,R), 1974 (T,R), 1977 (T,R), 1979 (T,R), 1981 (R), 1982 (R), 1983 (T,R), 1985 (R)

1965 (T,R), 1967 (R), 1971 (R), 1979 (R), 1985 (R)
1965 (T,R), 1966 (T,R), 1967 (T,R), 1969 (R), 1970 (T,R), 1974 (R), 1977 (T), 1993 (T,R), 2011 (T), 2012 (R)

1974 (T,R), 1977 (T,R), 1998 (T,R)
1968 (T,R), 1970 (R), 1974 (R), 1992 (T,R), 1994 (R), 2011 (T), 2012 (R)

1965 (T), 1966 (T,R), 1967 (T,R), 1968 (R), 1969 (R), 1970 (T,R), 1974 (R), 1975 (R), 1977 (T), 1990 (R), 1992 (R)

Difenoconazole (224)
Diflubenzuron (130)

Dimethenamid-P (214)
Dimethipin (151)

Dimethoate (027)

Dimethomorph (225)
Dimethrin
Dinocap (087)

Dinotefuran (255)
Dioxathion (028)
Diphenyl (029)
Diphenylamine (030)

Diquat (031)

Disulfoton (074)

Dithianon (180)

Dithiocarbamates (105)

4,6-Dinitro-ortho-cresol (DNOC)
Dodine (084)

Edifenphos (099)
Emamectin benzoate (247)
Endosulfan (032)

Endrin (033)

2007 (T,R), 2010 (R), 2013 (R), 2015 (R)
1981 (T,R), 1983 (R), 1984 (T,R), 1985 (T,R), 1988 (R), 2001 (T), 2002 (R), 2011 (R)

2005 (T,R)
1985 (T,R), 1987 (T,R), 1988 (T,R), 1999 (T), 2001 (R), 2004 (T)

1965 (T), 1966 (T), 1967 (T,R), 1970 (R), 1973 (R in evaluation of formothion), 1977 (R), 1978 (R), 1983 (R) 1984 (T,R), 1986 (R), 1987 (T,R), 1988 (R), 1990 (R), 1991 (corr. to 1990 evaluation), 1994 (R), 1996 (T), 1998 (R), 2003 (T,R), 2004 (corr. to 2003 report), 2006 (R), 2008 (R)

2007 (T,R), 2014 (R)
1965 (T)
1969 (T,R), 1974 (T,R), 1989 (T,R), 1992 (R), 1998 (R), 1999 (R), 2000 (T), 2001 (R)

2012 (T,R)
1968 (T,R), 1972 (R)
1966 (T,R), 1967 (T)
1969 (T,R), 1976 (T,R), 1979 (R), 1982 (T), 1984 (T,R), 1998 (T), 2001 (R), 2003 (R), 2008 (R)

1970 (T,R), 1972 (T,R), 1976 (R), 1977 (T,R), 1978 (R), 1994 (R), 2013 (T,R)

1973 (T,R), 1975 (T,R), 1979 ( R ), 1981 (R), 1984 (R), 1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1996 (T), 1998 (R), 2006 (R)

1992 (T,R), 1995 (R), 1996 (corr. to 1995 report), 2010 (T), 2013 (T,R)

1965 (T), 1967 (T,R), 1970 (T,R), 1983 (R propineb, thiram), 1984 ( R propineb), 1985 ( R ), 1987 ( T thiram), 1988 ( R thiram), 1990 (R), 1991 (corr. to 1990 evaluation), 1992 ( T thiram), 1993 (T,R), 1995 (R), 1996 (T,R ferbam, ziram; R thiram), 2004 (R), 2012 (R), 2014 (R)

1965 (T)
1974 (T,R), 1976 (T,R), 1977 (R), 2000 (T), 2003 (R), 2004 (corr. to 2003 report)

1976 (T,R), 1979 (T,R), 1981 (T,R)
2011 (T,R), 2014 (R)
1965 (T), 1967 (T,R), 1968 (T,R), 1971 (R), 1974 (R), 1975 (R), 1982 (T), 1985 (T,R), 1989 (T,R), 1993 (R), 1998 (T), 2006 (R), 2010 (R)

1965 (T), 1970 (T,R), 1974 (R), 1975 (R), 1990 (R), 1992 (R)

Esfenvalerate (204)
Ethephon (106)

Ethiofencarb (107)

Ethion (034)

Ethoprophos (149)
Ethoxyquin (035)
Ethylene dibromide
Ethylene dichloride
Ethylene oxide
Ethylenethiourea (ETU) (108)

Etofenprox (184)
Etoxazole (241)
Etrimfos (123)

Famoxadone (208)
Fenamidone (264)
Fenamiphos (085)

Fenarimol (192)
Fenbuconazole (197)
Fenbutatin oxide (109)
Fenchlorfos (036)
Fenhexamid (215)
Fenitrothion (037)

Fenpropathrin (185)
Fenpropimorph (188)
Fenpyroximate (193)

Fensulfothion (038)
Fenthion (039)

2002 (T,R)
1977 (T,R), 1978 (T,R), 1983 (R), 1985 (R), 1993 (T), 1994 (R), 1995 (T), 1997 (T), 2002 (T), 2015 (T, R)

1977 (T,R), 1978 (R), 1981 (R), 1982 (T,R), 1983 (R)

1968 (T,R), 1969 (R), 1970 (R), 1972 (T,R), 1975 (R), 1982 (T), 1983 (R), 1985 (T), 1986 (T), 1989 (T), 1990 (T), 1994 (R)

1983 (T), 1984 (R), 1987 (T), 1999 (T), 2004 (R)
1969 (T,R), 1998 (T), 1999 (R), 2005 (T), 2008 (R)
See 1,2-Dibromoethane
See 1,2-Dichloroethane
1965 (T,R), 1968 (T,R), 1971 (R)
1974 (R), 1977 (T,R), 1986 (T,R), 1987 (R), 1988 (T,R), 1990 (R), 1993 (T,R)

1993 (T,R), 2011 (T,R)
2010 (T,R), 2011 (R)
1980 (T,R), 1982 (T,R), 1986 (T,R), 1987 (R), 1988 (R), 1989 (R), 1990 (R)

2003 (T,R)
2013 (T), 2014 (T,R)
1974 (T,R), 1977 (R), 1978 (R), 1980 (R), 1985 (T), 1987 (T), 1997 (T), 1999 (R), 2002 (T), 2006 (R)

1995 (T,R,E), 1996 (R and corr. to 1995 report)
1997 (T,R), 2009 (R), 2012 (T), 2013 (R)
1977 (T,R), 1979 (R), 1992 (T), 1993 (R)
1968 (T,R), 1972 (R), 1983 (R)
2005 (T,R)
1969 (T,R), 1974 (T,R), 1976 (R), 1977 (T,R), 1979 (R), 1982 (T), 1983 ( R ), 1984 (T,R), 1986 (T,R), 1987 ( R and corr. to 1986 R evaluation), 1988 (T), 1989 (R), 2000 (T), 2003 (R), 2004 (R, corr. to 2003 report), 2007 (T,R)

1993 (T,R), 2006 (R), 2012 (T), 2014 (R)
1994 (T), 1995 (R), 1999 (R), 2001 (T), 2004 (T)
1995 (T,R), 1996 (corr. to 1995 report), 1999 (R), 2004 (T), 2007 (T), 2010 (R), 2013 (R)

1972 (T,R), 1982 (T), 1983 (R)
1971 (T,R), 1975 (T,R), 1977 (R), 1978 (T,R), 1979 (T), 1980 (T), 1983 (R), 1989 (R),

Fentin compounds (040)

Fenvalerate (119)

Ferbam

Fipronil (202)
Fipronil-desulfinyl
Flonicamid (282)
Flubendiamide (242)
Flucythrinate (152)

Fludioxonil (211)
Fluensulfone (265)
Flufenoxuron (275)
Flumethrin (195)
Fluopicolide (235)
Fluopyram (243)
Flupyradifurone (285)
Flusilazole (165)

Flutolanil (205)
Flutriafol (248)
Fluxapyroxad (256)
Folpet (041)

Formothion (042)

Glufosinate-ammonium (175)

Glyphosate (158)

Guazatine (114)

1995 (T,R,E), 1996 (corr. to 1995 report), 1997 (T), 2000 (R)

1965 (T), 1970 (T,R), 1972 (R), 1986 (R), 1991 (T,R), 1993 (R), 1994 (R)

1979 (T,R), 1981 (T,R), 1982 (T), 1984 (T,R), 1985 (R), 1986 (T,R), 1987 ( R and corr. to 1986 report), 1988 (R), 1990 (R), 1991 (corr. to 1990 R evaluation), 2012 (T,R)

See Dithiocarbamates, 1965 (T), 1967 (T,R), 1996 (T,R)

1997 (T), 2000 (T), 2001 (R)
1997 (T)
2015 (T,R)
2010 (T,R)
1985 (T,R), 1987 (R), 1988 (R), 1989 (R), 1990 (R), 1993 (R)

2004 (T,R), 2006 (R), 2010 (R), 2012 (R), 2013 (R)
2013 (T), 2014 (T,R)
2014 (T,R)
1996 (T,R)
2009 (T,R), 2014 (R)
2010 (T,R), 2012 (R), 2014 (R), 2015 (R)
2015 (T)
1989 (T,R), 1990 (R), 1991 (R), 1993 (R), 1995 (T), 2007 (T,R)

2002 (T,R), 2013 (R)
2011 (T,R), 2015 (R)
2012 (T,R), 2015 (R)
1969 (T,R), 1973 (T), 1974 (R), 1982 (T), 1984 (T,R), 1986 (T), 1987 (R), 1990 (T,R), 1991 (corr. to 1990 R evaluation), 1993 (T,R), 1994 (R), 1995 (T), 1997 (R), 1998 (R), 1999 (R) , 2002 (T), 2004 (T), 2007 (T)

1969 (T,R), 1972 (R), 1973 (T,R), 1978 (R), 1998 (R)

1991 (T,R), 1992 (corr. to 1991 report, Annex I), 1994 (R), 1998 (R), 1999 (T,R), 2012 (T,R), 2014 (R)

1986 (T,R), 1987 ( R and corr. to 1986 report), 1988 (R), 1994 (R), 1997 (T,R), 2004 (T), 2005 (R), 2011 (T,R), 2013 (R), 2016 (T)

1978 (T,R), 1980 (R), 1997 (T,R)

| Haloxyfop (194) | 1995 (T,R), 1996 ( R and corr. to 1995 report), 2001 (R), 2006 (T), 2009 (R) |
| :---: | :---: |
| Heptachlor (043) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T,R), } 1967 \text { (R), } 1968 \text { (R), } 1969 \text { (R), } \\ & 1970 \text { (T,R), } 1974 \text { (R), } 1975 \text { (R), } 1977 \text { (R), } 1987 \text { (R), } \\ & 1991 \text { (T,R), } 1992 \text { (corr. to } 1991 \text { report, Annex I), } \\ & 1993 \text { (R), } 1994 \text { (R) } \end{aligned}$ |
| Hexachlorobenzene (044) | $\begin{aligned} & 1969 \text { (T,R), } 1973 \text { (T,R), } 1974 \text { (T,R), } 1978 \text { (T), } \\ & 1985 \text { (R) } \end{aligned}$ |
| Hexaconazole (170) | 1990 (T,R), 1991 ( R and corr. to 1990 R evaluation), 1993 (R) |
| Hexythiazox (176) | 1991 (T,R), 1994 (R), 1998 (R), 2008 (T), 2009 (R) |
| Hydrogen cyanide (045) | 1965 (T,R) |
| Hydrogen phosphide (046) | $\begin{aligned} & 1965 \text { (T,R), } 1966 \text { (T,R), } 1967 \text { (R), } 1969 \text { (R), } \\ & 1971 \text { (R) } \end{aligned}$ |
| Imazalil (110) | $\begin{aligned} & 1977 \text { (T,R), } 1980(\mathrm{~T}, \mathrm{R}), 1984(\mathrm{~T}, \mathrm{R}), 1985(\mathrm{~T}, \mathrm{R}), \\ & 1986(\mathrm{~T}), 1988(\mathrm{R}), 1989(\mathrm{R}), 1991(\mathrm{~T}), 1994(\mathrm{R}), \\ & 2000(\mathrm{~T}), 2001(\mathrm{~T}), 2005(\mathrm{~T}) \end{aligned}$ |
| Imazamox (276) | 2014 (T,R) |
| Imazapic (266) | 2013 (T,R), 2015 (R) |
| Imazapyr (267) | 2013 (T,R), 2015 (R) |
| Imidacloprid (206) | $\begin{aligned} & 2001 \text { (T), } 2002 \text { (R), } 2006 \text { (R), } 2008 \text { (R), } 2012 \text { (R), } \\ & 2015 \text { (R) } \end{aligned}$ |
| Indoxacarb (216) | 2005 (T,R), 2007 (R), 2009 (R), 2012 (R), 2013 (R) |
| Iprodione (111) | $\begin{aligned} & 1977 \text { (T,R), } 1980 \text { (R), } 1992 \text { (T), } 1994 \text { (R), } 1995 \text { (T), } \\ & 2001 \text { (R) } \end{aligned}$ |
| Isofenphos (131) | $\begin{aligned} & 1981 \text { (T,R), } 1982 \text { (T,R), } 1984 \text { (R), } 1985 \text { (R), } \\ & 1986(T, R), 1988(R), 1992(R) \end{aligned}$ |
| Isopyrazam (249) | 2011 (T,R) |
| Isoxaflutole (268) | 2013 (T,R) |
| Kresoxim-methyl (199) | 1998 (T,R), 2001 (R) |
| Lead arsenate | 1965 (T), 1968 (T,R) |
| Leptophos (088) | 1974 (T,R), 1975 (T,R), 1978 (T,R) |
| Lindane (048) | 1965 (T), 1966 (T,R), 1967 (R), 1968 (R), 1969 (R), 1970 (T,R, published as Annex VI to 1971 evaluations), 1973 (T,R), 1974 (R), 1975 (R), 1977 (T,R), 1978 (R), 1979 (R), 1989 (T,R), 1997 (T), 2002 (T), 2003 (R), 2004 (corr. to 2003 report), 2015 (R) |
| Lufenuron (286) | 2015 (T, R) |
| Malathion (049) | 1965 (T), 1966 (T,R), 1967 (corr. to 1966 R evaluation), 1968 (R), 1969 (R), 1970 (R), 1973 (R), 1975 (R), 1977 (R), 1984 (R), 1997 (T), 1999 (R), |

Maleic hydrazide (102)

Mancozeb (050)

Mandipropamid (231)
Maneb

MCPA (257)
Mecarbam (124)

Meptyldinocap (244)
Mesotrione (277)
Metaflumizone (236)
Metalaxyl (138)

Metalaxyl -M (212)
Methacrifos (125)

Methamidophos (100)

Methidathion (051)

Methiocarb (132)

Methomyl (094)

Methoprene (147)

Methoxychlor
Methoxyfenozide (209)

Methyl bromide (052)
Metrafenone (278)
Metiram (186)
Mevinphos (053)
MGK 264

2000 (R), 2003 (T), 2004 (R), 2005 (R), 2008 (R), 2013 (R), 2016 (T)

1976 (T,R), 1977 (T,R), 1980 (T), 1984 (T,R), 1996 (T), 1998 (R)

1967 (T,R), 1970 (T,R), 1974 (R), 1977 (R), 1980 (T,R), 1993 (T,R)

2008 (T,R), 2013 (R)
See Dithiocarbamates, 1965 (T), 1967 (T,R), 1987 (T), 1993 (T,R)

2012 (T,R)
1980 (T,R), 1983 (T,R), 1985 (T,R), 1986 (T,R), 1987 (R)

2010 (T,R)
2014 (T,R)
2009 (T,R)
1982 (T,R), 1984 (R), 1985 (R), 1986 (R), 1987 (R), 1989 (R), 1990 (R), 1992 (R), 1995 (R)

2002 (T), 2004 (R)
1980 (T,R), 1982 (T), 1986 (T), 1988 (T), 1990 (T,R), 1992 (R)

1976 (T,R), 1979 (R), 1981 (R), 1982 (T,R), 1984 (R), 1985 (T), 1989 (R), 1990 (T,R), 1994 (R), 1996 (R), 1997 (R), 2002 (T), 2003 (R), 2004 (R, corr. to 2003 report)

1972 (T,R), 1975 (T,R), 1979 (R), 1992 (T,R), 1994 (R), 1997 (T)

1981 (T,R), 1983 (T,R), 1984 (T), 1985 (T), 1986 (R), 1987 (T,R), 1988 (R), 1998 (T), 1999 (R), 2005 (R)

1975 (R), 1976 (R), 1977 (R), 1978 (R), 1986 (T,R), 1987 (R), 1988 (R), 1989 (T,R), 1990 (R), 1991 (R), 2001 (T,R), 2004 (R), 2008 (R)

1984 (T,R), 1986 (R), 1987 (T and corr. to 1986 report), 1988 (R), 1989 (R), 2001 (T), 2005 (R)

1965 (T), 1977 (T)
2003 (T,R), 2004 (corr. to 2003 report), 2006 (R), 2009 (R), 2012 (R)

See Bromomethane
2014 (T,R)
1993 (T), 1995 (R)
1965 (T), 1972 (T,R), 1996 (T), 1997 (E,R), 2000 (R) 1967 (T,R)

| Monocrotophos (054) | $\begin{aligned} & 1972 \text { (T,R), } 1975 \text { (T,R), } 1991 \text { (T,R), } 1993 \text { (T), } \\ & 1994 \text { (R) } \end{aligned}$ |
| :---: | :---: |
| Myclobutanil (181) | $\begin{aligned} & 1992 \text { (T,R), } 1997 \text { (R), } 1998 \text { (R), (2001 (R)), } 2014 \\ & (T, R) \end{aligned}$ |
| Nabam | See Dithiocarbamates, 1965 (T), 1976 (T,R) |
| Nitrofen (140) | 1983 (T,R) |
| Novaluron (217) | 2005 (T,R), 2010 (R) |
| Omethoate (055) | $\begin{aligned} & 1971 \text { (T,R), } 1975 \text { (T,R), } 1978 \text { (T,R), } 1979 \text { (T), } \\ & 1981 \text { (T,R), } 1984(\mathrm{R}), 1985 \text { (T), } 1986 \text { (R), } 1987 \text { (R), } \\ & 1988 \text { (R), } 1990 \text { (R), } 1998 \text { (R) } \end{aligned}$ |
| Organomercury compounds | 1965 (T), 1966 (T,R), 1967 (T,R) |
| Oxamyl (126) | $\begin{aligned} & 1980 \text { (T,R), } 1983 \text { (R), } 1984 \text { (T), } 1985 \text { (T,R), } \\ & 1986 \text { (R), } 2002(\mathrm{~T}, \mathrm{R}) \end{aligned}$ |
| Oxydemeton-methyl (166) | 1965 (T, as demeton-S-methyl sulfoxide), 1967 (T), 1968 (R), 1973 (T,R), 1982 (T), 1984 (T,R), 1989 (T,R), 1992 (R), 1998 (R), 1999 (corr. to 1992 report), 2002 (T), 2004 (R) |
| Oxythioquinox | See Chinomethionat |
| Paclobutrazol (161) | 1988 (T,R), 1989 (R) |
| Paraquat (057) | $\begin{aligned} & 1970 \text { (T,R), } 1972 \text { (T,R), } 1976 \text { (T,R), } 1978 \text { (R), } \\ & 1981 \text { (R), } 1982 \text { (T), } 1985 \text { (T), } 1986 \text { (T), } 2003 \text { (T), } \\ & 2004 \text { (R), } 2009 \text { (R) } \end{aligned}$ |
| Parathion (058) | $\begin{aligned} & 1965 \text { (T), } 1967 \text { (T,R), } 1969 \text { (R), } 1970 \text { (R), } 1984 \text { (R), } \\ & 1991 \text { (R), } 1995 \text { (T,R), } 1997 \text { (R), } 2000 \text { (R) } \end{aligned}$ |
| Parathion-methyl (059) | $\begin{aligned} & 1965(\mathrm{~T}), 1968(\mathrm{~T}, \mathrm{R}), 1972(\mathrm{R}), 1975 \\ & 1978(\mathrm{~T}, \mathrm{R}), \quad 1979 \quad(\mathrm{~T}), \mathrm{R}), 1980(\mathrm{~T}), 1982(\mathrm{~T}), \\ & 1984(\mathrm{~T}, \mathrm{R}), 1991(\mathrm{R}), 1992(\mathrm{R}), 1994(\mathrm{R}), 1995(\mathrm{~T}), \\ & 2000(\mathrm{R}), 2003(\mathrm{R}) \end{aligned}$ |
| Penconazole (182) | 1992 (T,R), 1995 (R), 2015 (T) |
| Penthiopyrad (253) | 2011 (T), 2012 (R), 2013 (R) |
| Permethrin (120) | 1979 (T,R), 1980 (R), 1981 (T,R), 1982 (R), 1983 (R), 1984 (R), 1985 (R), 1986 (T,R), 1987 (T), 1988 (R), 1989 (R), 1991 (R), 1992 (corr. to 1991 report), 1999 (T) |
| 2-Phenylphenol (056) | $\begin{aligned} & 1969(\mathrm{~T}, \mathrm{R}), \quad 1975 \quad(\mathrm{R}), 1983 \quad(\mathrm{~T}), 1985 \quad(\mathrm{~T}, \mathrm{R}), \\ & 1989(\mathrm{~T}), 1990(\mathrm{~T}, \mathrm{R}), 1999(\mathrm{~T}, \mathrm{R}), 2002(\mathrm{R}) \end{aligned}$ |
| Phenothrin (127) | $\begin{aligned} & 1979 \text { (R), } 1980 \text { (T,R), } 1982 \text { (T), } 1984 \text { (T), } 1987 \text { (R), } \\ & 1988 \text { (T,R) } \end{aligned}$ |
| Phenthoate (128) | 1980 (T,R), 1981 (R), 1984 (T) |
| Phorate (112) | $\begin{aligned} & 1977 \text { (T,R), } 1982 \text { (T), } 1983 \text { (T), } 1984 \text { (R), } 1985 \text { (T), } \\ & 1990 \text { (R), } 1991 \text { (R), } 1992 \text { (R), } 1993 \text { (T), } 1994 \text { (T), } \\ & 1996 \text { (T), } 2004 \text { (T), } 2005 \text { (R), } 2012 \text { (R), } 2014 \text { (R) } \end{aligned}$ |
| Phosalone (060) | $\begin{aligned} & 1972 \text { (T,R), } 1975 \text { (R), } 1976 \text { (R), } 1993 \text { (T), } 1994 \text { (R), } \\ & 1997 \text { (T), } 1999 \text { (R), } 2001 \text { (T) } \end{aligned}$ |


| Phosmet (103) | 1976 (R), 1977 (corr. to 1976 R evaluation), 1978 (T,R), 1979 (T,R), 1981 (R), 1984 (R), 1985 (R), 1986 (R), 1987 ( R and corr. to 1986 R evaluation), 1988 (R), 1994 (T), 1997 (R), 1998 (T), 2002 (R), 2003 (R), 2007 (R) |
| :---: | :---: |
| Phosphine | See Hydrogen phosphide |
| Phosphamidon (061) | $\begin{aligned} & 1965 \text { (T), } 1966 \text { (T), } 1968 \text { (T,R), } 1969 \text { (R), } 1972 \text { (R), } \\ & 1974 \text { (R), } 1982 \text { (T), } 1985 \text { (T), } 1986 \text { (T) } \end{aligned}$ |
| Phoxim (141) | $\begin{aligned} & 1982 \text { (T), } 1983 \text { (R), } 1984 \text { (T,R), } 1986 \text { (R), } 1987 \text { (R), } \\ & 1988 \text { (R) } \end{aligned}$ |
| Picoxystrobin (258) | 2012 (T,R), 2013 (R) |
| Piperonyl butoxide (062) | 1965 (T,R), 1966 (T,R), 1967 (R), 1969 (R), 1972 (T,R), 1992 (T,R), 1995 (T), 2001 (R), 2002 (R) |
| Pirimicarb (101) | $\begin{aligned} & 1976 \text { (T,R), } 1978 \text { (T,R), } 1979 \text { (R), } 1981 \text { (T,R), } \\ & 1982 \text { (T), } 1985 \text { (R), } 2004 \text { (T), } 2006 \text { (R) } \end{aligned}$ |
| Pirimiphos-methyl (086) | 1974 (T,R), 1976 (T,R), 1977 (R), 1979 (R), 1983 (R), 1985 (R), 1992 (T), 1994 (R), 2003 (R), 2004 ( R , corr. to 2003 report), 2006 (T) |
| Prochloraz (142) | 1983 (T,R), 1985 (R), 1987 (R), 1988 (R), 1989 (R), $1990(\mathrm{R}), 1991$ (corr. to 1990 report, Annex I, and R evaluation), 1992 (R), 2001 (T), 2004 (R), 2009 (R) |
| Procymidone(136) | 1981 (R), 1982 (T), 1989 (T,R), 1990 (R), 1991 (corr. to 1990 Annex I), 1993 (R), 1998 (R), 2007 (T) |
| Profenofos (171) | $\begin{aligned} & 1990 \text { (T,R), } 1992 \text { (R), } 1994 \text { (R), } 1995 \text { (R), } 2007 \text { (T), } \\ & 2008 \text { (R), } 2011 \text { (R) } \end{aligned}$ |
| Propamocarb (148) | $\begin{aligned} & 1984 \text { (T,R), } 1986 \text { (T,R), } 1987 \text { (R), } 2005 \text { (T), } \\ & 2006 \text { (R), } 2014 \text { (R) } \end{aligned}$ |
| Propargite (113) | $\begin{aligned} & 1977 \text { (T,R), } 1978 \text { (R), } 1979 \text { (R), } 1980 \text { (T,R), } \\ & 1982 \text { (T,R), } 1999 \text { (T), } 2002 \text { (R), } 2006 \text { (R) } \end{aligned}$ |
| Propham (183) | 1965 (T), 1992 (T,R) |
| Propiconazole (160) | $\begin{aligned} & 1987 \text { (T,R), } 1991 \text { (R), } 1994 \text { (R), } 2004 \text { (T), } 2006 \text { (R), } \\ & 2007 \text { (R), } 2013 \text { (R), } 2014 \text { (R), } 2015 \text { (R) } \end{aligned}$ |
| Propineb | $\begin{array}{lllll} 1977 \text { (T,R), } 1980 \text { (T), } 1983 \text { (T), } 1984 \text { (R), } \\ 1985(T, R), 1993(T, R), 2004(R) \end{array}$ |
| Propoxur (075) | $\begin{aligned} & 1973 \text { (T,R), } 1977 \text { (R), } 1981 \text { (R), } 1983 \text { (R), } 1989 \text { (T), } \\ & 1991 \text { (R), } 1996 \text { (R) } \end{aligned}$ |
| Propylene oxide (250) | 2011 (T,R) |
| Propylenethiourea (PTU, 150) | 1993 (T,R), 1994 (R), 1999 (T) |
| Prothioconazole (232) | 2008 (T,R), 2009 (R), 2014 (R) |
| Pymetrozine (279) | 2014 (T,R) |
| Pyraclostrobin (210) | $\begin{aligned} & 2003 \text { (T), } 2004 \text { (R), } 2006 \text { (R), } 2011 \text { (R), } 2012 \text { (R), } \\ & 2014 \text { (R) } \end{aligned}$ |
| Pyrazophos (153) | 1985 (T,R), 1987 (R), 1992 (T,R), 1993 (R) |

Pyrethrins (063)

Pyrimethanil (226)
Pyriproxyfen (200)
Quinclorac (287)
Quinoxyfen (223)
Quintozene (064)

Saflufenacil (251)
Sedaxane (259)
Spices
Spinetoram (233)
Spinosad (203)
Spirodiclofen (237)
Spirotetramat (234)
Sulfoxaflor (252)
Sulfuryl fluoride (218)
2,4,5-T (121)
Tebuconazole (189)

Tebufenozide (196)

Tecnazine (115)

Teflubenzuron (190)
Temephos
Terbufos (167)
Thiabendazole (065)

Thiacloprid (223)
Thiamethoxam (245)
Thiodicarb (154)

Thiometon (076)

1965 (T), 1966 (T,R), 1967 (R), 1968 (R), 1969 (R), 1970 (T), 1972 (T,R), 1974 (R), 1999 (T), 2000 (R), 2003 (T,R), 2005 (R)

2007 (T,R), 2013 (R)
1999 (R,T), 2000 (R), 2001 (T)
2015 (T, R)
2006 (T,R)
1969 (T,R), 1973 (T,R), 1974 (R), 1975 (T,R), 1976 (Annex I, corr. to 1975 R evaluation), 1977 (T,R), 1995 (T,R), 1998 (R)

2011 (T,R)
2012 (T,R), 2014 (R)
2004 (R), 2005 (R), 2007 (R), 2010 (R), 2015 (R)
2008 (T,R), 2012 (R)
2001 (T,R), 2004 (R), 2008 (R), 2011 (R)
2009 (T,R)
2008 (T,R), 2011 (R), 2012 (R), 2013 (R), 2015 (R)
2011 (T,R), 2013 (R), 2014 (R)
2005 (T,R)
1970 (T,R), 1979 (T,R), 1981 (T)
1994 (T,R), 1996 (corr. to Annex II of 1995 report), 1997 (R), 2008 (R), 2010 (T), 2011 (R), 2015 (R)

1996 (T,R), 1997 (R), 1999 (R), 2001 (T,R), 2003 (T)

1974 (T,R), 1978 (T,R), 1981 (R), 1983 (T), 1987 (R), 1989 (R), 1994 (T,R)

1994 (T), 1996 (R)
2006 (T)
1989 (T,R), 1990 (T,R), 2003 (T), 2005 (R)
1970 (T,R), 1971 (R), 1972 (R), 1975 (R), 1977 (T,R), 1979 (R), 1981 (R), 1997 (R), 2000 (R), 2006 (T,R)

2006 (T,R)
2010 (T,R), 2011 (R), 2012 (R), 2014 (R)
1985 (T,R), 1986 (T), 1987 (R), 1988 (R), 2000 (T),
2001 (R)
1969 (T,R), 1973 (T,R), 1976 (R), 1979 (T,R), 1988 (R)

| Thiophanate-methyl (077) | $\begin{aligned} & 1973 \text { (T,R), } 1975 \text { (T,R), } 1977 \text { (T), } 1978 \text { (R), } \\ & 1988 \text { (R), } 2002(\mathrm{R}), 1990(\mathrm{R}), 1994 \text { (R), } 1995(\mathrm{~T}, \mathrm{E}), \\ & 1998(\mathrm{~T}, \mathrm{R}), 2006(\mathrm{~T}) \end{aligned}$ |
| :---: | :---: |
| Thiram (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1970 (T,R), 1974 (T), 1977 (T), 1983 (R), 1984 (R), 1985 (T,R), 1987 (T), 1988 (R), 1989 (R), 1992 (T), 1996 (R) |
| Tolclofos-methyl (191) | 1994 (T,R), 1996 (corr. to Annex II of 1995 report) |
| Tolfenpyrad (269) | 2013 (T) |
| Tolylfluanid (162) | $\begin{aligned} & 1988 \text { (T,R), } 1990 \text { (R), } 1991 \text { (corr. to } 1990 \text { report), } \\ & 2002(\mathrm{~T}, \mathrm{R}), 2003 \text { (R) } \end{aligned}$ |
| Toxaphene | See Camphechlor |
| Triadimefon (133) | 1979 (R), 1981 (T,R), 1983 (T,R), 1984 (R), 1985 (T,R), 1986 (R), 1987 ( R and corr. to 1986 R evaluation), 1988 (R), 1989 (R), 1992 (R), 1995 (R), 2004 (T), 2007 (R) |
| Triadimenol (168) | $\begin{aligned} & 1989 \text { (T,R), } 1992 \text { (R), } 1995 \text { (R), } 2004 \text { (T), } 2007 \text { (R), } \\ & 2014 \text { (R) } \end{aligned}$ |
| Triazolylalanine | 1989 (T,R) |
| Triazophos (143) | 1982 (T), 1983 (R), 1984 (corr. to 1983 report, Annex I), 1986 (T,R), 1990 (R), 1991 (T and corr. to 1990 R evaluation), 1992 (R), 1993 (T,R), 2002 (T), 2007 (R), 2010 (R), 2013 (R) |
| Trichlorfon (066) | 1971 (T,R), 1975 (T,R), 1978 (T,R), 1987 (R) |
| Trichloronat | 1971 (T,R) |
| Trichloroethylene | 1968 (R) |
| Tricyclohexyltin hydroxide | See Cyhexatin |
| Trifloxystrobin (213) | 2004 (T,R), 2012 (R), 2015 (R) |
| Triflumizole (270) | 2013 (T,R) |
| Triforine (116) | 1977 (T), 1978 (T,R), 1997 (T), 2004 (R), 2014 (T,R) |
| Trinexapac-ethyl (271) | 2013 (T,R) |
| Triphenyltin compounds | See Fentin compounds |
| Vamidothion (078) | $\begin{aligned} & 1973 \text { (T,R), } 1982 \text { (T), } 1985 \text { (T,R), } 1987 \text { (R), } \\ & 1988 \text { (T), } 1990(\mathrm{R}), 1992(\mathrm{R}) \end{aligned}$ |
| Vinclozolin (159) | 1986 (T,R), 1987 ( R and corr. to 1986 report and R evaluation), 1988 (T,R), 1989 (R), 1990 (R), 1992 (R), 1995 (T) |
| Zineb (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1993 (T) |
| Ziram (105) | See Dithiocarbamates, 1965 (T), 1967 (T,R), 1996 (T,R) |
| Zoxamide (227) | 2007 (T,R), 2009 (R) |

## Annex 3

ANNEX 3: INTERNATIONAL ESTIMATED DAILY INTAKES OF PESTICIDE RESIDUES

DIAZINON



## Annex 3


$\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg}$ bw


## Annex 3

| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expras | $\begin{gathered} \text { STMR } \\ \mathrm{mg} / \mathrm{kg} \end{gathered}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | G07 <br> diet | G07 <br> intake | G08 <br> diet | G08 <br> intake | G09 diet | G09 <br> intake | G10 <br> diet | $\begin{gathered} \text { G10 } \\ \text { intake } \end{gathered}$ | G11 diet | G11 <br> intake | G12 diet | $\begin{gathered} \text { G12 } \\ \text { intake } \end{gathered}$ |
| $\begin{aligned} & \text { FP } \\ & 0009 \end{aligned}$ | Pome fruit, raw (incl apple juice, excl cider) | RAC | 0.04 | 57.68 | 2.31 | 74.45 | 2.98 | 37.84 | 1.51 | 58.40 | 2.34 | 103.51 | 4.14 | 11.20 | 0.45 |
| $\begin{aligned} & \text { JF } \\ & 0226 \end{aligned}$ | Apple juice, single strength (incl concentrated) | PP | 0.0004 | 14.88 | 0.01 | 11.98 | 0.00 | 0.15 | 0.00 | 9.98 | 0.00 | 30.32 | 0.01 | 3.47 | 0.00 |
| $\begin{aligned} & \text { FS } \\ & 0013 \end{aligned}$ | Cherries, raw | RAC | 1 | 1.40 | 1.40 | 4.21 | 4.21 | 0.10 | 0.10 | 2.93 | 2.93 | 1.50 | 1.50 | NC | - |
| $\begin{aligned} & \text { FS } \\ & 0014 \end{aligned}$ | Plums, raw (incl dried plums, incl Chinese jujube) | RAC | 1 | 5.55 | 5.55 | 4.37 | 4.37 | 6.08 | 6.08 | 3.66 | 3.66 | 3.93 | 3.93 | 0.46 | 0.46 |
| $\begin{aligned} & \text { DF } \\ & 0014 \end{aligned}$ | Plum, dried (prunes) | PP | 2 | 0.61 | 1.22 | 0.35 | 0.70 | 0.10 | 0.20 | 0.35 | 0.70 | 0.49 | 0.98 | 0.13 | 0.26 |
| - | Peaches and nectarines, raw | RAC | 0.2 | 8.76 | 1.75 | 12.98 | 2.60 | 8.23 | 1.65 | 10.09 | 2.02 | 3.64 | 0.73 | 0.10 | 0.02 |
| $\begin{aligned} & \text { FB } \\ & 0264 \end{aligned}$ | Blackberries, raw | RAC | 0.1 | 0.10 | 0.01 | 0.52 | 0.05 | 0.14 | 0.01 | 0.24 | 0.02 | NC | - | 0.10 | 0.01 |
| $\begin{aligned} & \text { FB } \\ & 0266 \end{aligned}$ | Dewberries, incl boysen- \& loganberry, raw | RAC | 0.1 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 |
| $\begin{aligned} & \text { FB } \\ & 0272 \end{aligned}$ | Raspberries, red, black, raw | RAC | 0.2 | 0.47 | 0.09 | 0.91 | 0.18 | 0.10 | 0.02 | 0.99 | 0.20 | 1.14 | 0.23 | NC | - |
| $\begin{aligned} & \text { FB } \\ & 0021 \end{aligned}$ | Currants, red, black, white, raw | RAC | 0.2 | 0.48 | 0.10 | 4.23 | 0.85 | NC | - | 1.51 | 0.30 | 0.49 | 0.10 | NC | - |
| $\begin{aligned} & \text { FB } \\ & 0265 \end{aligned}$ | Cranberries, raw | RAC | 0.05 | 0.10 | 0.01 | 0.10 | 0.01 | 0.10 | 0.01 | 1.22 | 0.06 | 0.11 | 0.01 | NC | - |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry, raw | RAC | 0.1 | 4.49 | 0.45 | 5.66 | 0.57 | 0.10 | 0.01 | 6.63 | 0.66 | 5.75 | 0.58 | 0.10 | 0.01 |
| FI 0353 | Pineapple, raw (incl canned pineapple, incl pineapple juice, incl dried pineapple) | RAC | 0.1 | 13.13 | 1.31 | 11.13 | 1.11 | 6.94 | 0.69 | 14.36 | 1.44 | 36.74 | 3.67 | 18.81 | 1.88 |
| FI 0341 | Kiwi fruit, raw | RAC | 0.2 | 2.46 | 0.49 | 3.62 | 0.72 | 0.10 | 0.02 | 1.48 | 0.30 | 7.43 | 1.49 | 0.10 | 0.02 |
| - | Onions, mature bulbs, dry | RAC | 0.05 | 19.69 | 0.98 | 29.83 | 1.49 | 24.64 | 1.23 | 31.35 | 1.57 | 9.72 | 0.49 | 12.59 | 0.63 |
| - | Onions, green, raw | RAC | 1 | 1.55 | 1.55 | 0.74 | 0.74 | 1.05 | 1.05 | 3.74 | 3.74 | 0.94 | 0.94 | 6.45 | 6.45 |

$\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | Expr as | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G10 diet | $\begin{aligned} & \text { G10 } \\ & \text { intake } \end{aligned}$ | G11 diet | $\begin{aligned} & \text { G11 } \\ & \text { intake } \end{aligned}$ | G12 <br> diet | $\begin{gathered} \text { G12 } \\ \text { intake } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G07 diet | $\begin{aligned} & \text { G07 } \\ & \text { intake } \end{aligned}$ | G08 diet | G08 <br> intake | G09 diet | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ |  |  |  |  |  |  |
| $\begin{aligned} & \text { VB } \\ & 0041 \end{aligned}$ | Cabbages, head, raw | RAC | 0.01 | 8.97 | 0.09 | 27.12 | 0.27 | 1.44 | 0.01 | 24.96 | 0.25 | 4.55 | 0.05 | 11.23 | 0.11 |
| $\begin{aligned} & \text { VB } \\ & 0400 \end{aligned}$ | Broccoli, raw | RAC | 0.5 | 4.24 | 2.12 | 1.76 | 0.88 | NC | - | 0.51 | 0.26 | 3.79 | 1.90 | 0.26 | 0.13 |
| $\begin{aligned} & \text { VB } \\ & 0405 \end{aligned}$ | Kohlrabi, raw | RAC | 0.2 | NC | - | 3.25 | 0.65 | NC | - | NC | - | 0.10 | 0.02 | 0.36 | 0.07 |
| $\begin{aligned} & \text { VC } \\ & 0046 \end{aligned}$ | Melons, raw (excl watermelons) | RAC | 0.2 | 9.20 | 1.84 | 11.95 | 2.39 | 14.63 | 2.93 | 8.99 | 1.80 | 7.86 | 1.57 | 2.46 | 0.49 |
| $\begin{aligned} & \text { VC } \\ & 0424 \end{aligned}$ | Cucumber, raw | RAC | 0.1 | 6.72 | 0.67 | 11.03 | 1.10 | 32.10 | 3.21 | 15.10 | 1.51 | 4.05 | 0.41 | 9.57 | 0.96 |
| $\begin{aligned} & \text { VC } \\ & 0431 \end{aligned}$ | Squash, summer, raw (= courgette, zucchini) | RAC | 0.05 | NC | - | NC | - | 5.48 | 0.27 | NC | - | NC | - | 1.03 | 0.05 |
| $\begin{aligned} & \text { VO } \\ & 0445 \end{aligned}$ | Peppers, sweet, raw (incl dried) | RAC | 0.05 | 0.82 | 0.04 | 1.53 | 0.08 | 10.85 | 0.54 | 4.59 | 0.23 | 1.84 | 0.09 | 2.00 | 0.10 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.02 | 11.43 | 0.23 | 3.71 | 0.07 | 0.74 | 0.01 | 13.63 | 0.27 | 3.07 | 0.06 | 1.50 | 0.03 |
| $\begin{aligned} & \text { VO } \\ & 0448 \end{aligned}$ | Tomato, raw (incl juice, incl paste, incl canned) | RAC | 0.12 | 64.74 | 7.77 | 68.31 | 8.20 | 36.05 | 4.33 | 82.09 | 9.85 | 54.50 | 6.54 | 11.69 | 1.40 |
| $\begin{aligned} & \text { VL } \\ & 0480 \end{aligned}$ | Kale, raw (i.e. collards) (i.e. Brassica) | RAC | 0.05 | NC | - | NC | - | 14.54 | 0.73 | NC | - | NC | - | 2.32 | 0.12 |
| $\begin{aligned} & \text { VL } \\ & 0482 \end{aligned}$ | Lettuce, head, raw | RAC | 0.5 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf, raw | RAC | 0.5 | 14.50 | 7.25 | 11.76 | 5.88 | 13.14 | 6.57 | 19.50 | 9.75 | 4.81 | 2.41 | 2.23 | 1.12 |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach, raw | RAC | 0.5 | 2.20 | 1.10 | 1.76 | 0.88 | 13.38 | 6.69 | 2.94 | 1.47 | 5.53 | 2.77 | 0.10 | 0.05 |
| - | Chinese cabbage flowering stalk, raw | RAC | 0.05 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP | Beans, green, without pods, raw: beans | RAC | 0.2 | 2.21 | 0.44 | 5.25 | 1.05 | 4.17 | 0.83 | 1.61 | 0.32 | 16.95 | 3.39 | 0.17 | 0.03 |

Annex 3



Annex 3

| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |
|  |  |  |  | G13 <br> diet | $\begin{gathered} \text { G13 } \\ \text { intake } \end{gathered}$ | G14 <br> diet | G14 intake | $\begin{aligned} & \text { G15 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G15 } \\ \text { intake } \end{gathered}$ | G16 <br> diet | G16 <br> intake | G17 diet | G17 <br> intake |
| FP 0009 | Pome fruit, raw (incl apple juice, excl cider) | RAC | 0.04 | 2.43 | 0.10 | 11.06 | 0.44 | 79.27 | 3.17 | 1.64 | 0.07 | 19.56 | 0.78 |
| JF 0226 | Apple juice, single strength (incl concentrated) | PP | 0.0004 | 0.10 | 0.00 | 0.10 | 0.00 | 7.19 | 0.00 | 0.10 | 0.00 | NC | - |
| FS 0013 | Cherries, raw | RAC | 1 | 0.10 | 0.10 | 0.10 | 0.10 | 5.96 | 5.96 | 0.10 | 0.10 | NC | - |
| FS 0014 | Plums, raw (incl dried plums, incl Chinese jujube) | RAC | 1 | 0.10 | 0.10 | 0.10 | 0.10 | 16.65 | 16.65 | 0.10 | 0.10 | NC | - |
| DF 0014 | Plum, dried (prunes) | PP | 2 | 0.10 | 0.20 | 0.10 | 0.20 | 0.37 | 0.74 | 0.10 | 0.20 | NC | - |
| - | Peaches and nectarines, raw | RAC | 0.2 | 0.10 | 0.02 | 0.10 | 0.02 | 7.47 | 1.49 | 0.10 | 0.02 | NC | - |
| FB 0264 | Blackberries, raw | RAC | 0.1 | 0.10 | 0.01 | 7.29 | 0.73 | 0.25 | 0.03 | 0.10 | 0.01 | NC | - |
| FB 0266 | Dewberries, incl boysen- \& loganberry, raw | RAC | 0.1 | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | NC | - |
| FB 0272 | Raspberries, red, black, raw | RAC | 0.2 | 0.10 | 0.02 | 0.10 | 0.02 | 2.04 | 0.41 | 0.10 | 0.02 | NC | - |
| FB 0021 | Currants, red, black, white, raw | RAC | 0.2 | 0.10 | 0.02 | NC | - | 0.74 | 0.15 | NC | - | NC | - |
| FB 0265 | Cranberries, raw | RAC | 0.05 | NC | - | NC | - | 0.10 | 0.01 | NC | - | NC | - |
| FB 0275 | Strawberry, raw | RAC | 0.1 | 0.10 | 0.01 | 0.10 | 0.01 | 3.35 | 0.34 | 0.10 | 0.01 | 0.10 | 0.01 |
| FI 0353 | Pineapple, raw (incl canned pineapple, incl pineapple juice, incl dried pineapple) | RAC | 0.1 | 8.51 | 0.85 | 6.27 | 0.63 | 6.89 | 0.69 | 0.18 | 0.02 | 24.94 | 2.49 |
| FI 0341 | Kiwi fruit, raw | RAC | 0.2 | 0.10 | 0.02 | 0.10 | 0.02 | 2.00 | 0.40 | 0.10 | 0.02 | NC | - |
| - | Onions, mature bulbs, dry | RAC | 0.05 | 9.01 | 0.45 | 20.24 | 1.01 | 30.90 | 1.55 | 9.61 | 0.48 | 2.11 | 0.11 |
| - | Onions, green, raw | RAC | 1 | 1.43 | 1.43 | 0.10 | 0.10 | 0.20 | 0.20 | NC | - | 6.30 | 6.30 |
| VB 0041 | Cabbages, head, raw | RAC | 0.01 | 3.82 | 0.04 | 2.99 | 0.03 | 49.16 | 0.49 | 0.10 | 0.00 | NC | - |
| VB 0400 | Broccoli, raw | RAC | 0.5 | 0.10 | 0.05 | 0.10 | 0.05 | 2.13 | 1.07 | 0.10 | 0.05 | NC | - |
| VB 0405 | Kohlrabi, raw | RAC | 0.2 | 0.12 | 0.02 | 0.10 | 0.02 | 1.81 | 0.36 | 0.10 | 0.02 | NC | - |
| VC 0046 | Melons, raw (excl watermelons) | RAC | 0.2 | 0.19 | 0.04 | 0.10 | 0.02 | 4.98 | 1.00 | 0.10 | 0.02 | NC | - |
| VC 0424 | Cucumber, raw | RAC | 0.1 | 0.68 | 0.07 | 1.81 | 0.18 | 10.40 | 1.04 | 0.10 | 0.01 | 0.10 | 0.01 |


| DIAZINON (22) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.003 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | $\begin{gathered} \text { STMR } \\ \mathrm{mg} / \mathrm{kg} \end{gathered}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |
|  |  |  |  | G13 <br> diet | G13 <br> intake | G14 <br> diet | G14 <br> intake | G15 <br> diet | G15 <br> intake | G16 <br> diet | G16 <br> intake | G17 <br> diet | G17 <br> intake |
| VC 0431 | Squash, summer, raw (= courgette, zucchini) | RAC | 0.05 | 0.10 | 0.01 | 1.01 | 0.05 | NC | - | 1.91 | 0.10 | NC | - |
| VO 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.05 | 5.49 | 0.27 | 10.57 | 0.53 | 8.84 | 0.44 | 0.91 | 0.05 | NC | - |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.02 | 3.63 | 0.07 | 20.50 | 0.41 | 8.78 | 0.18 | 0.10 | 0.00 | 0.17 | 0.00 |
| VO 0448 | Tomato, raw (incl juice, incl paste, incl canned) | RAC | 0.12 | 15.50 | 1.86 | 5.78 | 0.69 | 71.52 | 8.58 | 2.00 | 0.24 | 12.50 | 1.50 |
| VL 0480 | Kale, raw (i.e. collards) (i.e. Brassica) | RAC | 0.05 | 0.79 | 0.04 | 0.62 | 0.03 | NC | - | 0.10 | 0.01 | NC | - |
| VL 0482 | Lettuce, head, raw | RAC | 0.5 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VL 0483 | Lettuce, leaf, raw | RAC | 0.5 | 0.29 | 0.15 | 0.10 | 0.05 | 6.71 | 3.36 | 0.10 | 0.05 | NC | - |
| VL 0502 | Spinach, raw | RAC | 0.5 | 0.17 | 0.09 | 0.10 | 0.05 | 0.81 | 0.41 | 0.10 | 0.05 | NC | - |
| - | Chinese cabbage flowering stalk, raw | RAC | 0.05 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP 0062 | Beans, green, without pods, raw: beans except broad bean \& soya bean (i.e. immature seeds only) (Phaseolus spp) | RAC | 0.2 | 0.30 | 0.06 | 3.13 | 0.63 | 4.11 | 0.82 | 0.10 | 0.02 | NC | - |
| VP 0064 | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) | RAC | 0.2 | 0.21 | 0.04 | 0.10 | 0.02 | 5.51 | 1.10 | 0.10 | 0.02 | NC | - |
| VR 0494 | Radish roots, raw | RAC | 0.1 | 3.96 | 0.40 | 2.86 | 0.29 | 3.30 | 0.33 | 2.67 | 0.27 | 5.34 | 0.53 |
| VR 0577 | Carrots, raw | RAC | 0.5 | 2.07 | 1.04 | 3.00 | 1.50 | 25.29 | 12.65 | 0.10 | 0.05 | NC | - |
| VR 0589 | Potato, raw (incl flour, incl frozen, incl starch, incl tapioca) | RAC | 0.01 | 23.96 | 0.24 | 13.56 | 0.14 | 213.41 | 2.13 | 104.35 | 1.04 | 8.56 | 0.09 |
| VR 0589 | Potato, raw (incl flour, incl frozen, incl tapioca, excl starch) | RAC | 0 | 23.96 | 0.00 | 13.54 | 0.00 | 213.41 | 0.00 | 104.35 | 0.00 | 8.56 | 0.00 |
| VR 0596 | Sugar beet, raw (incl sugar) | RAC | 0.1 | 3.93 | 0.39 | 1.68 | 0.17 | NC | - | NC | - | 36.12 | 3.61 |
| GC 0645 | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0 | 116.66 | 0.00 | 10.52 | 0.00 | 38.46 | 0.00 | 76.60 | 0.00 | 34.44 | 0.00 |
| TN 0660 | Almonds, nutmeat | RAC | 0.05 | 0.10 | 0.01 | 0.10 | 0.01 | 0.61 | 0.03 | 0.10 | 0.01 | NC | - |



## GLYPHOSATE

| GLYPHOSATE (158) |  |  | International Estimated Daily Intake (IEDI) |  |  |  | ADI $=0-1 \mathrm{mg} / \mathrm{kg}$ bw |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | G01 <br> diet | $\begin{aligned} & \text { G01 } \\ & \text { intake } \end{aligned}$ | G02 <br> diet | $\begin{aligned} & \text { G02 } \\ & \text { intake } \end{aligned}$ | $\begin{aligned} & \text { G03 } \\ & \text { diet } \end{aligned}$ | $\begin{aligned} & \text { G03 } \\ & \text { intake } \end{aligned}$ | G04 <br> diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | G05 <br> diet | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | G06 <br> diet | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
| FI 0327 | Banana, raw (incl plantains) (incl dried) | RAC | 0.05 | 5.06 | 0.25 | 6.91 | 0.35 | 37.17 | 1.86 | 31.16 | 1.56 | 40.21 | 2.01 | 18.96 | 0.95 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.325 | 0.14 | 0.05 | 0.94 | 0.31 | 5.70 | 1.85 | 2.61 | 0.85 | 1.94 | 0.63 | 0.22 | 0.07 |
| $\begin{aligned} & \text { VD } \\ & 0071 \end{aligned}$ | Beans, dry, raw (Phaseolus spp) | RAC | 0.17 | 2.39 | 0.41 | 1.61 | 0.27 | 10.47 | 1.78 | 1.84 | 0.31 | 12.90 | 2.19 | 7.44 | 1.26 |
| $\begin{aligned} & \text { VD } \\ & 0072 \end{aligned}$ | Peas, dry, raw (Pisum spp, Vigna spp): garden peas \& field peas \& cow peas | RAC | 0.5 | 1.67 | 0.84 | 3.22 | 1.61 | 2.66 | 1.33 | 1.51 | 0.76 | 2.91 | 1.46 | 0.24 | 0.12 |
| $\begin{aligned} & \text { VD } \\ & 0533 \end{aligned}$ | Lentil, dry, raw (Ervum lens) | RAC | 0.5 | 2.12 | 1.06 | 0.10 | 0.05 | 0.10 | 0.05 | 3.21 | 1.61 | 1.60 | 0.80 | 4.90 | 2.45 |
| $\begin{aligned} & \text { VD } \\ & 0541 \end{aligned}$ | Soya bean, dry, raw (incl flour, incl paste, incl curd, incl sauce, excl oil) | RAC | 5 | 0.63 | 3.15 | 1.09 | 5.45 | 0.40 | 2.00 | 1.40 | 7.00 | 1.68 | 8.40 | 0.48 | 2.40 |
| $\begin{aligned} & \text { OR } \\ & 0541 \end{aligned}$ | Soya oil, refined | PP | 0.1 | 12.99 | 1.30 | 10.43 | 1.04 | 3.63 | 0.36 | 13.10 | 1.31 | 10.70 | 1.07 | 13.10 | 1.31 |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet, raw | RAC | 3.4 | NC | - | NC | - | NC | - | NC | - | 0.10 | 0.34 | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0640 \end{aligned}$ | Barley, raw (incl malt extract, incl pot \& pearled, incl flour \& grits, incl beer, incl malt) | RAC | 3.7 | 19.91 | 73.67 | 31.16 | 115.29 | 5.04 | 18.65 | 3.10 | 11.47 | 9.77 | 36.15 | 4.31 | 15.95 |
| $\begin{aligned} & \mathrm{GC} \\ & 0641 \end{aligned}$ | Buckwheat, raw (incl flour) | RAC | 3.7 | NC | - | 0.40 | 1.48 | 0.10 | 0.37 | 0.10 | 0.37 | 0.10 | 0.37 | 0.10 | 0.37 |
| $\begin{aligned} & \mathrm{GC} \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl beer, incl germ, excl flour, excl oil, excl starch) | RAC | 0.12 | 0.84 | 0.10 | 0.24 | 0.03 | 1.56 | 0.19 | 0.46 | 0.06 | 2.44 | 0.29 | 13.13 | 1.58 |
| $\begin{aligned} & \mathrm{GC} \\ & 0656 \end{aligned}$ | Popcorn (i.e. maize used for preparation of popcorn) | RAC | 0.12 | - | - | - | - | - | - | - | - | - | - | - | - |
| CF 1255 | Maize, flour (white flour and wholemeal | PP | 0.13 | 22.72 | 2.95 | 35.61 | 4.63 | 87.27 | 11.35 | 34.92 | 4.54 | 46.71 | 6.07 | 49.12 | 6.39 |


$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G04 <br> diet | G04 <br> intake | $\underset{\text { Giet }}{\mathrm{G} 05}$ | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | $\begin{aligned} & \text { G06 } \\ & \text { diet } \end{aligned}$ | $\begin{aligned} & \text { G06 } \\ & \text { intake } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G01 <br> diet | $\begin{aligned} & \text { G01 } \\ & \text { intake } \end{aligned}$ | G02 <br> diet | $\begin{aligned} & \text { G02 } \\ & \text { intake } \end{aligned}$ | $\begin{aligned} & \text { G03 } \\ & \text { diet } \end{aligned}$ | $\begin{aligned} & \text { G03 } \\ & \text { intake } \end{aligned}$ |  |  |  |  |  |  |
| - | Fonio, raw (incl flour) | RAC | 3.7 | NC | - | NC | - | 1.01 | 3.74 | NC | - | NC | - | NC | - |
| - | Cereals, NES, raw (including processed) : canagua, quihuicha, Job's tears and wild rice | RAC | 3.7 | 2.04 | 7.55 | 2.99 | 11.06 | 1.86 | 6.88 | 19.17 | 70.93 | 3.33 | 12.32 | 1.66 | 6.14 |
| GS 0659 | Sugar cane, raw | RAC | 0.27 | 38.16 | 10.30 | NC | - | 12.58 | 3.40 | 0.34 | 0.09 | 17.79 | 4.80 | 42.78 | 11.55 |
| - | Sugar cane, molasses | PP | 2.3 | NC | - | NC | - | NC | - | NC | - | 0.10 | 0.23 | NC | - |
| - | Sugar cane, sugar (incl non-centrifugal sugar, incl refined sugar and maltose) | PP | 0.065 | 61.52 | 4.00 | 86.27 | 5.61 | 18.80 | 1.22 | 80.02 | 5.20 | 66.39 | 4.32 | 56.32 | 3.66 |
| SO 0495 | Rape seed, raw | RAC | 3 | 0.10 | 0.30 | NC | - | NC | - | 0.10 | 0.30 | 0.75 | 2.25 | 0.10 | 0.30 |
| OR $0495$ | Rape seed oil, edible | PP | 0.009 | 0.35 | 0.00 | 0.44 | 0.00 | 0.19 | 0.00 | 0.97 | 0.01 | 3.28 | 0.03 | 0.77 | 0.01 |
| SO 0691 | Cotton seed, raw | RAC | 5.2 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |
| $\begin{aligned} & \text { OR } \\ & 0691 \end{aligned}$ | Cotton seed oil, edible | PP | 0.52 | 3.22 | 1.67 | 1.54 | 0.80 | 1.01 | 0.53 | 0.74 | 0.38 | 1.12 | 0.58 | 2.93 | 1.52 |
| SO 0702 | Sunflower seed, raw (incl oil) | RAC | 0.395 | 7.40 | 2.92 | 35.86 | 14.16 | 1.15 | 0.45 | 8.76 | 3.46 | 5.45 | 2.15 | 13.62 | 5.38 |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $-80 \%$ as muscle | RAC | 0.05 | 24.96 | 1.25 | 57.95 | 2.90 | 16.70 | 0.84 | 38.38 | 1.92 | 26.46 | 1.32 | 29.00 | 1.45 |
| MM $0095$ | MEAT FROM MAMMALS other than marine mammals, raw (incl prepared meat) $-20 \%$ as fat | RAC | 0.05 | 6.24 | 0.31 | 14.49 | 0.72 | 4.18 | 0.21 | 9.60 | 0.48 | 6.62 | 0.33 | 7.25 | 0.36 |
| MO <br> 0105 | Edible offal (mammalian), raw | RAC | 2.9 | 4.79 | 13.89 | 9.68 | 28.07 | 2.97 | 8.61 | 5.49 | 15.92 | 3.84 | 11.14 | 5.03 | 14.59 |
| $\begin{aligned} & \text { ML } \\ & 0106 \end{aligned}$ | Milks, raw or skimmed (incl dairy products) | RAC | 0 | 289.65 | 0.00 | 485.88 | 0.00 | 26.92 | 0.00 | 239.03 | 0.00 | 199.91 | 0.00 | 180.53 | 0.00 |
| PM <br> 0110 | Poultry meat, raw (incl prepared) - $90 \%$ as muscle | RAC | 0 | 13.17 | 0.00 | 26.78 | 0.00 | 7.24 | 0.00 | 116.71 | 0.00 | 22.54 | 0.00 | 32.09 | 0.00 |
| PM | Poultry meat, raw (incl prepared) - 10\% | RAC | 0 | 1.46 | 0.00 | 2.98 | 0.00 | 0.80 | 0.00 | 12.97 | 0.00 | 2.50 | 0.00 | 3.57 | 0.00 |

Annex 3

GLYPHOSATE (158)
International Estimated Daily Intake (IEDI)
$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | Expras | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G01 <br> diet | $\begin{aligned} & \text { G01 } \\ & \text { intake } \end{aligned}$ | G02 <br> diet | $\begin{gathered} \text { G02 } \\ \text { intake } \end{gathered}$ | $\begin{aligned} & \text { G03 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G03 } \\ \text { intake } \end{gathered}$ | G04 <br> diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | $\begin{aligned} & \text { G05 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | G06 diet | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
| 0110 | as fat |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| PO 0111 | Poultry edible offal, raw (incl prepared) | RAC | 0.088 | 0.12 | 0.01 | 0.12 | 0.01 | 0.11 | 0.01 | 5.37 | 0.47 | 0.24 | 0.02 | 0.10 | 0.01 |
| PE 0112 | Eggs, raw (incl dried) | RAC | 0 | 7.84 | 0.00 | 23.08 | 0.00 | 2.88 | 0.00 | 14.89 | 0.00 | 9.81 | 0.00 | 14.83 | 0.00 |
| Total intake $(\mu \mathrm{g} /$ person $)=$ |  |  |  | 171.9 |  |  | 335.1 |  | 121.8 |  | 187.3 |  | 197.7 |  | 129.0 |
|  | Body weight per region (kg bw) = |  |  | 60 |  |  | 60 |  | 60 |  | 60 |  | 60 |  | 60 |
|  | $\operatorname{ADI}(\mu \mathrm{g} /$ person $)=$ |  |  |  | 60000 |  | 60000 |  | 60000 |  | 60000 |  | 60000 |  | 60000 |
|  | \% $\mathrm{ADI}=$ |  |  |  | 0.3\% |  | 0.6\% |  | 0.2\% |  | 0.3\% |  | 0.3\% |  | 0.2\% |
|  | Rounded \%ADI = |  |  |  | 0\% |  | 1\% |  | 0\% |  | 0\% |  | 0\% |  | 0\% |


| GLYPHOSATE (158) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | $\begin{gathered} \text { Expr } \\ \text { as } \end{gathered}$ | STMR $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} / \mathrm{person} /$ day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{gathered} \text { G07 } \\ \text { diet } \end{gathered}$ | $\begin{aligned} & \text { G07 } \\ & \text { intake } \end{aligned}$ | G08 diet | G08 <br> intake | G09 <br> diet | G09 <br> intake | G10 diet | G10 <br> intake | G11 <br> diet | G11 <br> intake | G12 <br> diet | $\begin{gathered} \text { G12 } \\ \text { intake } \end{gathered}$ |
| $\begin{aligned} & \text { FI } \\ & 0327 \end{aligned}$ | Banana, raw (incl plantains) (incl dried) | RAC | 0.05 | 25.14 | 1.26 | 23.37 | 1.17 | 23.06 | 1.15 | 23.40 | 1.17 | 18.44 | 0.92 | 39.29 | 1.96 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.325 | 11.43 | 3.71 | 3.71 | 1.21 | 0.74 | 0.24 | 13.63 | 4.43 | 3.07 | 1.00 | 1.50 | 0.49 |
| $\begin{aligned} & \text { VD } \\ & 0071 \end{aligned}$ | Beans, dry, raw (Phaseolus spp) | RAC | 0.17 | 1.51 | 0.26 | 1.50 | 0.26 | 1.90 | 0.32 | 5.11 | 0.87 | 1.36 | 0.23 | 23.43 | 3.98 |
| $\begin{aligned} & \text { VD } \\ & 0072 \end{aligned}$ | Peas, dry, raw (Pisum spp, Vigna spp): garden peas \& field peas \& cow peas | RAC | 0.5 | 3.80 | 1.90 | 1.25 | 0.63 | 1.06 | 0.53 | 2.33 | 1.17 | 2.70 | 1.35 | 3.83 | 1.92 |
| $\begin{aligned} & \text { VD } \\ & 0533 \end{aligned}$ | Lentil, dry, raw (Ervum lens) | RAC | 0.5 | 0.95 | 0.48 | 1.18 | 0.59 | 0.40 | 0.20 | 0.96 | 0.48 | 0.71 | 0.36 | 1.28 | 0.64 |
| $\begin{aligned} & \text { VD } \\ & 0541 \end{aligned}$ | Soya bean, dry, raw (incl flour, incl paste, incl curd, incl sauce, excl oil) | RAC | 5 | 0.47 | 2.35 | 0.77 | 3.85 | 9.12 | 45.60 | 8.05 | 40.25 | 0.10 | 0.50 | 6.06 | 30.30 |
| $\begin{aligned} & \text { OR } \\ & 0541 \end{aligned}$ | Soya oil, refined | PP | 0.1 | 19.06 | 1.91 | 21.06 | 2.11 | 5.94 | 0.59 | 33.78 | 3.38 | 40.05 | 4.01 | 13.39 | 1.34 |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet, raw | RAC | 3.4 | 0.10 | 0.34 | NC | - | 0.10 | 0.34 | 0.10 | 0.34 | NC | - | NC | - |
| $\begin{aligned} & \mathrm{GC} \\ & 0640 \end{aligned}$ | Barley, raw (incl malt extract, incl pot \& pearled, incl flour \& grits, incl beer, incl malt) | RAC | 3.7 | 36.18 | 133.87 | 53.45 | 197.77 | 9.39 | 34.74 | 35.25 | 130.43 | 46.68 | 172.72 | 15.92 | 58.90 |
| $\begin{aligned} & \mathrm{GC} \\ & 0641 \end{aligned}$ | Buckwheat, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | 0.79 | 2.92 | 0.18 | 0.67 | 0.35 | 1.30 | NC | - | NC | - |
| $\begin{aligned} & \text { GC } \\ & 0645 \end{aligned}$ | Maize, raw (incl glucose \& dextrose \& isoglucose, incl beer, incl germ, excl flour, excl oil, excl starch) | RAC | 0.12 | 0.10 | 0.01 | 9.93 | 1.19 | 1.40 | 0.17 | 10.26 | 1.23 | 0.33 | 0.04 | 0.10 | 0.01 |
| $\begin{aligned} & \text { GC } \\ & 0656 \end{aligned}$ | Popcorn (i.e. maize used for preparation of popcorn) | RAC | 0.12 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { CF } \\ & 1255 \end{aligned}$ | Maize, flour (white flour and wholemeal flour) | PP | 0.13 | 14.27 | 1.86 | 12.86 | 1.67 | 19.71 | 2.56 | 12.55 | 1.63 | 4.21 | 0.55 | 52.30 | 6.80 |


$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw

$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw

$\mathrm{ADI}=0-1 \mathrm{mg} / \mathrm{kg}$ bw

| Codex code | Commodity description | Expr as | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | G13 <br> diet | $\begin{gathered} \text { G13 } \\ \text { intake } \end{gathered}$ | G14 <br> diet | $\begin{gathered} \text { G14 } \\ \text { intake } \end{gathered}$ | G15 <br> diet | $\begin{gathered} \text { G15 } \\ \text { intake } \end{gathered}$ | G16 <br> diet | $\begin{gathered} \text { G16 } \\ \text { intake } \end{gathered}$ | $\begin{gathered} \text { G17 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G17 } \\ \text { intake } \end{gathered}$ |
| FI 0327 | Banana, raw (incl plantains) (incl dried) | RAC | 0.05 | 20.88 | 1.04 | 81.15 | 4.06 | 24.58 | 1.23 | 37.92 | 1.90 | 310.23 | 15.51 |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.325 | 3.63 | 1.18 | 20.50 | 6.66 | 8.78 | 2.85 | 0.10 | 0.03 | 0.17 | 0.06 |
| VD 0071 | Beans, dry, raw (Phaseolus spp) | RAC | 0.17 | 7.11 | 1.21 | 2.33 | 0.40 | 3.76 | 0.64 | 44.70 | 7.60 | 3.27 | 0.56 |
| VD 0072 | Peas, dry, raw (Pisum spp, Vigna spp): garden peas \& field peas \& cow peas | RAC | 0.5 | 14.30 | 7.15 | 3.51 | 1.76 | 3.52 | 1.76 | 7.89 | 3.95 | 0.74 | 0.37 |
| VD 0533 | Lentil, dry, raw (Ervum lens) | RAC | 0.5 | 0.67 | 0.34 | 7.26 | 3.63 | 0.37 | 0.19 | 0.10 | 0.05 | NC | - |
| VD 0541 | Soya bean, dry, raw (incl flour, incl paste, incl curd, incl sauce, excl oil) | RAC | 5 | 2.89 | 14.45 | 0.21 | 1.05 | 0.48 | 2.40 | 3.16 | 15.80 | 0.26 | 1.30 |
| OR 0541 | Soya oil, refined | PP | 0.1 | 2.32 | 0.23 | 2.54 | 0.25 | 18.70 | 1.87 | 2.51 | 0.25 | 6.29 | 0.63 |
| VR 0596 | Sugar beet, raw | RAC | 3.4 | 0.10 | 0.34 | NC | - | NC | - | NC | - | NC | - |
| GC 0640 | Barley, raw (incl malt extract, incl pot \& pearled, incl flour \& grits, incl beer, incl malt) | RAC | 3.7 | 11.58 | 42.85 | 2.33 | 8.62 | 46.71 | 172.83 | 3.72 | 13.76 | 16.26 | 60.16 |
| GC 0641 | Buckwheat, raw (incl flour) | RAC | 3.7 | 0.10 | 0.37 | 2.82 | 10.43 | 0.10 | 0.37 | 0.10 | 0.37 | NC | - |
| GC 0645 | Maize, raw (incl glucose \& dextrose \& isoglucose, incl beer, incl germ, excl flour, excl oil, excl starch) | RAC | 0.12 | 0.55 | 0.07 | 0.51 | 0.06 | 3.26 | 0.39 | 7.96 | 0.96 | NC | - |
| GC 0656 | Popcorn (i.e. maize used for preparation of popcorn) | RAC | 0.12 | - | - | - | - | - | - | - | - | - | - |
| CF 1255 | Maize, flour (white flour and wholemeal flour) | PP | 0.13 | 94.34 | 12.26 | 8.09 | 1.05 | 28.03 | 3.64 | 55.94 | 7.27 | 28.07 | 3.65 |
| - | Maize starch | PP | 0.04 | 0.10 | 0.00 | 0.10 | 0.00 | NC | - | NC | - | NC | - |
| OR 0645 | Maize oil | PP | 0.04 | 0.33 | 0.01 | 0.10 | 0.00 | 0.81 | 0.03 | 0.10 | 0.00 | NC | - |
| GC 0646 | Millet, raw (incl flour, incl beer) | RAC | 3.7 | 61.13 | 226.18 | 0.78 | 2.89 | NC | - | 33.55 | 124.14 | NC | - |
| GC 0647 | Oats, raw (incl rolled) | RAC | 3.7 | 0.37 | 1.37 | 0.10 | 0.37 | 2.79 | 10.32 | 0.10 | 0.37 | NC | - |
| GC 0648 | Quinoa, raw | RAC | 3.7 | NC | - | NC | - | NC | - | NC | - | NC | - |




## MALATHION

| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg}$ bw |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} /$ person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | G01 <br> diet | G01 intake | $\begin{gathered} \text { G02 } \\ \text { diet } \end{gathered}$ | $\begin{gathered} \text { G02 } \\ \text { intake } \end{gathered}$ | G03 <br> diet | G03 <br> intake | G04 <br> diet | $\begin{gathered} \text { G04 } \\ \text { intake } \end{gathered}$ | G05 <br> diet | $\begin{gathered} \text { G05 } \\ \text { intake } \end{gathered}$ | G06 <br> diet | $\begin{gathered} \text { G06 } \\ \text { intake } \end{gathered}$ |
| FC 0001 | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 34.91 | 0.70 | 16.51 | 0.33 | 17.23 | 0.34 | 104.48 | 2.09 | 35.57 | 0.71 | 98.49 | 1.97 |
| FP 0226 | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 13.94 | 1.53 | 30.81 | 3.39 | 15.14 | 1.67 | 23.10 | 2.54 | 6.86 | 0.75 | 55.48 | 6.10 |
| FB 0020 | Blueberries, raw | RAC | 2.27 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 | 0.10 | 0.23 |
| FB 0269 | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 16.25 | 2.60 | 28.96 | 4.63 | 2.87 | 0.46 | 24.22 | 3.88 | 9.33 | 1.49 | 68.64 | 10.98 |
| FB 0275 | Strawberry, raw | RAC | 0.25 | 0.70 | 0.18 | 2.01 | 0.50 | 0.10 | 0.03 | 1.36 | 0.34 | 0.37 | 0.09 | 2.53 | 0.63 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 29.36 | 6.75 | 37.50 | 8.63 | 3.56 | 0.82 | 34.78 | 8.00 | 18.81 | 4.33 | 43.38 | 9.98 |
| - | Onions, green, raw | RAC | 0.52 | 2.45 | 1.27 | 1.49 | 0.77 | 1.02 | 0.53 | 2.60 | 1.35 | 0.60 | 0.31 | 2.03 | 1.06 |
| VC 0424 | Cucumber, raw | RAC | 0.02 | 8.01 | 0.16 | 30.66 | 0.61 | 1.45 | 0.03 | 19.84 | 0.40 | 0.27 | 0.01 | 34.92 | 0.70 |
| VO 0444 | Peppers, chili, raw (incl dried) | RAC | 0.01 | 6.93 | 0.07 | 10.97 | 0.11 | 8.83 | 0.09 | 9.13 | 0.09 | 6.65 | 0.07 | 20.01 | 0.20 |
| VO 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 4.49 | 0.04 | 6.44 | 0.06 | 7.21 | 0.07 | 5.68 | 0.06 | 9.52 | 0.10 | 8.92 | 0.09 |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 0.14 | 0.00 | 0.94 | 0.01 | 5.70 | 0.06 | 2.61 | 0.03 | 1.94 | 0.02 | 0.22 | 0.00 |
| VO 0448 | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 42.04 | 8.83 | 76.13 | 15.99 | 10.69 | 2.24 | 84.59 | 17.76 | 24.92 | 5.23 | 203.27 | 42.69 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 2.34 | 0.16 | 1.33 | 0.09 | 1.57 | 0.11 | 4.24 | 0.30 | 0.34 | 0.02 | 2.83 | 0.20 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.29 | 0.00 | 0.29 | 0.00 | 0.10 | 0.00 | 0.38 | 0.00 | 0.10 | 0.00 | 0.14 | 0.00 |
| VL 0485 | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | 0.10 | 0.01 | 0.31 | 0.02 | 0.10 | 0.01 | 0.10 | 0.01 | 0.47 | 0.03 | 0.11 | 0.01 |
| VL 0502 | Spinach, raw | RAC | 0.35 | 0.74 | 0.26 | 0.22 | 0.08 | 0.10 | 0.04 | 0.91 | 0.32 | 0.10 | 0.04 | 2.92 | 1.02 |
| VL 0506 | Turnip greens, raw (i.e. Namenia, | RAC | 1.2 | NC | - | NC | - | NC | - | NC | - | NC | - | NC | - |



## Annex 3

| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | STMR <br> $\mathrm{mg} / \mathrm{kg}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |  |  |
|  |  |  |  | $\begin{aligned} & \text { G07 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G07 } \\ \text { intake } \end{gathered}$ | G08 <br> diet | G08 intake | G09 diet | $\begin{gathered} \text { G09 } \\ \text { intake } \end{gathered}$ | G10 <br> diet | G10 intake | G11 <br> diet | G11 intake | G12 <br> diet | G12 <br> intake |
| $\begin{aligned} & \text { FC } \\ & 0001 \end{aligned}$ | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 114.42 | 2.29 | 62.91 | 1.26 | 26.97 | 0.54 | 96.72 | 1.93 | 96.22 | 1.92 | 563.19 | 11.26 |
| $\begin{aligned} & \text { FP } \\ & 0226 \end{aligned}$ | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 61.44 | 6.76 | 72.81 | 8.01 | 26.84 | 2.95 | 45.18 | 4.97 | 93.28 | 10.26 | 7.78 | 0.86 |
| $\begin{aligned} & \text { FB } \\ & 0020 \end{aligned}$ | Blueberries, raw | RAC | 2.27 | 0.10 | 0.23 | 0.23 | 0.52 | 0.10 | 0.23 | 0.83 | 1.88 | 0.33 | 0.75 | NC | - |
| $\begin{aligned} & \text { FB } \\ & 0269 \end{aligned}$ | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 142.23 | 22.76 | 105.77 | 16.92 | 7.87 | 1.26 | 52.44 | 8.39 | 109.22 | 17.48 | 10.96 | 1.75 |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry, raw | RAC | 0.25 | 4.49 | 1.12 | 5.66 | 1.42 | 0.10 | 0.03 | 6.63 | 1.66 | 5.75 | 1.44 | 0.10 | 0.03 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 19.69 | 4.53 | 29.83 | 6.86 | 24.64 | 5.67 | 31.35 | 7.21 | 9.72 | 2.24 | 12.59 | 2.90 |
| - | Onions, green, raw | RAC | 0.52 | 1.55 | 0.81 | 0.74 | 0.38 | 1.05 | 0.55 | 3.74 | 1.94 | 0.94 | 0.49 | 6.45 | 3.35 |
| $\begin{aligned} & \text { VC } \\ & 0424 \end{aligned}$ | Cucumber, raw | RAC | 0.02 | 6.72 | 0.13 | 11.03 | 0.22 | 32.10 | 0.64 | 15.10 | 0.30 | 4.05 | 0.08 | 9.57 | 0.19 |
| VO <br> 0444 | Peppers, chili, raw (incl dried) | RAC | 0.01 | 6.36 | 0.06 | 15.46 | 0.15 | 10.74 | 0.11 | 7.28 | 0.07 | 8.21 | 0.08 | 3.58 | 0.04 |
| $\begin{aligned} & \text { VO } \\ & 0445 \end{aligned}$ | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 0.82 | 0.01 | 1.53 | 0.02 | 10.85 | 0.11 | 4.59 | 0.05 | 1.84 | 0.02 | 2.00 | 0.02 |
| $\begin{aligned} & \text { VO } \\ & 0447 \end{aligned}$ | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 11.43 | 0.11 | 3.71 | 0.04 | 0.74 | 0.01 | 13.63 | 0.14 | 3.07 | 0.03 | 1.50 | 0.02 |
| $\begin{aligned} & \text { VO } \\ & 0448 \end{aligned}$ | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 43.88 | 9.21 | 55.41 | 11.64 | 35.38 | 7.43 | 74.88 | 15.72 | 26.50 | 5.57 | 9.51 | 2.00 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 4.96 | 0.35 | 3.20 | 0.22 | 0.15 | 0.01 | 1.61 | 0.11 | 6.88 | 0.48 | 0.52 | 0.04 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.80 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 | 0.61 | 0.00 | 0.40 | 0.00 | 0.10 | 0.00 |
| VL | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | NC | - | NC | - | NC | - | NC | - | NC | - | 0.13 | 0.01 |



Annex 3


| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as g/person/day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  |  |  |  |  |
|  |  |  |  | G13 diet | G13 <br> intake | G14 diet | G14 <br> intake | G15 diet | G15 <br> intake | G16 <br> diet | $\begin{aligned} & \text { G16 } \\ & \text { intake } \end{aligned}$ | $\begin{aligned} & \text { G17 } \\ & \text { diet } \end{aligned}$ | $\begin{gathered} \text { G17 } \\ \text { intake } \end{gathered}$ |
| FC 0001 | Citrus fruit, raw (incl citrus fruit juice, incl kumquat commodities) | RAC | 0.02 | 21.16 | 0.42 | 2.94 | 0.06 | 58.52 | 1.17 | 0.44 | 0.01 | 5.13 | 0.10 |
| FP 0226 | Apple, raw (incl juice, incl cider) | RAC | 0.11 | 66.71 | 7.34 | 2.19 | 0.24 | 65.63 | 7.22 | 188.34 | 20.72 | 1.38 | 0.15 |
| FB 0020 | Blueberries, raw | RAC | 2.27 | NC | - | NC | - | 0.20 | 0.45 | NC | - | NC | - |
| FB 0269 | Grape, raw (incl must, incl dried, incl juice, incl wine) | RAC | 0.16 | 0.60 | 0.10 | 1.26 | 0.20 | 103.25 | 16.52 | 0.74 | 0.12 | 44.23 | 7.08 |
| FB 0275 | Strawberry, raw | RAC | 0.25 | 0.10 | 0.03 | 0.10 | 0.03 | 3.35 | 0.84 | 0.10 | 0.03 | 0.10 | 0.03 |
| - | Onions, mature bulbs, dry | RAC | 0.23 | 9.01 | 2.07 | 20.24 | 4.66 | 30.90 | 7.11 | 9.61 | 2.21 | 2.11 | 0.49 |
| - | Onions, green, raw | RAC | 0.52 | 1.43 | 0.74 | 0.10 | 0.05 | 0.20 | 0.10 | NC | - | 6.30 | 3.28 |
| VC 0424 | Cucumber, raw | RAC | 0.02 | 0.68 | 0.01 | 1.81 | 0.04 | 10.40 | 0.21 | 0.10 | 0.00 | 0.10 | 0.00 |
| VO 0444 | Peppers, chili, raw (incl dried) | RAC | 0.01 | 7.55 | 0.08 | 12.48 | 0.12 | 24.78 | 0.25 | 0.87 | 0.01 | NC | - |
| VO 0445 | Peppers, sweet, raw (incl dried) | RAC | 0.01 | 5.49 | 0.05 | 10.57 | 0.11 | 8.84 | 0.09 | 0.91 | 0.01 | NC | - |
| VO 0447 | Sweet corn on the cob, raw (incl frozen, incl canned) (i.e. kernels plus cob without husks) | RAC | 0.01 | 3.63 | 0.04 | 20.50 | 0.21 | 8.78 | 0.09 | 0.10 | 0.00 | 0.17 | 0.00 |
| VO 0448 | Tomato, raw (incl canned, excl juice, excl paste) | RAC | 0.21 | 13.10 | 2.75 | 4.90 | 1.03 | 62.16 | 13.05 | 1.04 | 0.22 | 0.10 | 0.02 |
| - | Tomato, paste (i.e. concentrated tomato sauce/puree) | PP | 0.07 | 0.58 | 0.04 | 0.22 | 0.02 | 2.21 | 0.15 | 0.24 | 0.02 | 3.10 | 0.22 |
| JF 0448 | Tomato, juice (single strength, incl concentrated) | PP | 0 | 0.10 | 0.00 | 0.10 | 0.00 | 0.42 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 |
| VL 0485 | Mustard greens, raw (i.e. Brassica) | RAC | 0.07 | 0.10 | 0.01 | 0.10 | 0.01 | NC | - | 0.10 | 0.01 | NC | - |
| VL 0502 | Spinach, raw | RAC | 0.35 | 0.17 | 0.06 | 0.10 | 0.04 | 0.81 | 0.28 | 0.10 | 0.04 | NC | - |
| VL 0506 | Turnip greens, raw (i.e. Namenia, Tendergreen) | RAC | 1.2 | NC | - | NC | - | NC | - | NC | - | NC | - |
| VP 0061 | Beans, green, with pods, raw: beans except broad bean \& soya bean (i.e. immature seeds + pods) (Phaseolus spp) | RAC | 0.31 | NC | - | NC | - | NC | - | NC | - | NC | - |

# Annex 3 

| MALATHION (49) |  |  | International Estimated Daily Intake (IEDI) |  |  |  |  |  | $\mathrm{ADI}=0-0.3 \mathrm{mg} / \mathrm{kg} \mathrm{bw}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity description | Expr as | $\begin{aligned} & \text { STMR } \\ & \mathrm{mg} / \mathrm{kg} \end{aligned}$ | Diets as $\mathrm{g} / \mathrm{person} /$ day |  |  | Intake as $\mu \mathrm{g} /$ person/day |  |  | G16 diet | $\begin{aligned} & \text { G16 } \\ & \text { intake } \end{aligned}$ | G17 <br> diet | $\begin{gathered} \text { G17 } \\ \text { intake } \end{gathered}$ |
|  |  |  |  | G13 <br> diet | G13 <br> intake | G14 <br> diet | G14 <br> intake | G15 diet | G15 <br> intake |  |  |  |  |
| VD 0071 | Beans, dry, raw (Phaseolus spp) | RAC | 0.36 | 7.11 | 2.56 | 2.33 | 0.84 | 3.76 | 1.35 | 44.70 | 16.09 | 3.27 | 1.18 |
| VR 0506 | Garden turnip, raw | RAC | 0.05 | 4.29 | 0.21 | 3.10 | 0.16 | 6.41 | 0.32 | 2.90 | 0.15 | 5.79 | 0.29 |
| VS 0621 | Asparagus | RAC | 0.305 | 0.10 | 0.03 | 0.10 | 0.03 | 0.17 | 0.05 | 0.10 | 0.03 | NC | - |
| GC 0645 | Maize, raw (incl glucose \& dextrose \& isoglucose, incl flour, incl oil, incl beer, incl germ, incl starch) | RAC | 0.01 | 116.66 | 1.17 | 10.52 | 0.11 | 38.46 | 0.38 | 76.60 | 0.77 | 34.44 | 0.34 |
| GC 0651 | Sorghum, raw (incl flour, incl beer) | RAC | 0.235 | 89.16 | 20.95 | 2.02 | 0.47 | NC | - | 35.38 | 8.31 | NC | - |
| GC 0654 | Wheat, raw (incl bulgur, incl fermented beverages, excl germ, excl wholemeal bread, excl white flour products, excl white bread) | RAC | 10 | 0.10 | 1.00 | 0.10 | 1.00 | 0.10 | 1.00 | 0.10 | 1.00 | 0.97 | 9.70 |
| CF 0654 | Wheat, bran | PP | 25 | NC | - | NC | - | NC | - | NC | - | NC | - |
| CF 1212 | Wheat, wholemeal flour | PP | 7.5 | NC | - | NC | - | NC | - | NC | - | NC | - |
| CP 1212 | Wheat, wholemeal bread | PP | 1.2 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 | 0.10 | 0.12 |
| CP 1211 | Wheat, white bread | PP | 0.2 | 0.43 | 0.09 | 0.41 | 0.08 | 1.56 | 0.31 | 0.11 | 0.02 | 0.10 | 0.02 |
| - | Wheat, gluten | PP | 0.012 | 0.10 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 | 0.10 | 0.00 | 0.19 | 0.00 |
| SO 0691 | Cotton seed, raw | RAC | 4.8 | NC | - | NC | - | NC | - | NC | - | NC | - |
| OR 0691 | Cotton seed oil, edible | PP | 3.12 | 1.28 | 3.99 | 0.10 | 0.31 | 0.45 | 1.40 | 0.42 | 1.31 | 0.15 | 0.47 |
| Total intake $(\mu \mathrm{g} /$ person $)=$ |  |  |  |  | 43.9 |  | 9.9 |  | 52.5 |  | 51.2 |  | 23.5 |
|  | Body weight per region ( kg bw ) $=$ |  |  |  | 60 |  | 60 |  | 60 |  | 60 |  | 60 |
|  | $\operatorname{ADI}(\mu \mathrm{g} /$ person $)=$ |  |  |  | 18000 |  | 18000 |  | 18000 |  | 18000 |  | 18000 |
|  | \%ADI = |  |  |  | 0.2\% |  | 0.1\% |  | 0.3\% |  | 0.3\% |  | 0.1\% |
|  | Rounded \%ADI = |  |  |  | 0\% |  | 0\% |  | 0\% |  | 0\% |  | $0 \%$ |

ANNEX 4: INTERNATIONAL ESTIMATES OF SHORT-TERM DIETARY INTAKES OF PESTICIDE RESIDUES

|  |  |  | DIAZINON (22) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | ARfD $=0.03 \mathrm{mg} / \mathrm{kg} \mathrm{bw}(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$ |  |  |  |  |  |  | Maximum \%ARfD: |  |  |  | $\begin{gathered} 100 \% \\ \text { all } \end{gathered}$ | $\begin{gathered} 100 \% \\ \text { gen pop } \end{gathered}$ | $\begin{aligned} & 100 \% \\ & \text { child } \end{aligned}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Codex code | Commodity | Processing | STMR or STMRP $\mathrm{mg} / \mathrm{kg}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, $\mathrm{g} /$ person | Unit weight, edible portion, g | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | ARfD rounded |  | ARfD rounded |
| FP 0226 | Apple <br> (all commodities) | highest utilization: Total | 0.004 | 0.24 | 1.000 | US | Child, 1-6 yrs | - | 624.45 | 127.0 | 3 | 2a | $\begin{aligned} & 0.04- \\ & 14.05 \end{aligned}$ | 0-50\% | 0-20\% | 0-50\% |
| FP 0227 | Crab-apple (all commodities) | highest utilization: raw with peel | 0 | 0.24 | 1.000 | CN | Gen pop, $>1 \mathrm{yrs}$ | 204 | 488.33 | - | - | - | 0-0 | 0-0\% | 0-0\% | 0-0\% |
| FP 0228 | Loquat (Japanese medlar) (all commodities) | highest utilization: raw without peel | 0 | 0.24 | 1.000 | JP | Gen pop, $>1$ yrs | 113 | 326.40 | 49.0 | 3 | 2a | $\begin{array}{r} 0.42- \\ 1.88 \end{array}$ | 1-6\% | 1-6\% | 0-0\% |
| FP 0229 | Medlar | Total |  | 0.24 | 1.000 | - | - | - | - | - | - | - | - | - | - | - |
| FP 0230 | Pear <br> (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0.004 | 0.24 | 1.000 | CN | Child, 1-6 yrs | 413 | 418.33 | 255.0 | 3 | 2a | 0-13.81 | 0-50\% | 0-20\% | 0-50\% |
| FT 0307 | Persimmon, Japanese (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0 | 0.24 | 1.000 | TH | Child, 3-6 yrs | 20 | 264.88 | 227.5 | 3 | 2a | $\begin{array}{r} 3.53- \\ 10.1 \end{array}$ | $\begin{gathered} 10- \\ 30 \% \end{gathered}$ | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ | $\begin{gathered} 30- \\ 30 \% \end{gathered}$ |
| FP 0231 | Quince (all commodities) | highest utilization: Total | 0 | 0.24 | 1.000 | DE | Child, 2-4 yrs | 16 | 26.30 | 301.2 | 3 | 2b | $\begin{array}{r} 1.17- \\ 1.17 \end{array}$ | 4-4\% | 0-0\% | 4-4\% |
| FS 0013 | Cherries <br> (all commodities) | highest utilization: raw | 0 | 0.73 | 1.000 | DE | Child, 2-4 yrs | 24 | 187.50 | 7.2 | NR | 1 | $\begin{array}{r} 0.83- \\ 8.48 \end{array}$ | 3-30\% | 3-30\% | 8-30\% |
| FS 0014 | Plums <br> (all commodities) | highest utilization: raw with peel (incl consumption without | 0 | $\begin{array}{r} 0.78- \\ 1.9 \end{array}$ | 1.000 | TH | Child, 3-6 yrs | 11 | 376.88 | 93.0 | 3 | 2a | $\begin{aligned} & 3.04- \\ & 25.68 \end{aligned}$ | $\begin{gathered} 10- \\ 90 \% \end{gathered}$ | 4-40\% | $\begin{gathered} 10- \\ 90 \% \end{gathered}$ |

DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$

| Codex code | Commodity | Processing |  |  |  |  |  |  |  |  |  |  |  | all | gen pop | child |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | STMR <br> or STMRP $\mathrm{mg} / \mathrm{kg}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, g/person | Unit weight, edible portion, g | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | ARfD rounded | ARfD rounded | ARfD rounded |
|  |  | peel) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FS 0247 | Peach <br> (all commodities) | highest utilization: raw with peel (incl consumption without peel) | 0 | 0.2 | 1.000 | JP | Child, 1-6 yrs | 76 | 306.00 | 255.0 | 3 | 2a | $\begin{aligned} & 1.09- \\ & 10.53 \end{aligned}$ | 4-40\% | 2-10\% | 4-40\% |
| $\begin{aligned} & \text { FB } \\ & 0264 \end{aligned}$ | Blackberries (all commodities) | highest utilization: raw with skin | 0 | 0.1 | 1.000 | DE | Gen pop, $14-80 \mathrm{yrs}$ | 35 | 460.00 | 2.4 | NR | 1 | 0.02-0.6 | 0-2\% | 0-2\% | 0-2\% |
| $\begin{aligned} & \text { FB } \\ & 0266 \end{aligned}$ | Dewberries, incl boysen- \& loganberry | Total |  | 0.1 | 1.000 | AU | Child, 2-6 yrs | 328 | 3.23 | $<25$ | NR | 1 | 0.02 | 0\% | - | 0\% |
| $\begin{aligned} & \text { FB } \\ & 0272 \end{aligned}$ | Raspberries, red, black <br> (all commodities) | highest utilization: Total | 0 | 0.2 | 1.000 | FR | Child, 3-6 <br> yrs | 0 | 157.50 | 4.3 | NR | 1 | $\begin{array}{r} 0.07- \\ 1.67 \end{array}$ | 0-6\% | 1-3\% | 0-6\% |
| $\begin{aligned} & \text { FB } \\ & 0021 \end{aligned}$ | Currants, red, black, white (all commodities) | highest utilization: Total | 0 | 0.21 | 1.000 | AU | $\begin{aligned} & \text { Gen pop, } \\ & >2 \text { yrs } \end{aligned}$ | 322 | 797.60 | 14.9 | NR | 1 | 0.14-2.5 | 0-8\% | 0-8\% | 0-7\% |
| $\begin{aligned} & \text { FB } \\ & 0265 \end{aligned}$ | Cranberry <br> (all commodities) | highest utilization: Total | 0 | 0.13 | 1.000 | AU | Child, 2- <br> 16 yrs | 103 | 279.66 | 1.8 | NR | 1 | $\begin{array}{r} 0.08- \\ 0.96 \end{array}$ | 0-3\% | 0-2\% | 3-3\% |
| $\begin{aligned} & \text { FB } \\ & 0275 \end{aligned}$ | Strawberry (all commodities) | highest utilization: Total | 0 | 0.12 | 1.000 | FR | Child, 3-6 yrs | 0 | 339.40 | 13.4 | NR | 1 | $\begin{array}{r} 0.14- \\ 2.15 \end{array}$ | 0-7\% | 0-4\% | 0-7\% |
| FI 0353 | Pineapple (all commodities) | highest utilization: raw without peel | 0 | $\begin{array}{r} 0.07- \\ 0.2 \end{array}$ | 1.000 | JP | Child, 1-6 yrs | 67 | 499.80 | 1116.0 | 3 | 2 b | $\begin{array}{r} 2.33- \\ 6.17 \end{array}$ | 8-20\% | 4-10\% | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ |
| $\begin{aligned} & \text { VA } \\ & 0385 \end{aligned}$ | Onion, bulb (all commodities) | highest utilization: raw without skin | 0 | 0.05 | 1.000 | JP | Child, 1-6 yrs | 748 | 102.00 | 244.4 | 3 | 2 b | $\begin{array}{r} 0.08- \\ 0.93 \end{array}$ | 0-3\% | 0-1\% | 0-3\% |
| $\begin{aligned} & \text { VA } \\ & 0389 \end{aligned}$ | Spring onion <br> (all commodities) | highest utilization: cooked/boiled | 0 | 0.65 | 1.000 | NL | Child, 2-6 yrs | E | 20.30 | 30.0 | 3 | 2 b | $\begin{array}{r} 1.66- \\ 2.15 \end{array}$ | 6-7\% | 3-3\% | 6-7\% |
| $\begin{aligned} & \text { VB } \\ & 0041 \end{aligned}$ | Cabbage, head (all commodities) | highest utilization: raw | 0 | 0.35 | 1.000 | CN | Child, 1-6 <br> yrs | 287 | 255.54 | 1402.5 | 3 | 2 b | $\begin{array}{r} 13.35- \\ 16.63 \end{array}$ | $\begin{gathered} 40- \\ 60 \% \end{gathered}$ | $\begin{array}{r} 20- \\ 30 \% \end{array}$ | $\begin{gathered} 40- \\ 60 \% \end{gathered}$ |

## Annex 4

DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$


DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$

|  |  |  |  |  |  |  |  |  |  |  |  |  |  | all | gen pop | child |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR- } \\ \mathrm{P} \\ \mathrm{mg} / \mathrm{kg} \end{array}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n |  | Unit weight, edible portion, | Variability factor | Case | $\begin{array}{r} \text { IESTI } \\ \mu \mathrm{g} / \mathrm{kg} \\ \mathrm{bw} / \mathrm{day} \end{array}$ | ARfD rounded | ARfD rounded |  |
| $\begin{aligned} & \text { VL } \\ & 0467 \end{aligned}$ | Chinese cabbage, type pe-tsai <br> (all commodities) | highest utilization: <br> Total | 0 | 0.02 | 1.000 | CN | Child, 1-6 yrs | 2788 | 336.16 | 1500.0 | 3 | 2 b | $\begin{array}{r} 0.25- \\ 1.25 \end{array}$ | 1-4\% | 1-3\% | 1-4\% |
| $\begin{aligned} & \text { VL } \\ & 0480 \end{aligned}$ | Kale (borecole, collards) (all commodities) | highest utilization: Total | 0 | 0.02 | 1.000 | DE | Gen pop, $14-80 \mathrm{yrs}$ | 123 | 669.80 | 672.0 | 3 | 2 b | $\begin{array}{r} 0.33- \\ 0.53 \end{array}$ | 1-2\% | 1-2\% | 1-2\% |
| $\begin{aligned} & \text { VL } \\ & 0482 \end{aligned}$ | Lettuce, head (all commodities) | highest utilization: raw | 0 | 0.5 | 1.000 | NL | Child, 2-6 yrs | 91 | 140.10 | 338.9 | 3 | 2 b | $\begin{array}{r} 4.7- \\ 11.42 \end{array}$ | $\begin{gathered} 20- \\ 40 \% \end{gathered}$ | $\begin{gathered} 10- \\ 20 \% \end{gathered}$ | $\begin{gathered} 20- \\ 40 \% \end{gathered}$ |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | Total |  | 0.5 | 1.000 | CN | Child, 1-6 yrs | 243 | 387.25 | 305.4 | 3 | 2a | 30.92 | 100\% | 30\% | 100\% |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | raw |  | 0.5 | 1.000 | NL | Child, 2-6 yrs | 91 | 140.10 | 117.8 | 3 | 2a | 10.21 | 30\% | 10\% | 30\% |
| $\begin{aligned} & \text { VL } \\ & 0483 \end{aligned}$ | Lettuce, leaf | cooked/boiled |  | 0.5 | 1.000 | NL | Gen pop, $>1$ yrs | 2 | 220.89 | 79.0 | 3 | 2a | 2.88 | 10\% | 10\% | NC |
| $\begin{aligned} & \text { VL } \\ & 0502 \end{aligned}$ | Spinach <br> (all commodities) | highest utilization: Total | 0 | 0.5 | 1.000 | ZA | Child, 1-5 yrs | - | 237.48 | 197.8 | 3 | 2a | $\begin{aligned} & 2.22- \\ & 22.29 \end{aligned}$ | 7-70\% | 7-20\% | 7-70\% |
| $\begin{aligned} & \text { VP } \\ & 0061 \end{aligned}$ | Beans, green, with pods, raw: beans except broad bean \& soya bean (i.e. immature seeds + pods) (Phaseolus spp) (all commodities) | highest utilization: canned/preserved | 0 | 0.2 | 1.000 | NL | Toddler, $8-20 \mathrm{~m}$ | E | 127.90 | 2.3 | NR | 1 | $\begin{array}{r} 0.76- \\ 2.51 \end{array}$ | 3-8\% | 3-5\% | 8-8\% |
| $\begin{aligned} & \text { VP } \\ & 0064 \end{aligned}$ | Peas, green, without pods, raw (i.e. immature seeds only) (Pisum spp) (all commodities) | highest utilization: <br> Total | 0 | 0.2 | 1.000 | UK | Child, 1.5-4.5 yrs | 57 | 174.00 | $<25$ | NR | 1 | 0.76-2.4 | 3-8\% | 2-6\% | 4-8\% |

# Annex 4 

DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$

|  |  |  |  |  |  |  |  |  |  |  |  |  |  | all | gen pop | child |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity | Processing | STMR <br> or STMRP $\mathrm{mg} / \mathrm{kg}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, $\mathrm{g} /$ person | Unit weight, edible portion, g | Variability factor | Case | IESTI $\mu \mathrm{g} / \mathrm{kg}$ bw/day | ARfD rounded | $A D f$ <br> ARfD rounded | ARfD rounded |
| $\begin{aligned} & \text { VR } \\ & 0494 \end{aligned}$ | Radish <br> (all commodities) | highest utilization: raw with skin | 0 | 0.1 | 1.000 | NL | Child, 2-6 <br> yrs | E | 64.40 | 172.0 | 3 | 2 b | $\begin{array}{r} 0.14- \\ 1.05 \end{array}$ | 0-4\% | 0-1\% | 0-4\% |
| $\begin{aligned} & \text { VR } \\ & 0577 \end{aligned}$ | Carrot <br> (all commodities) | highest utilization: raw with skin | 0 | 0.5 | 1.000 | CN | Child, 1-6 <br> yrs | 400 | 234.68 | 300.0 | 3 | 2 b | $\begin{aligned} & 4.08- \\ & 21.82 \end{aligned}$ | $\begin{gathered} 10- \\ 70 \% \end{gathered}$ | $\begin{gathered} 10- \\ 30 \% \end{gathered}$ | $\begin{gathered} 10- \\ 70 \% \end{gathered}$ |
| $\begin{aligned} & \text { VR } \\ & 0589 \end{aligned}$ | Potato <br> (all commodities) | highest utilization: Total | 0 | 0 | 1.000 | ZA | Child, 1-5 <br> yrs | - | 299.62 | 216.0 | 3 | 2a | $0-0$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { VR } \\ & 0596 \end{aligned}$ | Sugar beet (all commodities) | highest utilization: Total | 0 | 0.1 | 1.000 | DE | Gen pop, 14-80 yrs | 26295 | 161.79 | 160.0 | 3 | 2a | $\begin{array}{r} 0.63- \\ 0.63 \end{array}$ | 2-2\% | 2-2\% | 0-0\% |
| $\begin{aligned} & \text { GC } \\ & 0645 \end{aligned}$ | Maize (corn) | Total |  | 0 | 1.000 | CN | Child, 1-6 yrs | 166 | 524.69 | $<25$ | NR | 3 | ND | - | - | - |
| $\begin{aligned} & \text { TN } \\ & 0660 \end{aligned}$ | Almonds (all commodities) | highest utilization: raw incl roasted | 0 | 0.03 | 1.000 | DE | Women, $14-50 \mathrm{yrs}$ | 24 | 100.00 | 1.2 | NR | 1 | $\begin{array}{r} 0.03- \\ 0.04 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { TN } \\ & 0678 \end{aligned}$ | Walnut (all commodities) | highest utilization: raw incl roasted | 0 | 0 | 1.000 | DE | Child, 2-4 yrs | 75 | 49.40 | 7.0 | NR | 1 | 0-0 | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { DH } \\ & 1100 \end{aligned}$ | Hops, dry | Total |  | 0.45 | 1.000 | DE | Gen pop, 14-80 yrs | 5866 | 8.50 | $<25$ | NR | 3 | ND | - | - | - |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | Meat from mammals other than marine mammals | Total | NA | NA | 1.000 | CN | Child, 1-6 yrs | 302 | 264.84 | NR | NR | 1 | NA | 30\% | 20\% | 30\% |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | Meat from mammals other than marine mammals: $20 \%$ as fat | Total |  | 2 | 1.000 | CN | Child, 1-6 yrs | 302 | 52.97 | NR | NR | 1 | 6.57 | 20\% | 10\% | 20\% |
| $\begin{aligned} & \text { MM } \\ & 0095 \end{aligned}$ | Meat from mammals other than marine mammals: $80 \%$ as muscle | Total |  | 0.13333 | 1.000 | CN | Child, 1-6 yrs | 302 | 211.87 | NR | NR | 1 | 1.75 | 6\% | 4\% | 6\% |
| MO | Edible offal | Total |  | 0.03 | 1.000 | US | Child, 1-6 | - | 186.60 | NR | NR | 1 | 0.37 | 1\% | 1\% | 1\% |

DIAZINON (22)
ARfD $=0.03 \mathrm{mg} / \mathrm{kg}$ bw $(30 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$


MALATHION (49)
ARfD $=2 \mathrm{mg} / \mathrm{kg}$ bw $(2000 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw})$

|  |  |  |  |  |  |  |  |  |  |  |  |  |  | all | gen pop | child |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Codex code | Commodity | Processing | $\begin{array}{r} \text { STMR } \\ \text { or } \\ \text { STMR-P } \\ \mathrm{mg} / \mathrm{kg} \end{array}$ | HR or HR-P $\mathrm{mg} / \mathrm{kg}$ | DCF | Country | Population group | n | Large portion, g/person | Unit weight, edible portion, g | Variability factor | Case |  | \% <br> ARfD rounded | \% <br> ARfD <br> rounded | ARfD rounded |
| 001 | CITRUS FRUITS | - |  | 0.22 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { FC } \\ & 0303 \end{aligned}$ | Kumquats (all commodities) | highest utilization: Total | 0 | 0.22 | 1.000 | JP | Gen pop, $>1$ yrs | 135 | 120.00 | $<25$ | NR | 1 | $\begin{array}{r} 0.04- \\ 0.53 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { FC } \\ & 0204 \end{aligned}$ | Lemon (all commodities) | highest utilization: Total | 0.02 | 0.22 | 1.000 | FR | Child, 3-6 <br> yrs | 0 | 58.15 | 64.0 | 3 | 2 b | $\begin{array}{r} 0.01- \\ 2.03 \end{array}$ | 0-0\% | 0-0\% | 0-0\% |
| $\begin{aligned} & \text { FC } \\ & 0205 \end{aligned}$ | Lime <br> (all commodities) | highest utilization: Total | 0.02 | 0.22 | 1.000 | AU | $\begin{aligned} & \text { Gen pop, } \\ & >2 \text { yrs } \end{aligned}$ | 579 | 259.21 | 49.0 | 3 | 2a | 0-1.17 | 0-0\% | 0-0\% | 0-0\% |
| 001B | Mandarins | - |  | 0.22 | - | - | - | - | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { FC } \\ & 0003 \end{aligned}$ | Mandarins (incl mandarin-like hybrids) (all commodities) | highest utilization: raw, without peel | 0.02 | 0.22 | 1.000 | CN | Child, 1-6 <br> yrs | 151 | 586.75 | 124.3 | 3 | 2a | 0-11.39 | 0-1\% | 0-0\% | 0-1\% |
| $\begin{aligned} & \text { FC } \\ & 0004 \end{aligned}$ | Oranges, sweet, sour (incl orange-like hybrids) (all commodities) | highest utilization: Total | 0.02 | 0.22 | 1.000 | AU | Child, 2-6 yrs | 1735 | 800.83 | 155.8 | 3 | 2a | $\begin{aligned} & 0.01- \\ & 12.88 \end{aligned}$ | 0-1\% | 0-0\% | 0-1\% |
| $\begin{aligned} & \text { FC } \\ & 0005 \end{aligned}$ | Pummelo and Grapefruits (incl Shaddock-like hybrids, among others Grapefruit) (all commodities) | highest utilization: raw, without peel | 0.02 | 0.22 | 1.000 | DE | Child, 2-4 yrs | 12 | 358.60 | 178.5 | 3 | 2a | 0-9.75 | 0-0\% | $0-0 \%$ | 0-0\% |
| FP 0226 | Apple <br> (all commodities) | highest utilization: Total | 0.11 | 0.37 | 1.000 | US | Child, 1-6 yrs | - | 624.45 | 127.0 | 3 | 2a | $\begin{aligned} & 0.33- \\ & 21.67 \end{aligned}$ | 0-1\% | 0-0\% | 0-1\% |
| FB | Blueberries | highest utilization: | 0 | 7.5 | 1.000 | DE | Gen pop, | 70 | 388.00 | 1.8 | NR | 1 | 17.6- | 1-2\% | 0-2\% | 1-2\% |

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MALATHION (49)
ARfD $=2 \mathrm{mg} / \mathrm{kg}$ bw $(2000 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$
Maximum \%ARfD:
$9 \% \quad 5 \%$
all gen pop child


## Annex 4

MALATHION (49)
ARfD $=2 \mathrm{mg} / \mathrm{kg}$ bw $(2000 \mu \mathrm{~g} / \mathrm{kg}$ bw $)$


## ANNEX 5: REPORTS AND OTHER DOCUMENTS RESULTING FROM PREVIOUS JOINT MEETINGS OF THE FAO PANEL OF EXPERTS ON PESTICIDE RESIDUES IN FOOD AND THE ENVIRONMENT AND THE WHO CORE ASSESSMENT GROUP ON PESTICIDE RESIDUES

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2. Evaluation of the toxicity of pesticide residues in food. Report of a Joint Meeting of the FAO Committee on Pesticides in Agriculture and the WHO Expert Committee on Pesticide Residues. FAO Meeting Report, No. PL/1963/13; WHO/Food Add./23, 1964.
3. Evaluation of the toxicity of pesticide residues in food. Report of the Second Joint Meeting of the FAO Committee on Pesticides in Agriculture and the WHO Expert Committee on Pesticide Residues. FAO Meeting Report, No. PL/1965/10; WHO/Food Add./26.65, 1965.
4. Evaluation of the toxicity of pesticide residues in food. FAO Meeting Report, No. PL/1965/10/1; WHO/Food Add./27.65, 1965.
5. Evaluation of the hazards to consumers resulting from the use of fumigants in the protection of food. FAO Meeting Report, No. PL/1965/10/2; WHO/Food Add./28.65, 1965.
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#### Abstract

A Joint Meeting of the FAO Panel of Experts on Pesticide Residues in Food and the Environment and the WHO Core Assessment Group on Pesticide Residues (JMPR) was held in Geneva, Switzerland, from 9 to 13 May 2016. The three pesticides evaluated at the meeting were placed on the agenda by the JMPR Secretariat following the recommendation of an electronic task force of the WHO Core Assessment Group that they be re-evaluated due to public health concerns identified by the International Agency for Research on Cancer (IARC) and the availability of a significant number of new studies. During the meeting, the WHO Core Assessment Group was responsible for reviewing epidemiological, toxicological and related data in order to establish acceptable daily intakes (ADIs) and acute reference doses (ARfDs) of the pesticides for humans, where necessary. As no residue data were requested, the FAO Expert was responsible for estimating the dietary exposures (both short-term and long-term) to the pesticides reveiewed and, on this basis, performed dietary risk assessments in relation to their ADIs or ARfDs. This report contains information on ADIs, ARfDs and general principles for the evaluation of pesticides. The recommendations of the Joint Meeting, including further research and information, are proposed for use by Member governments of the respective agencies and other interested parties.


[^0]:    ** Evaluated following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that the compound be re-evaluated due to public health concerns identified by IARC and the availability of a significant number of new studies

[^1]:    ${ }^{1}$ Did not attend the meeting.
    ${ }^{2}$ Did not attend the meeting, but her valuable contributions to the methodological setup of the epidemiological evaluation are gratefully acknowledged.
    ${ }^{3}$ Attended part of the meeting only.

[^2]:    ${ }^{1}$ Did not attend the meeting.

[^3]:    ${ }^{1}$ See, for example, Clewell HJ, Crump KS. Quantitative estimates of risk for noncancer endpoints. Risk Anal. 2005;25(2):285-9; and Nachman KE, Fox MA, Sheehan MC, Burke TA, Rodricks JV, Woodruff TJ. Leveraging epidemiology to improve risk assessment. Open Epidemiol J. 2011;4:3-29.

[^4]:    ${ }^{1}$ IARC. Some organophosphate insecticides and herbicides: tetrachlorvinphos, parathion, malathion, diazinon and glyphosate. Lyon: International Agency for Research on Cancer; 2015 (IARC Monographs on the Evaluation of Carcinogenic Risks to Humans, Volume 112).
    ${ }^{2}$ Koutros S, Silverman DT, Alavanja MC, Andreotti G, Lerro CC, Heltshe S et al. Occupational exposure to pesticides and bladder cancer risk. Int J Epidemiol. 2015; pii: dyv195 [Epub ahead of print]; and Lerro CC, Koutros S, Andreotti G, Friesen MC, Alavanja MC, Blair A et al. Organophosphate insecticide use and cancer incidence among spouses of pesticide applicators in the Agricultural Health Study. Occup Environ Med. 2015; 72(10):736-44.
    ${ }^{3}$ Mills PK, Yang R. Prostate cancer risk in California farm workers. J Occup Environ Med. 2003; 45(3):249-58.

[^5]:    ${ }^{1}$ Guyatt GH, Oxman AD, Vist GE, Kunz R, Falck-Ytter Y, Alonso-Coello P et al. GRADE: an emerging consensus on rating quality of evidence and strength of recommendations. BMJ. 2008;336:924-6; and Morgan RL, Thayer KA, Bero L, Bruce N, Falck-Ytter Y, Ghersi D et al. GRADE: Assessing the quality of evidence in environmental and occupational health. Environ Int. 2016;doi: 10.1016/j.envint.2016.01.004 [Epub ahead of print].

[^6]:    Dietary administration.
    ${ }^{\text {b }}$ Acetylcholinesterase activity not measured.
    ${ }^{\text {c }}$ Highest dose tested.
    ${ }^{\mathrm{d}}$ Gavage administration.
    e Two or more studies combined.
    ${ }^{\mathrm{f}}$ Included in the overall NOAEL for rats and dogs.

[^7]:    ${ }^{\text {a }}$ Unlikely to pose a carcinogenic risk to humans via exposure from the diet.
    ${ }^{b}$ Secondary to local irritation of the colon.

[^8]:    ${ }^{\text {a }}$ Unlikely to pose a carcinogenic risk to humans via exposure from the diet.
    ${ }^{\mathrm{b}}$ Acetylcholinesterase activity not measured.
    ${ }^{\text {c }}$ Ninety-day neurotoxicity study in rats is covered by the overall oral NOAEL for repeated-dose studies of toxicity.

[^9]:    ${ }^{\text {a }}$ Group ADI for the sum of glyphosate, AMPA, $N$-acetyl-glyphosate and $N$-acetyl-AMPA.
    ${ }^{\mathrm{b}}$ Established for the sum of malathion and malaoxon (corrected for its potency), expressed as parent malathion.

[^10]:    ** Evaluated following the recommendation of an electronic task force of the WHO Core Assessment Group on Pesticide Residues that the compound be re-evaluated due to public health concerns identified by IARC and the availability of a significant number of new studies

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[^12]:    ${ }^{1}$ Did not attend the meeting.

[^13]:    ${ }^{1}$ See, for example, Clewell HJ, Crump KS. Quantitative estimates of risk for noncancer endpoints. Risk Anal. 2005;25(2):285-9; and Nachman KE, Fox MA, Sheehan MC, Burke TA, Rodricks JV, Woodruff TJ. Leveraging epidemiology to improve risk assessment. Open Epidemiol J. 2011;4:3-29.

[^14]:    ${ }^{1}$ IARC. Some organophosphate insecticides and herbicides: tetrachlorvinphos, parathion, malathion, diazinon and glyphosate. Lyon: International Agency for Research on Cancer; 2015 (IARC Monographs on the Evaluation of Carcinogenic Risks to Humans, Volume 112).
    ${ }^{2}$ Koutros S, Silverman DT, Alavanja MC, Andreotti G, Lerro CC, Heltshe S et al. Occupational exposure to pesticides and bladder cancer risk. Int J Epidemiol. 2015; pii: dyv195 [Epub ahead of print]; and Lerro CC, Koutros S, Andreotti G, Friesen MC, Alavanja MC, Blair A et al. Organophosphate insecticide use and cancer incidence among spouses of pesticide applicators in the Agricultural Health Study. Occup Environ Med. 2015; 72(10):736-44.
    ${ }^{3}$ Mills PK, Yang R. Prostate cancer risk in California farm workers. J Occup Environ Med. 2003; 45(3):249-58.

[^15]:    ${ }^{1}$ Guyatt GH, Oxman AD, Vist GE, Kunz R, Falck-Ytter Y, Alonso-Coello P et al. GRADE: an emerging consensus on rating quality of evidence and strength of recommendations. BMJ. 2008;336:924-6; and Morgan RL, Thayer KA, Bero L, Bruce N, Falck-Ytter Y, Ghersi D et al. GRADE: Assessing the quality of evidence in environmental and occupational health. Environ Int. 2016;doi: 10.1016/j.envint.2016.01.004 [Epub ahead of print].

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[^18]:    ${ }^{\text {a }}$ Unlikely to pose a carcinogenic risk to humans via exposure from the diet.
    ${ }^{\mathrm{b}}$ Acetylcholinesterase activity not measured.
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